

# Anvil Range Mining Complex 2005 Seepage Investigation at the Emergency Tailings Area

# 2005/06 - Task 20e

Prepared for

Deloitte & Touche Inc.

on behalf of

Faro Mine Closure Planning Office





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# **Deloitte & Touche Inc.**

On behalf of

# **The Faro Mine Closure Planning Office**

#### SRK Consulting (Canada) Inc.

Suite 800, 1066 West Hastings Street Vancouver, B.C. V6E 3X2

Tel: 604.681.4196 Fax: 604.687.5532 E-mail: vancouver@srk.com Web site: www.srk.com

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Authors Daniel Mackie, M.Sc. Quinn Jordan-Knox, M.Sc. Dr. Christoph Wels, Ph. D., M.Sc.

> Reviewed by Cam Scott, P.Eng.

# **Executive Summary**

An assessment of groundwater and surface water flow and water quality conditions at the Emergency Tailings Area (ETA) indicates that significant contaminant load (up to 70 t/yr zinc) is reporting to the Faro Creek Canyon and Rose Creek valley. The contaminated seepage discharging from the ETA is a combination of surface flow originating as toe seepage from the Faro waste rock dumps) and groundwater flow through highly permeable alluvial sediments in the historic Faro Creek valley. The tailings stored in the ETA also contribute significantly to the contaminant loading. Therefore, a seepage interception system (SIS) will be required in the ETA, probably for an indefinite period of time.

Drilling and hydraulic testing indicates that bedrock is significantly (> 3 orders of magnitude) less transmissive than overburden in this area. In addition, groundwater in bedrock also shows significantly lower contaminant concentrations than observed in the overburden. As a result, the estimated contaminant loads associated with groundwater flow in bedrock are 3-4 orders of magnitude lower than the contaminant loads observed in the overburden soils. These initial loading estimates suggest that interception of groundwater flow in bedrock in the ETA area may not be required.

Pumping test results and scoping level numerical modeling indicate that groundwater pumping alone in the ETA will likely not provide the necessary level of capture. Instead, a Primary SIS, a monitoring system and an Adaptive Management Program are recommended. As a first step, the tailings should be removed from the ETA itself. The primary SIS, consisting of a cut-off wall with up gradient pumping wells should then be installed. Limited sections of permeable trench should be installed along the fence of pumping wells in areas with relatively low permeability materials (till). A monitoring system and an adaptive management plan should be implemented to allow seepage cature efficiency to be monitored and upgraded as necessary.

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# 1 Introduction and Scope of Work

This report presents results of the 2005/2006 hydrogeology program for the Emergency Tailings Area (ETA) as part of Task 20e – Continued Seepage Investigations. Seepage from the Faro waste rock dump, and the subsequent influence of tailings in the ETA itself, have been identified as a source of contamination to the Rose Creek Aquifer. As a result, impacts and potential mitigation measures have been investigated. Figure 1 shows the location of the study area.

The scope of work for Task 20e was described in a memorandum dated June 8, 2005 and included investigations in four areas: the Emergency Tailings Area (ETA), Zone 2 Pit area, the S-cluster area, and the Grum area. This report covers the ETA component. The primary objectives addressed herein are:

- Drilling and completion of two inclined bedrock drillholes with core logging, packer testing and completion as bedrock monitoring wells,
- Additional hydraulic testing of the pumping well installed in 2004,
- A detailed stream survey, and
- Development of recommendation for a collection system.

A brief review of background information is presented in Section 2. Descriptions of field program methodologies and results are presented in Section 3, including a hydrogeologic conceptual model for the ETA area. Contaminant loading estimates in groundwater and surface water are presented in Section 4. Section 5 describes options and recommendations for a seepage interception system.

# 2 Background

## 2.1 Initial Data Review

An "Initial Review of Groundwater Quality downstream of Faro, Grum and Vangorda WRDs, Yukon Territory", by Robertson GeoConsultants Inc., dated July 14, 2004, is provided in Appendix A of the Preliminary Seepage Collection Options – Faro and Grum Waste Rock Dumps (SRK, 2006). Available groundwater quality data was reviewed and priorities for further work were assigned based on the observed trends in acid rock drainage (ARD) related contaminants, specifically, zinc and sulphate, as well as parameters such as pH and alkalinity. The following paragraphs summarize the findings relevant to the ETA.

### Surface Seepage in Faro Creek Channel (X23)

Water quality has been monitored regularly since 1986 at station X23 where seepage from the Main Dump "daylights" year-round (Figure 2a and 2b). Sulphate concentrations increased from around 1,500 mg/L to more than 4,000 mg/L with spikes to above 6,000 mg/L. Zinc concentrations increased from around20 mg/L in the late 1980's to around200-400 mg/L with spikes up to greater than 1,000 mg/L since about 2001. Total and dissolved iron concentrations show 10-100 fold increases since about 2001 (Figure 2b). Interception and treatment of surface seepage at this location was recommended.

#### Subsurface Seepage in Faro Creek Channel (ETA)

Two monitoring wells were installed close to X23 in 1996 (P96-8A and 8B), completed in overburden and fractured bedrock. Sulphate and zinc concentrations show similar trends to surface water quality at X23. Sulphate increased from around 2,000 mg/L in 1996 to around 4,000 mg/L in 2003. Zinc increased from around2 mg/L to greater than100 mg/L in the same 8-year time span. Alkalinity ranged from 150-350 mg/L and pH from 6-7, suggesting possible buffering within the WRD itself. Further characterization studies to quantify subsurface seepage from the WRD, as well as detailed flow measurements along the creek downstream from X23 were recommended.

Potential seepage from the Faro Pit along the old Faro Creek channel was also identified as an area for further investigation. While seepage from Faro Pit into the old Faro creek channel is not considered a current issue due to the lower elevation of the water level in the pit relative to the topographic low on the pit wall along the creek channel, increased load to the creek channel could become an issue if the pit water level is allowed to rise. Further studies of subsurface conditions and installation of a monitoring well were suggested, but were considered a low priority relative to areas where seepage had already been observed.

## 2.2 2004 Field Program

The recommendations of the initial data review provided a framework for developing the 2004 hydrogeological field investigation that is described in Preliminary Seepage Collection Options – Faro and Grum Waste Rock Dumps (SRK, 2006). The 2004 program included drilling, well installation, hydraulic testing, test pitting and geophysical surveying of targets based as prioritized by the initial review of groundwater quality. The final program also included reduction of all field data and the identification of preliminary seepage collection options. Preliminary costs were estimated and a preliminary preferred option was presented.

Analysis of water quality data at the ETA, from new (2004) drillholes, older (1996) drillholes and surface water, indicated that the tailings in the ETA are contributing significant contaminant load to the Rose Creek Valley tailings impoundment and underlying aquifer.

Hydraulic testing and measurements of surface flow suggested that approximately 6 l/s of highly contaminated water is discharging to the Rose Creek Valley. This water is likely a combination of seepage from the main Faro waste rock dump and mill area, both located up gradient from the ETA, and precipitation that falls directly on the ETA. Water quality is poorer at the downstream side of the ETA than on the upstream side, suggesting that the tailings in the ETA are contributing to the overall contaminant loading.

The report from the 2004 program recommended a seepage collection system consisting of a line of pumping wells located adjacent to the mine access road, combined with surface water collection sumps.

# 3 Field Investigation

## 3.1 Monitoring Well Drilling and Installation

Seven new monitoring wells were drilled in the ETA in 2005.Boreholes SRK05-ETA-BR1 and – BR2 were completed in bedrock as part of Task 22e to characterize potential bedrock groundwater flow. The remaining five (ETA05-01 to -05) were completed as part of a separate investigation of the ETA tailings, but were also utilised for the groundwater program.

Locations for all ETA monitoring wells are shown in Figure 1. Table 1 lists completion details for the 2005 monitoring wells, as well as the five previously existing 2004 monitoring wells. Borehole logs for all newly completed monitoring wells are provided in Appendix A.

2005 Monitoring Wells	Easting	Northing	Total Depth (m)	Top of Casing Elevation (m.a.s.l.)	Screen Interval (m.b.g.s.)			
ETA05-01	582953	6913851	9.1	1105.13	4.3-7.3			
ETA05-02	582993	6913849	7.6	1105.06	5.0-7.5			
ETA05-03	582975	6913808	9.0	1103.98	7.0-8.8			
ETA05-04	583045	6913858	9.0	1105.4	6.7-9.0			
ETA05-05	582977	6913856	7.5	1105.44	2.0-5.2			
SRK05-ETA-BR1	582972	6913846	13.0	1105.21	9-12			
SRK05-ETA-BR2	582987	6913825	23.6	1103.75	14.6-18.9			
	200	4 Monitoring	Wells					
SRK-04-04	582,977	6,913,837	11.9	1104.80	7.6 – 11.6			
SRK04-3A	582,977	6,913,824	13.4	1104.55	10.4 – 11.9			
SRK04-3B	582,977	6,913,824	13.4	1104.63	5.5 – 7.0			
Historic Monitoring Wells								
P96-8A	577,050	6,911,223	~4.5	1109.39	1.0-4.5			
P96-8B	577,050	6,911,223	~8.5	1109.48	5.5-8.5			

Table 1: ETA Monitoring Well Summary

The 2005 drilling program was conducted using two different drill types. A track mounted sonic drill owned and operated by SDS Sonic Drilling out of Calgary, Alberta, was used to drill shallow boreholes ETA05-01 through -05. A track-mounted Maxidrill coring drill, owned and operated by Midnight Sun Drilling of Whitehorse, Yukon, was used to drill the two deeper bedrock holes.

The sonic rig was equipped with a 4x6 system (4" core barrel and 6" casing) that allowed for continuous sampling in 10 foot runs (1 core barrel; approximately 3 metres) by advancing the core barrel using ultra-sonic vibrations. Casing is advanced over the core barrel to below the bit to keep the hole open during barrel retrieval. Water is only used during casing advancement to prevent

heave between barrel and casing. Run samples are extruded into 4" diameter clear plastic sample bags for logging and grab sampling. Rods and casing were in imperial units and all units have been converted to metric. Final drillhole diameter was 152 mm (six-inches).

The Maxidrill was a combination reverse circulation / diamond drilling machine, but only the diamond drill component was employed. Alluvium was drilled with NW size casing (88.9 mm O.D.) and shoe to bedrock to stabilize the overburden. After casing was set into bedrock, NQ drillrods and drill bit (75.7 mm O.D.) were lowered through the casing for bedrock coring. Casing could not be set into bedrock in SRK05-ETA-BR1 because of coarse gravel or boulders. This borehole was terminated in the alluvium. SRK05-ETA-BR2 was successfully completed in bedrock.

All drillholes were logged. Packer injection tests were completed in bedrock during drilling of SRK05-ETA-BR2.

Piezometers in ETA05-01 to -05 were constructed with standard (Schedule 80), two-inch diameter, PVC components (solid and screen) installed through the Sonic drill-rods. Sand packs were emplaced around the PVC screens and then a grout seal poured to the surface. Depths of materials were measured during installation.

Piezometers in SRK05-ETA-BR1 and –BR2 were completed differently than the others. Sections of 2.5 cm (1") PVC solid pipe and screen were installed through the Maxidrill drill rods to hole bottom. A rubber "shale trap" (funnel-shaped rubber seal with outer diameter of the drillhole) was fixed to the solid PVC above the screen zone. A bentonite grout (mix approximately ½ bag grout to 5 US gallons of water) was used to seal the well annulus above the shale trap. Grout was mixed at surface and pumped down a 1" (26 mm) PVC tremmie pipe to the bottom of the hole using a pneumatic mud pump.

Two monitoring wells were developed using a peristaltic pump: SRK05-ETA-BR1 and SRK05-ETA-BR2. Water was pumped from the wells until field conductivity and pH readings stabilized, and at least three well volumes had been removed. Monitoring wells ETA05-01 through-05 have not been developed.

# 3.2 Hydrostratigraphy

The interpreted cross-sections shown in Figures 3 and 4 were developed using the borehole logs and results of a ground penetrating radar survey, which was conducted as part of the 2004 investigations. Cross-section locations are shown on Figure 1. The cross-sections show the primary geologic units identified in the ETA area, which from top to bottom are: tailings, natural soil consisting of either alluvium or till, and bedrock.

## 3.2.1 Tailings

The tailings vary in texture from gravel and sand to silt with lenses of visible pyrite-rich sand/gravel observed in drill-core. The tailings deposit is approximately 6.5m thick near the access road and

appears to thin up-valley and towards the valley sides. Coarser materials (gravels) observed within the tailings deposit may represent waste rock placed onto the tailings surfaces after intermittent discharges to the ETA. These coarser units cannot be accurately delineated with the available data.

### 3.2.2 Alluvium and Till

The alluvium unit is dominantly comprised of coarse (sand and gravel sized) alluvium interpreted to have been deposited by the old Faro Creek near the mine access road, but is also interpreted to include some till of sandy or gravelly silt texture further up-valley (towards the Faro waste dumps) and towards the valley sides. The alluvial unit is interpreted to continue under the access road, sloping downwards with topography and is interpreted to vary between 2-6 m in thickness within the ETA. Fine grained (silt) till is assumed to blanket the adjacent hillsides bordering the ETA, but the distribution/thickness of alluvial and till deposits up-valley from the ETA is uncertain. Drill logs for P96-8, located up gradient of the ETA at the toe of the waste rock dump, indicate bedrock at approximately 8 m depth, overlain by coarse alluvial sand and gravel.

### 3.2.3 Bedrock

Bedrock in the ETA is characterised as phyllite, with a weathered zone ranging in thickness from 0.10 to 1.2 meters determined by the presence of iron staining on fractures in BR2. Drilling at most locations was terminated at the overburden-bedrock interface and did not provide detailed characterisation of weathered bedrock. The bedrock channel observed below the mine access road is interpreted to extend up-valley under the mine access road into the ETA to at least SRK04-4. The bedrock surface is interpreted to raise up-valley from the access road to the waste rock dumps and valley-sides.

Drilling at SRK05-ETA-BR2 was continued to a total depth of 23.6 m.b.g.s., intersecting approximately 12 meters of bedrock. Bedrock was characterised using standard geotechnical methods, including rock quality designation (RQD), solid core recovery (SCR) and fracture frequency (fractures/meter). Table 2 summarises bedrock characteristics.

DEPTH		RUN	RECOVERY		RQD		SCR		Fracture
FROM	то	LENGTH							frequency
(m)	(m)	(m)	(m)	(%)	(m)	(%)	(m)	(%)	(Joints/III)
10.20	11.60	1.40	1.33	95.0	0.34	24.0	0.70	50.0	14.0
11.60	13.10	1.52	1.52	100.0	0.26	17.0	1.02	67.0	18.0
13.10	14.60	1.52	1.52	100.0	0.71	47.0	1.01	66.0	15.0
14.60	16.15	1.52	1.52	100.0	0.15	10.0	0.89	58.0	20.0
16.15	17.80	1.65	1.60	97.0	0.17	10.0	0.97	59.0	21.0
17.80	19.33	1.52	1.48	100.0	0.33	22.0	0.94	61.0	17.0
19.33	20.60	1.27	1.27	100.0	0.32	25.0	0.76	59.0	20.0
20.60	22.10	1.52	1.52	100.0	0.72	47.0	1.08	71.0	16.0
22.10	23.60	1.52	1.52	100.0	0.36	23.0	1.15	76.0	15.0

Table 2: SRK05-ETA-BR2 Bedrock Characteristics

Bedrock in the ETA area is characterised as phyllite or schist, both of which have shown at least one relatively well developed foliation surface. The presence of foliation can complicate geotechnical and hydrogeological characterisation due to the inherent weakness, and subsequent "apparent" high fracture frequency that can result. Many breaks, if not most, are caused by the drilling process itself, which may cause the core to break along the foliation surface. Plate 1 shows drillcore from SRK05-ETA-BR2. The preferential fracture orientation parallel to foliation suggests that measured RQD and fracture frequency values are conservative and rock quality is likely better than observed in core.



Plate 1: Drillcore from SRK05-ETA-BR2

## 3.3 Hydrology

## 3.3.1 Methodology

Surface water and seepage flows, as well as water quality, were monitored in 2005 at multiple locations above and below the ETA (shown on Figure 1). As part of this program, the weir at monitoring station X23 was improved and an additional weir installed at the mouth of the Faro Creek canyonby Laberge Environmental Services (LES). Continuous recording water level data loggers were installed at each weir and calibrated to manual discharge measurement on each site visit to develop stage-discharge relationships. The resulting data were used to assess the hydrologic dynamics of the ETA area, and, using regional analysis, develop preliminary estimates of 100-year flood flows.

Appendix B includes memoranda prepared by LES describing the weir installation and periodic site visits. Appendix C provides a memorandum prepared by Mr. Pat Bryan on the regional analysis and flood estimates.

### 3.3.2 Manual Streamflow Surveys

Discharge measurements and water quality (SO<sub>4</sub>, Zn-T and Fe-T) from the four stations along the Faro Creek channel (from X23 at the toe of the WRDs to the mouth of the Faro Creek Canyon) are summarised in Table 3. The following conclusions can be drawn with respect to streamflows and surface water quality in the ETA area:

- Toe seepage from the waste rock (FCS-1 at X23) occurs year-round with some decrease in seepage during winter baseflow; concentrations of sulphate, zinc and iron are highly elevated but have remained relatively steady during the period of observation;
- Surface runoff from the ETA area (FCS-2 at culvert) shows more variable flow than toe seepage up-gradient of the ETA area but concentrations of sulphate, zinc and iron are very similar to those observed in toe seepage up-gradient of the ETA area (except for a small decrease in total iron concentrations);
- Subsurface seepage from the ETA area (FCS-3 at downstream seepage face) flows year-round with flow estimates ranging from 2.3 4.8 L/s; water quality of this seepage is generally similar to that observed in the alluvial wells in the ETA area; iron concentrations in this seepage are consistently about one order of magnitude higher than in waste rock seepage entering (and leaving) the ETA area;
- Surface runoff at the mouth of the Faro Creek canyon (FCS-4) also flows year-round with flow rates ranging from 4.5 to 11.7 L/s (peak flows during snowmelt and/or storm events may be higher but have not yet been measured); the water quality represents a mixture of surface runoff and subsurface seepage from the ETA area with intermediate concentrations of total iron;
- The two most reliable flow surveys (May and October 2005) indicate that incremental gains in streamflow along the Faro Creek Canyon (between FCS-2/3 and FCS-4) are very small

(0.1 to 0.7 L/s) suggesting that there is little or no groundwater discharge along the Faro Creek canyon;

Station ID	Date	Flow L/s	S04 mg/L	Zn-T mg/L	Fe-T mg/L
	10-Apr-05	1.3	5,030	295	88.9
RD	13-May-05	4.6		no sample	
s-1 S-1	18-Oct-05	1.2	6,200	477	157
Ge C	21-Nov-05	1.0	6,370	458	131
At to	19-Dec-05	0.6	5,920	516	180
	Average	1.7	5,880	437	139
pad	10-Apr-05	frozen		no sample	
N IC	13-May-05	9.0		no sample	
S-2 belo	18-Oct-05	3.4	6,210	459	61.6
ertk	21-Nov-05	>1.0	5,890	437	36.8
N N N	19-Dec-05	frozen	no sample		
At 6	Average	n/a	6,050	448	49
MO	10-Apr-05	4.8	5,550	309	1,210
bel	13-May-05	2.6		no sample	
S-3 ace ad	18-Oct-05	3.1	6,570	222	1,120
ge f C	21-Nov-05	3.4	7,460	371	1,790
eba	19-Dec-05	2.3	7,030	430	1,990
Š	Average	3.2	6,653	333	1,528
c	10-Apr-05	6.6	4,170	174	801
nyo	13-May-05	11.7		no sample	
S-4 Ca	18-Oct-05	7.2	5,750	310	773
h of	21-Nov-05	5.4	5,610	266	940
lout	19-Dec-05	4.2	5,540	278	1,220
2	Average	7.1	5,268	257	934

Table 3: Faro Creek Flow and Water Quality

## 3.3.3 Streamflow Gauging

In September 2005, continuous monitoring of flow rates along the Faro Creek channel was initiated. For this purpose, the existing V-notch weir at monitoring stations FCS-1 (X23) was upgraded and a new V-notch weir was constructed at the mouth of the Faro Creek canyon (monitoring station FCS-4). Both weirs were equipped with a PT2X sensor/data logger to allow continuous monitoring of stage height. Rating curves were developed to convert the recorded stage height to streamflow (see Appendix B for details).

Figure 5 shows the observed flow rates between mid-October 2005 and mid-January 2006 at the two monitoring stations. Seepage flow at the toe of the WRD (at X23) has gradually decreased from about 1.3 L/s in mid-September 2005 to about 0.6 L/s in early January 2006. The occasional "spikes" are believed to be related to ice jams rather than runoff events.

Seepage flows at the mouth of the Faro Canyon (at FCS-4) have also gradually decreased from about 8 L/s in mid-September 2005 to 3.5 L/s in early January 2006. The observed sudden spikes in flow are believed to be a result of mill discharge into the ETA area (in mid-September 2005) or blockage of the weir due to ice build-up.

## 3.4 Hydraulic Testing

Hydraulic testing during the 2005 field program consisted of packer injection tests completed in SRK05-ETA-BR2 during drilling and, after completion of all monitoring wells, a 24-hour pumping test.

## 3.4.1 Packer Injection Testing

During drilling of SRK05-ETA-BR2, packer injection testing was completed in bedrock portions of the drillhole. Packer injection test equipment was provided by Midnight Sun Drilling, who conducted the tests under the supervision of SRK field staff. A wireline pneumatic packer system was used for testing, incorporating nitrogen-inflated packers sealing off the test interval below the drill rods, and injection of water under known pressure into the test zone using the drill water pump. A total of four packer tests were completed in SRK05-ETA-BR2, commencing close to the bottom of the surface casing at 10.2 meters below ground surface (m.b.g.s.). Details and test results are summarised in Table 4. Data sheets for packer tests are included in Appendix D-1

Test No.	Test Interval (m.b.g.s.)	Weighted Average RQD (%)*	Hydraulic Conductivity (m/s)
1	11.7 – 15.0	30	1.3 x 10 <sup>-8</sup>
2	16.0 – 19.0	15	1.1 x 10 <sup>-7</sup>
3	19.1 – 23.6	32	< 1.0 x 10 <sup>-8</sup>
4	12.0 – 23.6	25	4.3 x 10 <sup>-8</sup>

Table 4: Packer Testing Summary

\*Weighted Average RQD calculated by wieghting RQD % by run lengths

Despite the low RQD and high fracture frequency values determined for this rock (see Table 2), packer testing results suggest the bedrock has relatively low hydraulic conductivity values. This suggests that, overall, the observed bedrock fracturing is not indicative of actual subsurface conditions. In contrast, the highest hydraulic conductivity value corresponds to an area with the lowest RQD. It is difficult to determine the precise location of the more permeable fractures, but the packer testing data suggest that the bedrock may have zones with slightly higher hydraulic conductivity in bedrock in this borehole (~1x10<sup>-7</sup> m/s) is more than 3 orders of magnitudes lower than the hydraulic conductivity in the alluvial sediments overlying the bedrock (see below).

### 3.4.2 24-hour Pumping Test

#### Methodology

The pumping test was conducted by Aquatech of Whitehorse, Yukon, with field supervision by SRK staff. The test was completed using a submersible pump rated to greater than 4.7 L/s (75 USGPM). Pump flow rates were measured with an inline manometer attached to the discharge line, which was directed to the culvert passing underneath the mine access road, and from there down slope to the tailings impoundment, below the FCS4 Faro Creek Canyon monitoring station.

Prior to initiation of the pumping test, a step test was conducted in SRK04-4 to determine an appropriate test discharge rate. Pumping rates of 1.9, 3.2, 3.8 and 4.7 L/s (30, 50, 60 and 75 USgpm) were used over time intervals of 20 to 50 minutes and water level changes recorded. Water level data are shown on Figure 6. A test pumping rate of ~3.2 L/s (50 USgpm) was chosen as a rate for the 24-hour test.

A 24-hr constant rate pumping test was conducted on SRK-04-4 at an average rate of approximately 3.2 L/s (50 U.S.gpm) between 09:22 on October 2and 09:50 October 3. Water levels were recorded at 11 groundwater monitoring locations and two weirs on Faro Creek, one at FCS-1 (X23) and another at FCS-4 (discussed in Section 3.3). Visual observations were taken intermittently at the groundwater seep below the mine access road (station FCS3/X7) to determine if pumping resulted in a decrease of the seepage rate. Static groundwater levels were monitored immediately prior to the pumping test and are shown on Figure 7. Table 5 summarises the monitoring locations and measurement type.

Station Type	Monitoring Station ID	Measurement Method		
	P96-8A	Water Level Tape		
	P96-8B	M10 Level-logger / Water Level Tape		
	SRK04-4	Water Level Tape		
	SRK04-3A	M10 Level-logger / Water Level Tape		
	SRK04-3B	Water Level Tape		
Croupdwater Wall	ETA-05-1	Water Level Tape		
Groundwater weir	ETA-05-2	Water Level Tape		
	ETA-05-3	Water Level Tape		
	ETA-05-4	M10 Level-logger / Water Level Tape		
	ETA-05-5	Water Level Tape		
	SRK-05-ETA-BR1	M10 Level-logger / Water Level Tape		
	SRK-05-ETA-BR2	M10 Level-logger / Water Level Tape		
Surface Water Weir	FCS1/X23	Datalogger		
Surface Water Well	FCS4	Datalogger		

#### **Table 5: Monitoring Station Measurements**

#### Results

The 24-hour pumping test was successfully completed without any significant interruption or pump breakdowns. Discharge data during the testing period is included in Appendix D-2. In general, water levels at groundwater monitoring stations did not reach steady-state or indicate intersection with aquifer boundaries. Water levels and flow at each of the weirs indicated no apparent changes related to the pumping test (Figures 8 and 9). Seepage from below the mine access road was not observed to noticeably change during the testing period (based on visual observations). Figure 10 shows drawdown immediately prior to termination of the test.

Aquifer properties were determined from drawdown and recovery data with standard analytical techniques using the Waterloo Hydrogeologic software AquiferTest V3.5 and V4. Table 6 summarises the estimated aquifer properties (T and S) inferred from the drawdown/recovery data observed in the pumping well and the various monitoring wells. The best fits of the analytical solution to the drawdown/recovery data are provided in Appendix D-3.

Monitoring Well	Average Transmissivity (m²/d)	Average Storativity	Comment
Wells screened in All	luvial Aquifer		
SRK-04-4	50.0	NA	Pumping Well, only recovery data used in analysis
SRK-04-3A	77.9	0.018	Monitoring well at 13 m distance; drawdown and recovery data used
ETA-05-2	52.2	0.013	Monitoring well at20 m distance; drawdown and recovery data used
ETA-05-3	93.0 0.012		Monitoring well at29 m distance; drawdown and recovery data used
SRK-05-ETA-BR1	62.4	0.014	Monitoring well at 10 m distance; drawdown and recovery data used
Average	67.1	0.014	Arithmetic Average

Table 6: Results of Hydraulic Analyses from Pumping Test at SRK04-4

Note: only those wells screened in the alluvial aquifer were analysed; an interpretation of pump test response in wells screened in tailings and bedrock is considered questionable as modeling assumptions are not met

The results of the pump test can be summarized as follows:

- Aquifer transmissivity (T) was estimated to be approximately 67.1 m<sup>2</sup>/d; assuming an average aquifer thickness of 4m this represents an average hydraulic conductivity (K) of approximately  $2x10^{-4}$  m/s for the alluvial sediments.
- Aquifer storativity was estimated to be approximately 0.014; this estimate is significantly lower than typical values for an unconfined sand and gravel aquifer (0.1-0.2) and greater than typical values for confined aquifers (<0.001) suggesting leakage to the alluvial aquifer from the overlying tailings (semi-confined conditions).

- Wells screened across or within tailings showed drawdown and suggest potential for desaturation of the tailings.
- The bedrock monitoring well (SRK05-ETA-BR2) showed delayed but similar drawdown response to shallower wells screened within the aquifer (SRK05-ETA-BR1).

Note that packer testing in the bedrock borehole (SRK05-ETA-BR2) suggested an average hydraulic conductivity of  $4.3 \times 10^{-8}$  m/s for the bedrock in the ETA area. Assuming these packer test results are representative of the bedrock in the ETA area and further assuming a bedrock aquifer thickness of about 11.6m (equivalent to the testing interval), the bedrock would have a bulk transmissivity of 0.043 m<sup>2</sup>/day, or approximately three to four orders of magnitude lower than the overburden materials.

## 3.5 Hydrogeologic Conceptual Model

Based on geologic data, hydraulic testing results and hydrology data, the ETA is interpreted to have three hydrostratigraphic units and a limited, but in certain areas direct, connection of groundwater with surface flow in the former Faro Creek channel. A summary of ETA hydrostratigraphy is as follows:

Unit 1 – Tailings: silt to gravel size tailings up to 6.5 meters thick, thinning up-valley and towards the valley sides, with coarser, gravel-size, waste rock interbeds. Tailings may represent a partially confining layer to the primary aquifer, which may desaturate during pumping. Coarse waste rock interbeds may represent preferential pathways within the tailings. Hydraulic conductivity unknown, but interpreted to be lower than primary aquifer unit.

Unit 2 – Primary Aquifer: sand to gravel size alluvium interpreted to be deposits of historic Faro Creek, with sandy or gravelly-silt till in up-valley sections of the ETA and along valley walls. Includes weathered bedrock where present. Alluvium thickness varies between 2-6 meters near the mine access road, but is uncertain up-valley.

Average transmissivity = 67 m<sup>2</sup>/day (conductivity =  $2x10^{-4}$  m/s); average storativity = 0.014.

Unit 3 – Bedrock: phyllitic bedrock is relatively minor aquifer unit only. Intermittent discrete fracturing present but with no known preferred orientation. Average hydraulic conductivity =  $4.3 \times 10^{-8}$  m/s with higher zones (up to  $1.1 \times 10^{-7}$  m/s)

Review of available hydraulic gradient and stream flow data suggests that the Faro Creek channel gains significant flow from groundwater along its length. While the exact location of gaining or losing reaches is unknown, data suggests this occurs in the upper reaches, but below X23. Below the mine access road, Faro Creek does not appear to gain significantly from upwelling groundwater emanating from bedrock.

## 3.6 Water Quality

Tables 7 and 8 summarize selected groundwater quality data collected in 2005 from existing and newly installed monitoring wells in the ETA area. Complete water quality data is included in Appendix E. The following conclusions can be drawn about the groundwater quality in the ETA area:

- All groundwater samples exceed the CCME limit of 0.05 mg/l for Zinc
- Groundwater quality in the two up-gradient wells (P96-6A&B) showed significant variations over time suggesting variable contributions of (more dilute) surface water recharge;
- Concentrations of sulphate, zinc, and in particular iron were generally higher in alluvial wells located in the ETA area compared to those wells located up-gradient; furthermore, the highest concentrations of sulphate, zinc and iron were observed in the single well screened in tailings (SRK04-03B); these observations suggest that the ETA tailings represent a significant source of contaminant loading (in particular iron) to the Faro Creek seepage;
- The sample from the bedrock well, SRK05-ETA-BR2, had the lowest contaminant concentrations of all wells in the ETA. This information suggests that bedrock groundwater is significantly less affected by ARD seepage from the Faro mine site than groundwater in the alluvial aquifer;
- Contaminant concentrations in the pumping well decreased slightly with time during the pumping test. This result may indicate a declining influence of very poor quality waters leaking from overlying tailings.

ID	Date	Lab pH	Lab Conductivity (µS/cm)	SO4 (mg/L)	Zn (mg/L)	Fe (mg/L)
		November	2005 Sampling			
SRK05-ETA-BR1	11/2005	5.42	9750	9250	681	3100
SRK05-ETA-BR2	11/2005	6.8	2040	1200	7.41	22.5
	Octo	ber 2005 SF	RK04-04 Pumping	Test		
SRK04-4	10/2/2005	5.44	7630	7370	438	1950
	0	ctober 2005	Sampling (Pre-Te	st)		
P96-8A	9/10/2005	6.50	6370	5040	604	0.061
P96-8B	9/10/2005	6.35	6620	4980	368	9.85
		May 20	05 Sampling			
SRK04- 04-04	5/5/2005	5.23	n/a	7080	350	1630
SRK04-03A	5/5/2005	5.87	n/a	5480	233	693
SRK04-03B	5/5/2005	3.72	n/a	16700	749	6610
P96- 8A	5/3/2005	6.76	197	71.2	1.67	0.064
P96- 8B	5/3/2005	7.01	5540	4520	173	0.22

#### Table 7: Monitoring Well Water Quality

ID	Date	Lab pH	Lab Conductivity (µS/cm)	SO4 (mg/L)	Zn (mg/L)	Fe (mg/L)			
	October 2005 SRK04-04 Pumping Test								
SRK04- 04-1HR	10/2/2005	4.84	8390	8100	461	2380			
SRK04- 04-10HR	10/2/2005	5.34	7780	7460	447	2020			
SRK04- 04-24HR	10/3/2005	5.39	7610	7460	444	1950			
SRK04- 04-36HR*	10/3/2005	5.44	7630	7370	438	1950			

Table 8: Pumping Well Water Quality during Test

\*NOTE: SRK04-04-36HR is a duplicate sample of SRK04-04-24HR

Comparison of groundwater quality with surface water quality (Table 3) indicates that groundwater from monitoring wells completed in the aquifer unit is slightly more impacted than surface water quality, with the notable exception of SRK05-ETA-BR2, the bedrock monitoring well, which has significantly better water quality than that of the overlying alluvial aquifer or Faro Creek.

# 4 Assessment of Contaminant Sources and Loading

## 4.1 Contaminant Sources

### 4.1.1 Waste Rock Seepage

Waste rock seepage represents a significant contaminant source for ETA groundwater and surface water (Faro Creek). Two monitoring wells installed close to X23 in 1996 (P96-8A and 8B) show high concentrations of sulphate, zinc and other metals. Sulphate and zinc concentrations show similar trends to surface water quality at X23. Sulphate has increased from about 2,000 mg/L in 1996 to about 4,000 mg/L in 2003. Zinc has increased from about 2 mg/L to greater than 100 mg/L in the same 8-year time span.

## 4.1.2 ETA Tailings

Groundwater quality results from ETA monitoring wells indicate that the tailings also represent a significant source of contaminants (in particular iron) to the Faro Creek seepage. ETA wells have significantly higher concentrations than upstream wells and a well completed in the tailings (SRK04-3B) had the highest observed concentrations and lowest pH. Concentrations were also highest in the pumping well at early times, potentially when tailings leakage likely represented a larger proportion of discharged water.

## 4.1.3 Other Sources

The waste rock seepage at X23 represents only some of the seepage that enters the ETA. There are also intermittent flows from the waste rock north of X23 and from the mill area. Site staff also report that other wastes, such as the concentrate removed during thickener cleaning, have been introduced into the ETA. These wastes presumably remain mixed in with the tailings. No independent sampling of these sources has been completed; therefore, their contribution to contaminant loadings is not distinguishable from the contribution by tailings.

# 4.2 Contaminant Loading in Groundwater and Surface Water

## 4.2.1 Faro Creek and Alluvial Groundwater

Table 9 summarizes the observed surface water flows and calculated loads of sulphate, total zinc and total iron at the four sampling locations along the Faro Creek channel. The following conclusions can be drawn with respect to contaminant loading in surface water upstream and downstream of the ETA area:

• The contaminant load associated with surface runoff entering the ETA area (i.e. waste rock dump seepage reporting to X23) is generally much smaller (<25%) than the combined

contaminant load discharging from the ETA area (in surface runoff and groundwater discharge combined);

- The contaminant load associated with surface runoff from the ETA area varies significantly (in relative and absolute terms), due to significant seasonal variations in both flow and contaminant concentrations (see table 3);
- Groundwater discharge along the seepage face downstream of the ETA area (immediately below the access road) represents the primary source of contaminant loading during winter baseflow and a significant source of loading during the remainder of the year;
- The total sulphate and zinc load discharging from the ETA area (at FCS-4) varies with flow conditions; for example, zinc loading ranged from 70 t/yr during the (wet) fall to 36 t/yr during the winter baseflow; the total iron load at FCS-4 remained surprisingly constant over time (~170 t/yr);
- During the October 2005 survey, the combined sulphate and zinc loads from FCS-2 and FCS-3 agreed very well with the observed total loads at FCS-4. The load estimates for the other surveys are considered too uncertain (in particular for FCS-2 due to freezing of the culvert) to allow similar mass balance calculations.

Date	Station	Flow L/s	S04 Load t/yr	Zn Load t/yr	Fe Load t/yr
5	FCS-1	1.3	206	12.1	3.6
or-0	FCS-2	frozen	0	0	0
0-Al	FCS-3	4.8	840	46.8	183
1	FCS-4	6.6	868	36.2	167
5	FCS-1	1.15	225	17.3	5.7
ct-0	FCS-2	3.4	666	49.2	6.6
0-8	FCS-3	3.1	642	21.7	109
<del>, -</del>	FCS-4	7.2	1306	70.4	176
15	FCS-1	1.01	203	14.6	4.2
00	FCS-2	<mark>1.0 to 2.0</mark>	196 – 392	14 – 29	1.9 – 3.9
1-Nc	FCS-3	3.4	800	39.8	192
5	FCS-4	5.4	955	45.3	160
15	FCS-1	0.63	118	10.3	3.6
-De	FCS-2	frozen	0	0	0
9-D	FCS-3	2.3	510	31.2	144
1	FCS-4	4.2	734	36.8	162

Table 9: ETA Surface Water and Seepage Loading, October 2004

Estimated range

#### 4.2.2 Bedrock Groundwater

Table 10 presents loading estimates for the bedrock system, based on available data from SRK05-ETA-BR2. These estimates assume: the hydraulic gradient is the same as the overlying

alluvial aquifer under static conditions (~0.064); a width of 200m (~50m wider than the interpreted alluvial system); and a thickness in bedrock of 20m (total area =  $4000m^2$ ). The maximum and minimum hydraulic conductivities from packer testing were used for calculations.

The estimated contaminant loads associated with groundwater flow in bedrock are 3-4 orders of magnitude lower than the contaminant loads observed in the alluvial aquifer (at FCS-3). The much lower contaminant loads are a result of the orders-of-magnitude lower bedrock permeability plus significantly lower contaminant concentrations.

	K (m/s) Flux Concentration (m3/s) (mg/L)		SO4 Concentration (mg/L)	Zn Load (tonnes/yr)	SO4 Load (tonnes/yr)	
1-min	1.8x10 <sup>-8</sup>	4.6x10 <sup>-3</sup>	7.41	1200	0.0011	0.1714
2-max	1.1x10 <sup>-7</sup>	2.8x10 <sup>-2</sup>	7.41	1200	0.0065	1.048

Table 10: Groundwater Loading in Bedrock

# 4.3 Contaminant Loading to Rose Creek Valley

The calculated loads at FCS4 are interpreted to represent most of the total load to Rose Creek Valley from the ETA. As discussed above, upstream loads from FCS2 and FCS3 balance very well with observed (combined) load at FCS4 and groundwater moving in the deeper bedrock is estimated to carry only a very small, even negligible load.

# 5 Conceptual Design of Seepage Interception System

The primary sources of contamination in the ETA are the up-gradient waste rock dumps, the mill area and the ETA tailings themselves. While removal of the ETA tailings is a feasible option for partial source control, the waste rock dumps and mill site will not likely be removed as part of the mine closure, though options to reduce infiltration through these areas and, subsequently, load from them, are being considered. Consequently, collection systems to intercept the main flow of contaminated groundwater will be required for an indefinite period of time.

This extended period of time will allow any capture system to be refined or upgraded. Various technologies installed in phases, as necessary, maximize the overall capture efficiency of the total system.

# 5.1 Assessment of Previous Design Concepts

SRK (2004) proposed a SIS consisting of a fence of pumping wells up-gradient of the mine access road for groundwater control, in combination with a sump for collection of Faro Creek surface water flows. As part of the 2005 investigation, further work was completed on assessing the viability of this option.

Utilising the additional data collected during the 2005 investigation, a scoping level numerical model was created for the ETA to determine the number of pumping wells that would be required for the recommended SIS. The numerical model was a 2-D, finite element model using the code Feflow, produced by WASY (WASY, 2006). An initial model was calibrated to steady-state static conditions and then run transiently for comparison with results of the 24-hour pumping test. This initial model was used to assess longer term interception using different combinations and positioning of pumping wells (see Appendix F for details).

The numerical modeling suggested that pumping wells alone would not provide adequate capture of contaminated groundwater passing through the ETA. A "stagnation point" or line beyond which the pumping wells had either no, or only a minimal, effect developed down-gradient of the fence of pumping wells, leading to bypass of some contaminated groundwater. The primary factor affecting the ability of the pumping wells to capture all groundwater was related to the limited available drawdown and elevation difference between the location of the pumping wells and the down-gradient seepage face. Complete capture could only be achieved by allowing drawdown in the pumping wells to go below the elevation of the seepage face, an operational measure that is not practical considering the local geology and the potential for "well fouling", i.e. blinding of the well screen with iron precipitates, due to aeration of the well screen when water levels are drawn down significantly.

## 5.2 Recommended Approach

The approach recommended herein has three components:

- 1. Construction of a "Primary Seepage Interception System" that will be as effective as possible given the current understanding of the site conditions;
- 2. Installation of a monitoring system to assess performance of the Primary SIS; and
- 3. A series of contingency measures to be implemented according to a well-defined Adaptive Management Program.

## 5.2.1 Primary SIS

The primary seepage interception system will have groundwater and surface water collection components, which will be constructed after removal of the ETA tailings. Conceptual layout of the SIS is shown on Figure 11. Figure 12 shows two longitudinal cross-sections through the ETA including the proposed SIS.

Currently, a large volume of tailings are present (scoping estimate of >64,000 m<sup>3</sup>) and it is assumed that tailings will be removed by hydraulic mining and or excavation (truck and shovel) with some additional materials cleanup required (~10,000-20,000 m<sup>3</sup>). Once the tailings have been removed, a berm will be built out approximately 20 meters from the edge of the mine access road, into the ETA area, that would act as both a platform for construction of the interception system and support for the flood-retention structure.

#### **Groundwater Interception**

Groundwater will be intercepted using a combination of slurry cut-off wall and pumping wells upstream of the mine access road, constructed from the top of the berm. The slurry wall would be emplaced using trenching equipment and a soil-bentonite slurry, which would be keyed into weathered bedrock. Pumping wells would be installed up gradient of the cut-off wall. The pumping wells would be screened in the overburden soils and the underlying weathered bedrock. In areas where the overburden soils along the proposed alignment are comprised of lower permeability till material (in particular, close to the valley sides), a permeable trench would be installed down to bedrock using the same trenching equipment as the slurry-wall. In those areas, the pumping wells would be screened directly into this permeable trench and underlying weathered bedrock. In the central portions of the valley, near SRK04-4, where permeable alluvium has been identified during drilling, the permeable trench would likely not be required, but could be installed if deemed appropriate by field engineers.

Based on the results of our initial loading calculations, direct interception of groundwater flow in deeper bedrock (below weathered bedrock) using a grout curtain plus pumping wells screened in deeper bedrock would not be included in the primary SIS. However, these system components were included as optional components of the adaptive management plan. This is described further in Sections 5.2.2 and 5.2.3.

Based on the available hydraulic information, a well spacing of approximately 50m (between two neighbouring pumping wells) is proposed, requiring a total of 5 pumping wells (Figure 11). This spacing should provide some redundancy in the system to allow maintenance of individual pumping wells without complete system shutdown. The screening intervals of the pumping wells would be selected sufficiently below the water table to minimize aeration of the well screen during active pumping. Pumping wells would be outfitted with automatic controllers, including water level recorders and flow meters, to maintain a positive hydraulic gradient toward the up gradient side of the barrier and to provide information regarding total load captured. The system would be equipped with an alarm system to alert the operator if the system was malfunctioning (e.g. pump breakdown).

The proposed combination of a hydraulic barrier with a fence of pumping wells located up-gradient of the barrier provides a very high collection efficiency. The hydraulic barrier improves the hydraulics of the capture system (by essentially ponding the groundwater up-gradient for collection) and also provides a secondary containment in the case of temporary shut-down of the pumping system (e.g. due to power failure). One of the authors of this report recently designed, constructed and operated a similar hydraulic barrier system at a development site in North Vancouver, BC to prevent hydrocarbon contamination. This hydraulic barrier system was operated successfully for over three years with no detectable by-pass of contaminated groundwater (Wels, 2002).

Surface water collection in the ETA would have three components: (i) flood detention storage, (ii) collection of contaminated surface runoff (Faro Creek seepage), and (iii) bypass of clean run-off from surrounding areas.

#### **Flood Detention Storage**

Scoping level estimates indicate that removing the tailings and associated cleanup could provide 100,000 m<sup>3</sup> of storage behind the mine access road. This volume is estimated to be adequate to store pre-closure flood flows up to the 100-yr event and most post-closure flows (Appendix C). Only the estimated "long-duration" post-closure events, such as a 1-week 100-yr event (average 278 l/s), would exceed this storage. Flows greater than the 100-yr event, or long duration events that exceed storage, would be passed downstream via an emergency spillway.

Consequently, it is proposed that all pre-closure storm flows be stored in the ETA detention pond until they can be pumped to a water treatment plant or the Faro Pit. The interim water quality of the runoff flows post-closure (after cover construction) is unknown, but it is conservatively assumed that runoff water quality will initially remain poor and will continue to require treatment and storage in the ETA. Once runoff has sufficiently improved, it could be allowed to directly discharge to Faro Creek canyon (i.e. no storage). The mine access road, berm and cut-off wall are envisioned to hold back water in the ETA detention area. No assessment has been made regarding any geotechnical upgrades to the mine access road that may be required. The suitability of this structure will require assessment in the detailed design phase. It should be noted that significant storage of flood water in the ETA over extended periods of time may temporarily compromise the performance of the groundwater SIS. The ponding of surface water would tend to increase the groundwater level on the up gradient side of the barrier for the period of ponding. In order to prevent this upstream mounding, the following control measures would be implemented:

- Provide adequate pumping capacity to reduce the pond height relatively quickly after a significant storm event; and/or
- Line the upstream face of the berm and the foot print area of the detention pond within a narrow corridor (say 20m distance) upstream of the berm using a synthetic liner (see Figures 11 and 12).

The final design of a flood-storage structure in the ETA will require integration with overall site water treatment plans. If the structure is to operate only as a short-term, emergency retention structure, with water pumped as quickly as possible to the treatment facility, the total volume of treatable water would increase substantially. Alternatively, if the ETA were designed such to provide medium-term flood water storage, the additional water could be directed to the treatment plant at a more constant rate. A third option would be to utilize a long term storage pond elsewhere, either in the Faro Pit or constructed elsewhere. Decision on how storage should be accommodated will have to be addressed after the final closure plan has been determined and will then include final sizing of pumps and pipelines.

### Faro Creek Seepage Collection

Surface runoff in the historic Faro Creek channel (primarily seepage from the Faro waste rock dumps) will likely remain contaminated for the foreseeable future. Faro Creek water will be realigned towards a sump in the ETA, once tailings have been removed (see Figures 11 and 12). Water will be directed from the sump to the main pumping well discharge line, from which it will be directed to the water treatment plant.

#### **Run-off Bypass**

Currently, there is a component of the ETA catchment, located southeast of the ETA, which likely generates relatively clean run-off. A shallow ditch will be constructed along the southeast margin of the ETA to divert this water from entering the ETA. Water in this ditch would be directed past the mine access road and Faro Creek Canyon to a reasonable discharge location. Under closure conditions, when covers have been installed over waste rock dumps, the drainage ditch would be extended to allow capture of additional clean water prior to it entering the ETA area.

## 5.2.2 Initial Monitoring System

Both groundwater and surface water monitoring will continue for the duration of interception. Monitoring locations would be focused in four general areas: (i) up-gradient of the cut-off wall, (ii) immediately down gradient of the cut-off wall, (iii) in the area of the seepage face below the access road and (iv) at the mouth of the Faro Creek Canyon. Monitoring station locations are shown on Figure 11. Combined with surface water monitoring stations, the multiple tiers of groundwater monitoring wells would allow for the assessment of SIS performance by observing different scale flow systems, from small scale in close proximity to the cut-off wall to larger scale between the ETA and the mouth of the Faro Creek Canyon. Monitoring stations would be used to provide assessment of system performance via four general parameters:

- 1. Groundwater gradients;
- 2. Groundwater quality;
- 3. Surface water discharge rate;
- 4. Surface water load.

All groundwater monitoring wells would be nested or multi-level, with zones in overburden, weathered bedrock and deeper bedrock (Figure 12). Packer testing would be conducted during installation to provide additional information regarding bedrock permeability. Monitoring wells located between the pumping wells up-gradient of the cut-off wall and immediately down gradient of the cut-off wall would be monitored to assess the hydraulic gradients created by the SIS. Vertical gradients in monitoring wells located between pumping wells, when combined with water quality data, would allow detection of underflow or other forms of cut-off wall bypass.

Note that downstream-directed gradients across the cut-off wall may still occur due to elevation differences and dewatering of the down gradient side, even if all groundwater was captured in the SIS. Hence, additional monitoring of groundwater quality in these wells (plus surface water monitoring at the seepage face, see below) will be required to assess system performance.

Three monitoring wells will be installed in bedrock in the upper Faro Creek Canyon, (i.e. immediately down gradient of the mine access road - see Figure 11 for location). During installation, packer tests will be carried out in the bedrock to assess the bedrock permeability and to assist with selection of the screening intervals. These monitoring wells will be used to monitor groundwater gradients and groundwater quality in downstream bedrock in response to system operation. Information from these monitoring wells, in conjunction with information from the bedrock wells completed immediately down gradient of the SIS, would be used to assess the potential for seepage by-pass in bedrock (see section 5.2.3).

A groundwater monitoring station would also be located at the mouth of the Faro Creek Canyon. This station would be used primarily to monitor absolute and relative water quality. Hydraulic gradient information would be obtained that could be used to further characterise the larger-scale flow system, but would not be an integral part of the monitoring system. While it may take considerable time for water quality at this location to improve, due to its proximity to the tailings impoundment and its current poor water quality, water quality is unlikely to worsen. Surface water monitoring would include discharge rate and water quality at three of the four pre-SIS monitoring stations: FCS-1, FCS-3 and FCS-4. FCS-2, currently located on the downstream end of the culvert passing under the mine access road, would be deactivated once this culvert is blocked by the berm and cut-off wall. If the system is working properly, there should be no more discharge of seepage along the current seepage face monitored at FCS-3, except perhaps for some minor flow of "clean" groundwater from the surrounding areas. The presence of highly contaminated water at this location would indicate cut-off wall bypass. Monitoring at FCS-4 will provide a final check on seepage collection (groundwater and surface water combined). Assessment of load at FCS-4 will improve the understanding of contributions from the ETA area itself, which should not contribute significantly if the SIS is working properly, and from the relatively small catchment downstream of the ETA. Monitoring at FCS-1 (X23) would continue to improve understanding of the ETA hydrology and would detect potential improvement in water quality assuming covers are installed on the waste rock dumps. This information would be used to determine the timing for extension of the shallow runoff collection ditches to bypass relatively clean water around the ETA.

## 5.2.3 Adaptive Management Program

Performance of the Primary SIS would be assessed using surface water and groundwater data from the monitoring system. If surface water or groundwater parameters reach or surpass monitoring triggers, contingency measures would be implemented.

The monitoring triggers and contingency measures can be grouped into three main areas:

- 1. Area adjacent to the cut-off wall,
- 2. Down gradient of the cut-off wall (at seepage face), and
- 3. Mouth of Faro Creek Canyon.

In all areas, the first action upon reaching a monitoring trigger would be an investigation into the cause and determination of appropriate remedial steps. If trigger values are reached after these initial remedial action steps, then additional investigations or remedial actions would be taken as described below.

#### Area adjacent to cut-off wall

In this area, the monitoring system would consist primarily of multi-level groundwater monitoring wells, located both immediately up gradient and down gradient of the wall itself. Groundwater monitoring would include two components:

- 1. Hydraulic gradients
- 2. Groundwater concentrations

Hydraulic gradients would be monitored primarily between pumping wells and monitoring wells located on the up gradient side of the cut-off wall. Lack of adequate drawdown in water levels at these monitoring locations would suggest pump inefficiency or improper spacing. Remedial actions

would include an investigation of the pumping system and repair or installation of additional pumping wells, as necessary.

In the event of a significant increase in groundwater contaminant concentrations from baseline levels, two responses are possible. If the initial investigation suggested edge bypass, additional pumping wells would be installed closer to the ends of the cut-off wall. If the investigation suggested that underflow through the bedrock system was occurring, options for grouting and/or pumping of the bedrock would be investigated.

Bedrock grouting could be completed along either of two alignments: immediately below the cut-off wall, or down gradient of the mine access road, where Faro Creek Canyon begins to narrow. At this time, there is insufficient bedrock data to provide a recommended option. Data should be collected in both locations during installation of the monitoring system. If the cut-off wall alignment were chosen, monitoring would continue as before. If the down gradient alignment were chosen, the interim collection sump would likely require upgrading to allow continued operation, but monitoring would continue as prior to the trigger, as described in the following.

#### Area down gradient of the cut-off wall (at seepage face)

Surface water and groundwater would be monitored in the vicinity of the seepage face in the Upper Faro Creek Canyon (near monitoring station FCS-3). The flow rate and water quality of seepage at this location (FCS-3) would be monitored as part of surface monitoring. If the SIS is working properly, there should be no surface flow at this monitoring station, besides local inputs from precipitation recharge. If contaminated seepage was identified, a remedial investigation would be conducted. Remedial actions could include, as necessary, repair or improvement of the pumping system or re-activation of the interim collection system. If seepage could be linked to underflow at the cut-off wall, an investigation into bedrock grouting options would be completed.

Groundwater monitoring would focus on water quality. If contaminant concentrations were observed to increase from baseline levels, an investigation would be completed. If surface seepage were not observed, changes in contaminant concentrations would likely be related to underflow. In this case, options for bedrock grouting would be investigated.

#### Area at mouth of Faro Creek Canyon

Both surface water and groundwater would be monitored at the mouth of the Faro Creek Canyon (at monitoring station FCS-4). Surface water monitoring would include water quality, discharge and load. If the SIS is working properly, there should be only minimal surface flow at this location, particularly during baseflow conditions. Surface water monitoring would include continued operation of the FCS-4 weir and datalogger. The trigger at this location would be the presence of contaminated seepage. If contaminated seepage were observed, an investigation of the upstream SIS would be completed. Remedial actions for upstream areas would be implemented as appropriate. If contaminated seepage were determined to be the result of upwelling groundwater, pumping wells

could be installed in this area. Further measures would include bedrock grouting options, if appropriate, and/or secondary containment structures.

Groundwater quality would be monitored and compared to baseline values. The trigger for investigation and remedial action would be an increase above baseline conditions. Due to its location on the edge of the tailings impoundment and the current presence of contamination, improvements in water quality here are unlikely to occur for a significant period of time. In fact, it may even be possible that contaminant concentrations in groundwater (in particular in bedrock) may even increase before they decrease, even if the SIS is working as intended. A detailed investigation into the likely causes for any increase in contaminant concentrations at this location (including a comparison with up gradient system monitoring), and an assessment of the associated loading to the Rose Creek valley aquifer would be required before additional remedial action would be taken.

If such an investigation concluded that contamination could not be effectively intercepted at the up gradient SIS and that the contaminant load represented a risk to the downstream environment, a secondary groundwater interception system (e.g. consisting of a cut-off wall and/or pumping wells) would be installed at the mouth of Faro Creek canyon.

As a final contingency, if initial remedial actions at any or all monitoring areas do not provide adequate capture to protect the downstream aquatic environment, a secondary containment structure or capture system could be developed in or below the Faro Creek canyon.

## 5.3 Further Work

Additional design work should be completed on flood detention structure options at the ETA during the detailed design phase. The ability to design appropriate control structures for estimated flood discharge is limited by the relatively short and intermittent record of flow for the historic Faro Creek channel at FCS-1 (X23), uncertainty regarding actual volume of tailings and related materials that will be removed from the ETA and uncertainty about the final design and effectiveness of waste rock covers. Each of these items should be further assessed prior to final design.

Monitoring of Faro Creek flows should continue and improvements of flood hydrology completed. Uncertainty regarding volumes of water that could potentially be bypassed during a large flood event, and the consequent loading, will be decreased only by improvement in flood forecasting. As part of flood forecasting, an assessment of risks associated with bypass of excess flood waters should be completed.

Regardless of these uncertainties, the mine access road is likely to be needed to perform as a control structure to some degree. An assessment of what the access road may be used as, or what level of upgrading would be necessary, should be undertaken by a Professional Geotechnical Engineer.

Prior to additional work in the ETA area, a detailed ground survey should be conducted of the historic Faro Creek channel, focusing on the canyon and seepage face areas downstream of the mine access road.

Routine groundwater and surface water monitoring should continue to further develop baseline conditions for use with the Adaptive Management Program.

The interim collection sump will be installed in 2006/2007 and operated during summer months. Contaminant capture monitoring at this system should be used to improve the understanding of system dynamics and loading estimates. Water quality and discharge monitoring should occur at a location down-gradient of the interim collection sump to assess capture efficiency.

This report, **"2005/06 Task 20e - 2005 Seepage Investigation at the Emergency Tailings Area"**, has been prepared by SRK Consulting (Canada) Inc.

Daniel Mackie, M.Sc.

Dr. Christoph Wels, Ph. D., M.Sc.

Quinn Jordan-Knox, M.Sc.

**Reviewed by** 

Cam Scott, P.Eng.

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SRK, 2006 Preliminary Seepage Collection Options – Faro and Grum Waste Rock Dumps; prepared for Deloitte and Touche, Interim Receiver for Anvil Range Mining.

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Figures


























Appendix A Borehole Drill Logs

	B	ORE	Er HOL	Co ngine LE	nsulting ers and Scientists LOG STRATIGE	PROJECT: LOCATION: ETA Area FILE No: FARO (1Cl BORING DATE: 2005-08 DIP: AZIMU COORDINATES: 69138	D00 3-1: 1TH	03.73) 5 To 1: .00 N	5829	53.00	e da	BOREHOLE: ETA-05-01 PAGE: 1 OF 2 DRILL TYPE: DRILL: CASING: TUM:	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
DEPTH - ft	DEPTH - m	DETAIL & WATE LEVEL -	S T NOILY A T	DEPTH - M	DESCRI Natural ground sur	IPTION S		TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
X/00 KEFERRIVCE MM / E-KM-15 Above roduction and the manual structure a	- 1	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{c} 1104 \\ \hline 0 \hline \hline 0 \\ \hline 0 \\ \hline 0 \hline \hline 0 \hline \hline 0 \\ \hline 0 \hline \hline \hline 0 $	<u>1.59</u> 90	Gravel, silty, sandy.	nes.		ò				0 Light brown, slightly moist sand mixed with dark gray discrete pods. Surface has subangular gravel. 0.152000 Dark grey to black, fresh sulphides (sand), no oxidation, moist (sl.). 0.305000 Light yellowish brown gravelly sand, anulgar gravel, slightly moist, gravel up to 2cm. 0.610000 Dark grey Sx sand, some lenses up to 20cm with more fines, fresh Py (sand). Bottom of unit has more Py sand, less fines. Slightly moist. 2.134000 Brownish orange, gravelly sand. Silt, moist to very moist. Angular gravel (20%) up to 3cm. No fresh Sx observed. 2.896000 Mottled white/orange/grey/brown zone, dry to slightly moist, silty sandy gravel (ang., up to full core diam. (4")). Lenses full of visible Py up to 20cm. Dry to sl. moist.	2	

	B	/ SRI oreho	K Ca Engir DLE	<b>DISULTING</b> meers and Scientists <b>LOG</b>	PROJECT: LOCATION: ETA Are FILE No: FARO ( BORING DATE: 2008 DIP: AZI COORDINATES: 69	a (1CE 5-08 <b>IMU</b> 1388	0003.73) -15 T TH: 51.00 N	<b>O</b> 5829	53.00	E DAT	BOREHOLE: ETA-05-01 PAGE: 2 OF 2 DRILL TYPE: DRILL: CASING: SAMPLE CONDITION Remoulded Undisturbed Lost CASING: TUM:	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
		WELL	E	STRATIG	RAPHY			SAMI	PLES	;			
DEPTH - ft	DEPTH - m	& WATER LEVEL - m	ELEVATION - n DEPTH - m	DESCR	IPTION	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
25	- 7	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	<u>1097.18</u> 7.32	Sand, gravelly, silty.			8				7.315000 Original ground. Peat layer ~ 5cm, coarse fibrous. Smells like diesel. 7.620000 Dark chl green to olive green gravelly silty sand, moist to wet, mottled rusty zone at 8.53m. Smells like diesel from 7.62-8.53m.	7	
COB REFERENCE MATERIALS/geotec.log/tempates/og/tempates/og/tempates/ 36	) - - - - - - - - - - - - - - - - - - -		9.14	END OF BOREHOLI	E	1.12						10	

	B	ore	R HC	CCC Engin	<b>Stratic</b>	PROJECT: LOCATION: ETA Area FILE No: FARO (10 BORING DATE: 2005-0 DIP: AZIMU COORDINATES: 69130	2D) )8-' <b>UT</b>	003.73) 15 T H: 9.00 N	O 5829	993.00	DE DA	BOREHOLE: ETA-05-02 PAGE: 1 OF 2 DRILL TYPE: DRILL: CASING:	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
DEPTH - ft	DEPTH - m	WELL DETAIL & WATE LEVEL -	S ER - m	ELEVATION - M DEPTH - M	DESCRI Natural ground sur			TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
X106 REFERENCE MAILERING Supported to the second se	- 1	P         P         P           P         P         P		<u>1100.27</u> 3.96 <u>1099.35</u> 4.88	Tailings Sandy gravel			1 2 3 4 5 6 7				0       Light brown gravelly sand, dry angular - subangular gravel; ETA         surface material       0.30000         Yellowish gray moist tailings, 95% fresh, sandy silt, vertical black and orange laminations         1.200000         Yellowish brown sand (tailings), very slightly moist, 5% loosely cemented, no silt.         3.000000         Dark gray tailings sand, little to no silt, moist. WL appeared to be at 4.0m.         4.300000         Mixed tailings and organics.         4.500000         Sandy gravel, light brown, subangular - subrounded, moist.         5.500000         Olive green till, silty sandy gravel, moist, diesel smell.	22	

	B	/= SRI oreh(	<b>É Co</b> Engine <b>DLE</b>	nsulting eers and Scieniists LOG	PROJECT: LOCATION: ETA Area FILE No: FARO (1 BORING DATE: 2005- DIP: AZIM COORDINATES: 691	1 1CD -08- MUT .384	0003.73) 15 T F <b>H:</b> 9.00 N	<b>O</b> 5829	93.00	E DAI	BOREHOLE: ETA-05-02 PAGE: 2 OF 2 DRILL TYPE: DRILL: CASING: CMS: BOREHOLE: ETA-05-02 SAMPLE CONDITION Constant Remoulded Undisturbed Lost Rock core CASING:	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
		WELL	e 1	STRATIGE	RAPHY		5	SAMF	PLES				
DEPTH - ft	DEPTH - m	& WATER LEVEL - m	ELEVATION - n DEPTH - m	DESCRI		SYMBUL	TYPE AND NUMBER	CONDITION	<b>RECOVERY %</b>	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
-11-03 13:20hrs	. 7		1096.63				8 9				6.500000 Coarse sand with gravel, no silt. 6.700000 Wet till, silty sandy gravel, light orange, brown.	7.	
mp-Lab.sty PLOINED: 2000	5		7.60	END OF BOREHOLE	<u>-</u>							8	
Vog/PMWell-Strat-RQD-Sa	0											9	
RIAL Styreotec. log vemplates	- 10 5											10	
X:106 REFERENCE MATE	- 11 								Address - A			11	

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B	SOREHO	K Cc Engin DLE	onsulting eers and Scientists LOG	PROJECT: LOCATION: ETA A FILE No: FARO BORING DATE: 20 DIP: A COORDINATES: 6	rea (1CI 05-08 <b>ZIMU</b> 59138	0003.73) -15 T TH: 08.00 N	O 5829	975.00	) E DA	BOREHOLE: ETA-05-03 PAGE: 1 OF 2 DRILL TYPE: DRILL: CASING: TUM: BOREHOLE: ETA-05-03 Remoulded Undisturbed Lost Rock core	TYP DC GS SS	E OF SAMPLER Diamond core barrel Grab sample Split spoon
DEPTH - ft DEPTH - m	WELL DETAILS & WATER LEVEL - m	000 DEPTH - m	STRATIGF DESCRI Natural ground sur	RAPHY PTION face	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$		Tailings			1 2 3 4 4				0 Dry gravel, subangular, 40% sand, orange. 0.400000 Yellowish brown coarse tailings, <5% angular gravel mixed in, dry to slightly moist. 1.500000 Dark gray fresh tailings, silt with 5% sand, wet. 3.000000 Yellowish gray tailings sand, fresh, wet, very little silt. 5.250000 Wet silt (tailings-fresh) as from 1.5-3m.	2	

	B	/= SRI OREHO	Engine	<b>Insulting</b> eers and Scientists LOG	PROJECT: LOCATION: ETA A FILE No: FARO BORING DATE: 20 DIP: A COORDINATES:	Area (1CI 005-08 <b>AZIMU</b> 69138	D003.73) -15 T TH: 08.00 N	O 5829	75.00	E DAT	BOREHOLE: ETA-05-03 PAGE: 2 OF 2 DRILL TYPE: DRILL: CASING: UM: BOREHOLE: ETA-05-03 Remoulded Undisturbed Lost CASING: UM:	TYPE OF SAMPLER DC Diamond core barrel GS Grab sample SS Split spoon
		WELL DETAILS	ε	STRATIG				SAIVII	LES	)		
DEPTH - ft	DEPTH - m	& WATER LEVEL - m	ELEVATION - DEPTH - m	DESCR	IPTION	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	E H LABORATORY and IN SITU TESTS G
25	7	Por         Por         Por         Por           Por         Por         Por         Por         Por	-6.71 6.71 -7.92 7.92 -8.53 8.53 8.53 -9.00	Sand/ gravel Till Bedrock			6 7 8 9				6.750000 Fibrous peat. 6.780000 Coarse sandy gravel with silt; light brown, wet. Angular-subrounded gravel. 7.500000 Dark chl green till, gravelly sandy silt, moist, angular gravel chips to 3cm diam. 8.400000 Looks like bedrock. Powder rock flour and chips that look like quartzite - siliceous.	8
20 Kulo REFERENCE MAI LEVIALS George Log tempares vegr-minettrant	10 11				_							10

	B	<b>SRI</b> OREHO	CCO Engine DLE	eers and Scientists	PROJECT: LOCATION: ETA Area FILE No: FARO (10 BORING DATE: 2005-0 DIP: AZIM COORDINATES: 6913	CD( )8-1 IUT	003.73) 15 <b>T</b> i T <b>i:</b> 8.00 N	O 5830	45.00	E DA	BOREHOLE: ETA-05-04 PAGE: 1 OF 2 DRILL TYPE: DRILL: CASING: TUM:	SAMPLE CONDITION Remoulded Undisturbed Lost Rock core	TYP DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
DEPTH - ft	DEPTH - m	WELL DETAILS & WATER LEVEL - m	M - M DEPTH - M 1104.51	STRATIGF DESCRI Natural ground sur	RAPHY		TYPE AND NUMBER		RECOVERY %	RQD %	SAMPLE DESCRIPTION	Ι	DEPTH - m	LABORATORY and IN SITU TESTS
- 5	· · · · · · · · · · · · · · · · · · ·	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	0.00	Tailings			2				0 Pale orangeytan silty gravel; streaks of yellow 0.200000 Fine to medium sand tailings. Transitions from through brownish orange, yellowish. Gray, bro depth. Moist.	and orangey silt. n orange at top wnish yellow with		
	- 2	$\begin{array}{cccccccccccccccccccccccccccccccccccc$					3				Uniform yellowish dark gray fresh medium sar	nd tailings, moist.	2	
- 15	- 4	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	1099.94 4.57	Tailings/OVB contact	t		4				4.500000 Run was lost. Driller says cobble layer preven Transition from tailings to overburden occurred Driller thinks overburden starts at 6m. Based o	ted any recovery. d over this interval. on drill response.	4 5	
	-	DB .DD DB .DD DB DD DD .DD DD DD DD .DD DD .DD DD .DD											-	

		B	SRF OREHC	Engin	<b>DISULTING</b> eers and Scientists LOG	PROJECT: LOCATION: ETA Area FILE No: FARO (10 BORING DATE: 2005-0 DIP: AZIMI COORDINATES: 69133	DDC 08-1 01 858	003.73) 15 Ti H: 8.00 N	O 5830	045.00	e da	BOREHOLE: ETA-05-04 PAGE: 2 OF 2 DRILL TYPE: DRILL: CASING: TUM: BOREHOLE: ETA-05-04 SAMPLE CONDITION Remoulded Undisturbed Lost Rock core	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
		ц	WELL DETAILS	E	STRATIGE	RAPHY		S	SAM	PLES			E	
DEPTH - f		DEPTH - n	LEVEL - m	ELEVATION DEPTH - r	DESCRI	NBOL AMBOL S		TYPE AND NUMBER	CONDITION	RECOVERY	RQD %	SAMPLE DESCRIPTION	DEPTH - I	LABORATORY and IN SITU TESTS
-		1	P P P P P P P P P P P P P P P P P P P	6.10	Till		2 t	5				6.000000 Orangey brown sandy gravel with silt, gravel subangular, wet.		
00-140 Sty PLOTIED, 2003-11-03 13,21113	225	7	P2 P2 P2 P2 P2 P3 P2 P2 P3 P3 P3 P2 P2 P2 P2 P3 P3 P3 P2 P2 P2 P2 P2 P3 P					6				<ul> <li>6.400000</li> <li>Transition from gravel through bony till to silty till. Moist, yellowish brown.</li> <li>8.000000</li> <li>Yellowish brown silty till, sandy gravelly silt, moist.</li> </ul>	- - - - - - - - - - - - - - - - - - -	
-1707-1	-	9		1095.51 9.00	END OF BOREHOLE								9	
ERENCE MATERIALS/geotec.logvemplates/ogv/m/vell-stra	30 35 - - -	10 - 11											10	

	B	SRI OREHO	Engini	<b>Insulting</b> eers and Scientists LOG	PROJECT: LOCATION: ETA Are FILE No: FARO BORING DATE: 200 DIP: AZ COORDINATES: 69	ea (1CE 5-08 <b>(IMU</b> ) 9138	0003.73) -15 T TH: 56.00 N	<b>О</b> 5829	77.00	E DA	BOREHOLE: ETA-05-05 PAGE: 1 OF 2 DRILL TYPE: DRILL: CASING: SAMPLE CONDITION CASING: SAMPLE CONDITION CASING: SAMPLE CONDITION CASING: SAMPLE CONDITION CONDI CONDITION CONDI CONDI CONDITION CO	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
		WELL		STRATIGE	RAPHY			SAME	PLES	5			
DEPTH - ft	DEPTH - m	DETAILS & WATER LEVEL - m	ELEVATION - m DEPTH - m	DESCR	IPTION	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
-		292292 292292 29. 29 29. 29 297 2.95 27 2.97	<u>1104.66</u> 0.00	Tailings		8, 8, 8 8 8	1				0 Tan sandy gravel, dry to slightly moist, subangular - subrounded	-	
		Do     Do     Do     Do       Do     Do     Do     Do <td><u>1099.48</u> 5.18</td> <td>Peat, original ground</td> <td></td> <td></td> <td>2 3 4 5 6 7 7 8 8</td> <td></td> <td></td> <td></td> <td>gravel         0.5         Yellowish, brown tailings, fine sand with silt, moist. Isolated         0.700000         Granite cobble, stuck in bit. Lost 0.7-1.5m. Driller indicated that it felt like moist tailings as from 0.5-0.7m.         1.500000         Moist silty fine sand tailings, yellowish brown.         1.700000         Tan sandy gravel as from 0-0.5m.         2.000000         Yellowish brown fine to medium sand tailings, slightly moist, little to no fines.         3.800000         Tailings as above - transition to yellowish gray in colour.         5.200000         Dark gray to black organics and silt. Moist.         5.400000</td> <td>2 - 3 - 5 - 5</td> <td></td>	<u>1099.48</u> 5.18	Peat, original ground			2 3 4 5 6 7 7 8 8				gravel         0.5         Yellowish, brown tailings, fine sand with silt, moist. Isolated         0.700000         Granite cobble, stuck in bit. Lost 0.7-1.5m. Driller indicated that it felt like moist tailings as from 0.5-0.7m.         1.500000         Moist silty fine sand tailings, yellowish brown.         1.700000         Tan sandy gravel as from 0-0.5m.         2.000000         Yellowish brown fine to medium sand tailings, slightly moist, little to no fines.         3.800000         Tailings as above - transition to yellowish gray in colour.         5.200000         Dark gray to black organics and silt. Moist.         5.400000	2 - 3 - 5 - 5	
X:106	-	DA DA DA DA DAT DAT DAT DAT				8 . ¢. . r . . r .						-	

	B	SRI OREHO	Engin	<b>DINSULT TING</b> Teers and Scientists <b>LOG</b>	PROJECT: LOCATION: ETA Area FILE No: FARO (1 BORING DATE: 2005- DIP: AZIM COORDINATES: 691	1CD -08- <b>MUT</b> 385	0003.73) -15 T FH: 56.00 N	O 5829	77.00	E DAT	BOREHOLE: ETA-05-05 PAGE: 2 OF 2 DRILL TYPE: DRILL: CASING: SAMPLE CONDITION Remoulded Undisturbed Lost Rock core CASING:	TYP DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
		WELL	1	STRATIG	RAPHY	_	5	SAME	PLES				
DEPTH - ft	DEPTH - m	DETAILS & WATER LEVEL - m	ELEVATION - m DEPTH - m	DESCR		SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
	- 7	\$\$\begin{aligned}{c} \$\$\begin{aligned}{c} \$			6 6 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7	· · · · · · · · · · · · · · · · · · ·	10				6.200000 Chl green silty sandy gravel, moist. Particles subrounded to subangular. Becomes sandier towards end of borehole.		
- 25	-      	<u>Po</u> <u>Po</u> <u>Po</u> <u>Po</u> T <u>Po</u> T <u>Po</u> T <u>Po</u> T	7.50	END OF BOREHOLE	Ξ							8	
	- - - -										т И	9 -	
CIAL Digeorec. log wernplates w	- 10							ANY -				10	
X-106 REFERENCE WATE	<b>11</b> <b>1</b>											11	

	B	SRK OREHO	Engine	nsulting eers and Scientists LOG	PROJECT: Faro M LOCATION: Emerg FILE No: FARO BORING DATE: 20 DIP: 90.00 A COORDINATES: 6	line Se ency <sup>-</sup> (1CI 04-08 <b>ZIMU</b> 39138:	eepage II Tailings A 0003.053 30 T TH: 24.47 N	nvestig Area ) <b>O</b> 2 5829	gation 2004-( 177.28	09-01 E <b>DA</b> 1	BOREHOLE: SRK04-03 PAGE: 1 OF 2 DRILL TYPE: Odex 6" DRILL: Air Rotary CASING: 6" UM: BOREHOLE: SRK04-03 SAMPLE CONDITION CASING: 0 SAMPLE CONDITION SAMPLE CONDITION	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
DEPTH - ft	DEPTH - m	WELL DETAILS & WATER LEVEL - m	DEPTH - m	STRATIGF DESCRI Natural ground sur	PTION	SYMBOL	TYPE AND NUMBER		RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
	- 1	1 2 1 2 1 2 1 2 1 2 1 2 1 2 1 2 1 2 1 2	0.00	Sand and gravel fill. Stickup Heights: 0.59 (deep). Tailings Wells are 2" Sched. 4	m (shallow), 0.51m 0 PVC							1 2 3	
2000 - 21	5		1097.54 6.50 1096.14 7.90	0.020 Slot 2" PVC sc Alluvium with tailings Sand and gravel alluv	reen							5- 6- 7- 8- 9-	

	B	SRI OREHO	Engine	eers and Scientists	PROJECT: Faro M LOCATION: Emerg FILE No: FARO BORING DATE: 20 DIP: 90.00 A COORDINATES: 6	line S ency (1CI 04-08 <b>ZIMU</b> 59138	eepage II Failings A 2003.053 -30 T TH: 24.47 N	nvestig vrea ) <b>*O</b> 2 5829	gation 2004-1 77.28	09-01 E <b>DA</b> T	BOREHOLE: SRK04-03 PAGE: 2 OF 2 DRILL TYPE: Odex 6" DRILL: Air Rotary CASING: 6" SAMPLE CONDITION SAMPLE CONDITION Conditional Statements SAMPLE CONDITION SAMPLE CONDITION Conditional Statements CASING: 6" CUM:	N TY DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
		WELL		STRATIGE	RAPHY	1		SAMF	PLES	;			
DEPTH - ft	DEPTH - m	DETAILS & WATER LEVEL - m	ELEVATION - m DEPTH - m	DESCRI	IPTION	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %	SAMPLE DESCRIPTION	DEPTH - m	LABORATORY and IN SITU TESTS
-	-												
- 35	- 11		1093.34 10.70	Weathered bedrock, Actual start of weather ambiguous.	sand and gravel. ered zone							11-	
- 40	- 12		1092.14 11.90	2" Sched. 40 PVC sc Bedrock	reen							12	
	- 13	4. 4. 8. 4. 4. 4. 5. 6. 7. 5. 8. 6. 6. 6. 6. 7. 6. 8. 8. 7. 6. 7. 7. 7. 8. 8.	1090.64		-							13-	
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- 5	- 1 - 2	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	$\begin{array}{c} \mathcal{P} \rho & \mathcal{P} \rho \\ \mathcal{P} \sigma & \mathcal{P} \rho \\ \mathcal{P} \rho & \mathcal{P} \rho \\ \mathcal{P} $					1					2	
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	-	P D.0 D.0. D.0 D.0. D.0	DO DO DRD DO RO DO	5.70	Tailings; alluvium			1				5.700000		
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DEPTH - ft	DEPTH - m	WELL DETAILS & WATER LEVEL - m	ELEVATION - m DEPTH - m	DESCR	IPTION	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %		SAMPLE DESCRIPTIO	Ν	DEPTH - m	LABORATORY and IN SITU TESTS
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	B	SRI OREHO	Engine	eers and Scientists	PROJECT: LOCATION: ETA Are FILE No: FARO BORING DATE: 200 DIP: AZ COORDINATES: 69	ea (1CE 95-09 <b>2IMU</b> 91384	0003.73) -25 T TH: 46.00 N	<b>O</b> 25829	2005-0	99-28 E <b>DA</b> T	BOREHOLE: SRK05-ETA-BR2 PAGE: 3 OF 4 DRILL TYPE: DRILL: CASING: TUM: BOREHOLE: SRK05-ETA-BR2 SAMPLE CONDITION Remoulded Undisturbed Lost Rock core	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
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											14.600000 Sheared/crushed zone 15.24-15.39m. Weathered zone at 16.15m.	15	
MMell							7		100	10		-	
AL Sigeotec. logiemplates vogi-	- 16										16.150000 Schist with quartz lenses, up to 5cm thick. +/- vertical joint ~1.46m	16	
NCE MATERI	- 17						8		97	10		17-	
EFEREI												-	k = 1.10E-07
X:106 R	-										17.799999	-	k = 4.33E-08

	B	/= SRI oreh(	Engini	eers and Scientists	PROJECT: LOCATION: ETA Are FILE No: FARO BORING DATE: 200 DIP: AZ COORDINATES: 65	ea (1CE 05-09 <b>2IMU</b> 91384	0003.73) -25 T TH: 16.00 N	<b>O</b> 2 5829	2005-0	09-28 E <b>DA</b> T	B P D C C	OREHOLE: SRK05-ETA-BR2 AGE: 4 OF 4 PRILL TYPE: PRILL: CASING:	SAMPLE CONDITION Remoulded Undisturbed Lost Rock core	TYF DC GS SS	PE OF SAMPLER Diamond core barrel Grab sample Split spoon
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DEPTH - ft	DEPTH - m	& WATER LEVEL - m	ELEVATION - n DEPTH - m	DESCRI	PTION	SYMBOL	TYPE AND NUMBER	CONDITION	RECOVERY %	RQD %		SAMPLE DESCRIPTIO	DN	DEPTH - m	LABORATORY and IN SITU TESTS
- 60							9		100	22					
	19										19.330000			19	
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	5 - 23		1079.55				12		100	23				23-	
X:106 KE			23.60	END OF BOREHOLI										-	

Appendix B Laberge Environmental Services Memorandums



Memorandum

To:Dan MackieSRKChristoph WellsRGCSeptember 4, 2005Copies:Deloitte, GLL, RGC, BGC, Faro Project Office, Water ResourcesFrom:Ken NordinLES

## Re: Installation of Weirs in the Emergency Tailings Area (ETA)

This is a brief description of two weir installations in the ETA; at X\_23 and in the Faro Creek canyon at FCS\_4.

## X\_23 Old Waste Rock Dump Seep near remnant Faro Creek channel

A one foot rectangular weir was replaced with a 90° V-notch weir on August 30, 2005 using a 235 CAT excavator. The bulkhead was made by the welding shop at the Faro mine site.



Old rectangular weir at X\_23



replacement 90° weir September 1, 2005





Two sumps were dug upstream of the head pond and a small berm was placed upstream to help trap sediment. The weir was tested on September  $2^{nd}$ , using two different containers for volumetric measurements, and using the standard weir formula Q= 1362.9 (H<sup>2.5</sup>). 26 L container = 0.6 L/sec 9.6 L container = 0.59 L/sec Formula (H=0.045m) = 0.59 L/sec

A staff gauge was installed as shown, and adjusted to read 0.045m (H).

## FCS\_4 Remnant Faro Creek Channel at end of Canyon

A standard 90<sup>°</sup> V-notch weir was installed in the Faro Creek channel at the end of the canyon on August 30, 2005 using a 235 CAT excavator. The bulkhead was made by the welding shop at the Faro mine site.



Weir at bottom of Faro Creek canyon





The weir was tested on September  $2^{nd}$ , using a container for volumetric measurements, and using the standard weir formula Q L/sec = 1362.9 (H<sup>2.5</sup>). 9.6 L container = 7.62 L/sec Formula (H=0.125m) = 7.56 L/sec

A staff gauge was installed as shown, and adjusted to read 0.125 m (H).

The head on a standard 90<sup>°</sup> weir (H) is to be measured as a depth above the elevation of the crest or vertex of the notch. H should be measured at a distance upstream of four times the maximum H expected. In practice, H can be measured closer to the bulkhead. One way to measure H at the notch is to use a set square or thin ruler and measure H just upstream of the plate.

Provided conditions for a standard weir are met, a rating table or curve can be used with the formula above. Standard weirs must have a sharp edge, 90 degree angle, close to zero approach velocity, free falling nappe, and be level and plumb. A rating table and chart follow which can be used for any standard 90 degree V-notch weir. The chart covers low level discharges up to 20 L/sec.

H (m)	L/sec												
0.030	0.21	0.073	1.97	0.108	5.25	0.143	10.59	0.178	18.30	0.223	32.15	0.258	46.29
0.035	0.31	0.074	2.04	0.109	5.37	0.144	10.77	0.179	18.56	0.224	32.52	0.259	46.74
0.040	0.44	0.075	2.11	0.110	5.49	0.145	10.96	0.180	18.82	0.225	32.88	0.260	47.20
0.041	0.47	0.076	2.18	0.111	5.62	0.146	11.15	0.181	19.08	0.226	33.25	0.270	51.87
0.042	0.49	0.077	2.25	0.112	5.75	0.147	11.34	0.182	19.35	0.227	33.61	0.280	56.80
0.043	0.52	0.078	2.33	0.113	5.88	0.148	11.54	0.183	19.62	0.228	33.99	0.290	62.01
0.044	0.56	0.079	2.40	0.114	6.01	0.149	11.73	0.184	19.88	0.229	34.36	0.300	67.49
0.045	0.59	0.080	2.48	0.115	6.14	0.150	11.93	0.185	20.16	0.230	34.74	0.310	73.26
0.046	0.62	0.081	2.56	0.116	6.27	0.151	12.13	0.186	20.43	0.231	35.12	0.320	79.31
0.047	0.66	0.082	2.64	0.117	6.41	0.152	12.33	0.187	20.70	0.232	35.50	0.330	85.65
0.048	0.69	0.083	2.72	0.118	6.55	0.153	12.54	0.188	20.98	0.233	35.88	0.340	92.29
0.049	0.73	0.084	2.80	0.119	6.69	0.154	12.74	0.189	21.26	0.234	36.27	0.350	99.23
0.050	0.77	0.085	2.88	0.120	6.83	0.155	12.95	0.190	21.55	0.235	36.66	0.360	106.47
0.051	0.80	0.086	2.97	0.121	6.97	0.156	13.16	0.191	21.83	0.236	37.05	0.370	114.02
0.052	0.84	0.087	3.06	0.122	7.12	0.157	13.37	0.192	22.12	0.237	37.44	0.380	121.88
0.053	0.89	0.088	3.15	0.123	7.26	0.158	13.59	0.193	22.41	0.238	37.84	0.390	130.06
0.054	0.93	0.089	3.24	0.124	7.41	0.159	13.80	0.194	22.70	0.239	38.24	0.400	138.55
0.055	0.97	0.090	3.33	0.125	7.56	0.160	14.02	0.195	22.99	0.240	38.64	0.410	147.38
0.056	1.02	0.091	3.42	0.126	7.72	0.161	14.24	0.196	23.29	0.241	39.04	0.420	156.53
0.057	1.06	0.092	3.52	0.127	7.87	0.162	14.46	0.197	23.58	0.242	39.45	0.430	166.01
0.058	1.11	0.093	3.61	0.128	8.03	0.163	14.69	0.198	23.89	0.243	39.85	0.440	175.83
0.059	1.16	0.094	3.71	0.129	8.18	0.164	14.91	0.199	24.19	0.244	40.27	0.450	185.99
0.060	1.21	0.095	3.81	0.130	8.34	0.165	15.14	0.200	24.49	0.245	40.68	0.500	242.04
0.061	1.26	0.096	3.91	0.131	8.50	0.166	15.37	0.201	24.80	0.246	41.10		
0.062	1.31	0.097	4.01	0.132	8.67	0.167	15.60	0.202	25.11	0.247	41.52		
0.063	1.36	0.098	4.12	0.133	8.83	0.168	15.84	0.203	25.42	0.248	41.94		
0.064	1.42	0.099	4.22	0.134	9.00	0.169	16.08	0.204	25.57	0.249	42.36		
0.065	1.47	0.100	4.33	0.135	9.17	0.170	16.32	0.205	26.05	0.250	42.79		
0.066	1.53	0.101	4.44	0.136	9.34	0.171	16.56	0.206	26.37	0.251	43.22		
0.067	1.59	0.102	4.55	0.137	9.51	0.172	16.80	0.207	26.69	0.252	43.65		
0.068	1.65	0.103	4.66	0.138	9.69	0.173	17.04	0.208	27.02	0.253	44.08		
0.069	1.71	0.104	4.78	0.139	9.86	0.174	17.29	0.209	27.34	0.254	44.52		
0.070	1.78	0.105	4.89	0.140	10.04	0.175	17.54	0.210	27.67	0.255	44.96		
0.071	1.84	0.106	5.01	0.141	10.22	0.176	17.79	0.211	28.00	0.256	45.40		
0.072	1.90	0.107	5.13	0.142	10.40	0.177	18.05	0.222	31.79	0.257	45.85		

Rating Table for Standard 90 degree V-notch Weir where Q (L/sec) =1362.9H^2.5

ETA Weirs




Memorandum

To:Christoph Wels RGC, Dan Mackie SRKOctober 23, 2005Copies:Deloitte, GLL, Faro Project Office, Water ResourcesFrom:From:Ken Nordin LES

#### Re: Installation of Weir at FCS 5 and Initial Monitoring of Faro Creek Seepage Losses

This is a brief description of the results of the project *Additional Monitoring of Faro Creek Seepage Losses* for October 2005.

#### X\_23 FCS\_1 Old Waste Rock Dump Seep near remnant Faro Creek channel

A one foot rectangular weir was replaced with a  $90^{\circ}$  V-notch weir on August 30, 2005 using a 235 CAT excavator. The bulkhead was made by the welding shop at the Faro mine site. Two sumps were dug upstream of the head pond and a small berm was placed upstream to help trap sediment. The weir was tested on September  $2^{nd}$ , using two different containers for volumetric measurements, and using the standard weir formula Q= 1362.9 (H<sup>2.5</sup>). A staff gauge was installed and adjusted to read (H). On September 15, 2005 a PT2X sensor/logger was installed. At 11:30 AM the measured H over the notch was 0.059 m. On October 18, 2005 the datalogger was downloaded. The data was sent to RGC and SRK in Excel format on October 21.

#### FCS\_4 Remnant Faro Creek Channel at end of Canyon

A standard 90<sup>°</sup> V-notch weir was installed in the Faro Creek channel at the end of the canyon on August 30, 2005 using a 235 CAT excavator. The bulkhead was made by the welding shop at the Faro mine site. A staff gauge was installed and adjusted to read (H). On September 15, 2005 a PT2X sensor/logger was installed. At 13:37 the measured H over the notch was 0.130 m. On October 18, 2005 the datalogger was downloaded. The data was sent to RGC and SRK in Excel format on October 21.

#### FCS\_5 Old Tailings dam decant X\_1 at end of ditch

A standard 90<sup>°</sup> V-notch weir was installed near the old decant formerly known as X\_1 on October 18, 2005. A chart pac data logger and PS9800 pressure transducer were installed temporarily while waiting for the INW PT2X sensor/logger. Then installation consists of a 1/4 "metal weir bulkhead with 30 mil polypropylene liner and earth embankments. The 235 CAT excavator was used to build the installation. The bulkhead was made by the welding shop at the Faro mine site. The installation was provided with a pressure sensor and datalogger which had been deployed at X\_14 for the summer. A dedicated INW PT2X sensor/logger is on order and will be installed in November.

#### Rating Table

Provided conditions for a standard weir are met, a rating table or curve can be used with the formula above. Standard weirs must have a sharp edge, 90 degree angle, close to zero approach velocity, free falling nappe, and be level and plumb. On September 3, a rating table was supplied with a memo on the ETA weirs. There was a transcription error in the table (the coefficient 136**9.2** was used instead of 136**2.9**). A corrected rating table and chart are included.

### Results of measurements at recording stations

The following is a summary of discharge measurements taken to date. An Excel file of containing the downloaded data from FCS\_1 and FCS\_4 was forwarded to RGC and SRK on October 21, 2005.

### FCS\_1 or X\_23

Date/Time	H (m)	Q (L/sec	Sensor (m)	Offset (m)	comment
Aug 30 11:00	0.045	0.585			fmla
Aug 30 11:00		0.6			Site bucket 26 L
Aug 30 11:00		0.59			LES bucket 9.6 L
Sept 15 11:32	0.059	1.15			fmla
Sept 15 11:32		1.14	.2478	.1888	21.5 L bucket
Oct 3 07:30?	0.061	1.25	.2450?		fmla
Oct 18 13:35	0.059	1.15	.2443	.1853	fmla
Oct 18 13:35		1.15			20.5 L in dedicated
					bucket

## FCS\_4 Faro Creek Canyon

Date/Time	H (m)	Q (L/sec	Sensor (m)	Offset (m)	comment
Sept 1 16:00	0.125	7.56			fmla
Sept 1 16:00		7.62			9.6 L bucket
Sept 15 15:30	0.130	8.30	.2394	.1091	fmla
Sept 15 15:30		8.6			9.6 L bucket
Oct 3 07:30?	0.115	6.11	.2219	.1069	fmla
Oct 18 15:23	0.123	7.23	.2291	.1061	fmla
Oct 18 15:30		7.17			21.0 L dedicated bucket

## Results of initial survey ETA/Faro Creek seepage

The following table shows the results of discharge measurements taken during the October 18 initial survey of stations FCS\_1 through to FCS\_7 in L/sec. Scanned pages of field notes, site photos and the ALS chain of custody forms are also included. Measurements at each of the weir sites were made using the formula for a standard  $90^{\circ}$  V-notch weir, and compared with volumetric measurements at the same time. Volumetric measurements were made using a calibrated bucket. In most cases five trials were made with the lowest and highest result discarded and the remaining three measurements averaged to determine the rate of flow. A portable  $90^{\circ}$  V-notch weir box was used at FCS\_6. It should be noted that this device reached a maximum head of 0.094 m before the containment dam broke, and that there was some seepage through this dam (difficult to estimate but likely less than 5% of the flow).

Station ID	Location	October 2005 DD/Time	Trial 1	Trial 2	Cond. <i>u</i> S/cm	pН	TDS mg/L
Reach 3							
FCS1	WRD seepage in old Faro Creek channel (at X23)	18/13:3 5	1.15 <sup>v</sup>	1.15 <sup>w</sup>	4230	6.21	2060
FCS2	surface seepage discharging below road (below road at culvert)	18/14:3 0	3.39 <sup>v</sup>		4890	7.47	2200
FCS3	subsurface seepage discharging at seepage face below road (at X7) (w/ organic smell)	18/15:0 0	3.12 <sup>v</sup>		5010	6.07	2390
FCS4	combined seepage below confluence of X7 and X23 (at mouth of Faro Creek canyon)	18/15:2 3	7.23 <sup>w</sup>	7.17 <sup>v</sup>	4610	6.82	2120
FCS5	Seepage flow at end of diversion ditch (prior to discharge into Interm. Impoundment) new weir installed Oct.18	18/16:0 0	6.38 <sup>w</sup>	6.28 <sup>v</sup>	4350	6.02	1940
FCS5		19/16:4 5	6.11 <sup>w</sup>				
FCS6	Seepage flow appr. Halfway towards Interm. Pond. Station established and flagged Oct 18. Q by portable weir box	18/18:0 0	3.71 <sup>wb</sup>		4840	5.02	2200
FCS7	seepage flow near pond (but u/s of inflow from Guardhouse Creek) Station established and flagged Oct 18	18/17:3 0	2.7 <sup>efa</sup>		5220	3.51	2360
GHC	Guardhouse Creek before discharge into Intermediate Impoundment (at road)	18/1730	5 <sup>e</sup>				

<sup>e</sup> estimated, judgmental

<sup>efa</sup> estimated velocity float x area
 <sup>v</sup> volumetric (average trials time to fill calibrated

bucket)

<sup>w</sup> weir

<sup>wb</sup> portable weir box

H (m)	L/sec												
0.030	0.21	0.073	1.96	0.108	5.22	0.143	10.54	0.178	18.22	0.223	32.01	0.258	46.08
0.035	0.31	0.074	2.03	0.109	5.35	0.144	10.72	0.179	18.48	0.224	32.37	0.259	46.53
0.040	0.44	0.075	2.10	0.110	5.47	0.145	10.91	0.180	18.73	0.225	32.73	0.260	46.98
0.041	0.46	0.076	2.17	0.111	5.59	0.146	11.10	0.181	19.00	0.226	33.09	0.270	51.63
0.042	0.49	0.077	2.24	0.112	5.72	0.147	11.29	0.182	19.26	0.227	33.46	0.280	56.54
0.043	0.52	0.078	2.32	0.113	5.85	0.148	11.48	0.183	19.53	0.228	33.83	0.290	61.72
0.044	0.55	0.079	2.39	0.114	5.98	0.149	11.68	0.184	19.79	0.229	34.20	0.300	67.18
0.045	0.59	0.080	2.47	0.115	6.11	0.150	11.88	0.185	20.06	0.230	34.58	0.310	72.92
0.046	0.62	0.081	2.54	0.116	6.25	0.151	12.08	0.186	20.34	0.231	34.95	0.320	78.95
0.047	0.65	0.082	2.62	0.117	6.38	0.152	12.28	0.187	20.61	0.232	35.33	0.330	85.26
0.048	0.69	0.083	2.70	0.118	6.52	0.153	12.48	0.188	20.89	0.233	35.72	0.340	91.87
0.049	0.72	0.084	2.79	0.119	6.66	0.154	12.68	0.189	21.17	0.234	36.10	0.350	98.77
0.050	0.76	0.085	2.87	0.120	6.80	0.155	12.89	0.190	21.45	0.235	36.49	0.360	105.98
0.051	0.80	0.086	2.96	0.121	6.94	0.156	13.10	0.191	21.73	0.236	36.88	0.370	113.49
0.052	0.84	0.087	3.04	0.122	7.09	0.157	13.31	0.192	22.01	0.237	37.27	0.380	121.32
0.053	0.88	0.088	3.13	0.123	7.23	0.158	13.52	0.193	22.30	0.238	37.66	0.390	129.46
0.054	0.92	0.089	3.22	0.124	7.38	0.159	13.74	0.194	22.59	0.239	38.06	0.400	137.92
0.055	0.97	0.090	3.31	0.125	7.53	0.160	13.96	0.195	22.88	0.240	38.46	0.410	146.70
0.056	1.01	0.091	3.40	0.126	7.68	0.161	14.18	0.196	23.18	0.241	38.86	0.420	155.81
0.057	1.06	0.092	3.50	0.127	7.83	0.162	14.40	0.197	23.48	0.242	39.26	0.430	165.25
0.058	1.10	0.093	3.59	0.128	7.99	0.163	14.62	0.198	23.78	0.243	39.67	0.440	175.02
0.059	1.15	0.094	3.69	0.129	8.15	0.164	14.84	0.199	24.08	0.244	40.08	0.450	185.14
0.060	1.20	0.095	3.79	0.130	8.30	0.165	15.07	0.200	24.38	0.245	40.49	0.500	240.93
0.061	1.25	0.096	3.89	0.131	8.47	0.166	15.30	0.201	24.69	0.246	40.91		
0.062	1.30	0.097	3.99	0.132	8.63	0.167	15.53	0.202	24.99	0.247	41.32		
0.063	1.36	0.098	4.10	0.133	8.79	0.168	15.77	0.203	25.30	0.248	41.74		
0.064	1.41	0.099	4.20	0.134	8.96	0.169	16.00	0.204	25.62	0.249	42.17		
0.065	1.47	0.100	4.31	0.135	9.13	0.170	16.24	0.205	25.93	0.250	42.59		
0.066	1.53	0.101	4.42	0.136	9.30	0.171	16.48	0.206	26.25	0.251	43.02		
0.067	1.58	0.102	4.53	0.137	9.47	0.172	16.72	0.207	26.57	0.252	43.45		
0.068	1.64	0.103	4.64	0.138	9.64	0.173	16.97	0.208	26.89	0.253	43.88		
0.069	1.70	0.104	4.75	0.139	9.82	0.174	17.21	0.209	27.22	0.254	44.31		
0.070	1.77	0.105	4.87	0.140	10.00	0.175	17.46	0.210	27.54	0.255	44.75		
0.071	1.83	0.106	4.99	0.141	10.17	0.176	17.71	0.211	27.87	0.256	45.19		
0.072	1.90	0.107	5.10	0.142	10.36	0.177	17.96	0.222	31.65	0.257	45.63		

## (Corrected) Rating Table for Standard 90 degree V-notch Weir where Q (L/sec) =1362.9H^2.5

Additional Monitoring Faro Creek Seepage Losses October 23, 2005



bucket 12 V= 20.0 -. 5= 20.5 FCS-1 (X-23) Oct. 18, 2005 T. 17.88 (bucket on slant) Oct. 18,2005 Tz 17.75 Ta - 2°C wind SIS PICI T3 17.80 AVE= 17.81 Q= 1.15 samples 4 se 12, 100 ml, 250 mc PH Hon weir = 0.059m 6.21 Fm/a Q = 1.3692(059 2.5) = 1.16 LISEC COND 4,230 uslom · two pictures - stu measuring H TDS 2060 mg/L TOC 4.20 3.2 FCS-2 suicide auve, Download. masure + sample D'S Battery 82% end alvert > chop out ice Free 83,157 04 7.47 Time 13:35:39 LOND 4890 us/cm Real 13:35:19 TDS 2200 mg/1 amont 24.43 cm TUC - 0.1°C start new session samples; 12, 250 ml, 100 ml X-23 FCS-1 Winter saved pile as X-23 FCS-100F18 · chop outice, flows freely, able to put bucket under Q. " measure bucket = 10,52 2005 in AQ4 and Excel (used Imhoff come) T. S. 31 T3 2.99 T5 3.09 Tz 3. 13 T4 3.09 - X highthown, AVE= 3.10 Q= 3.39 L/SEC

6 Oct. 18 2005 FCS\_ 3 u/s canyon FCS.4 Dowload: pH 6.07 Time 15:20:12 real 15:20:06 10ND 5010 5020 MS/CM 22.93 cm 82% battery TDS 2390 mg/L 82, 828 Free records TE 3.3 (WeirH = 0, 123 m) · Samples; 12, 250 ml, 100 ml · Q Fmla= 7.26 L/Sec · Q by blue bucket U= 10.5 L · start new session 30 mm intervals for 80,00 records, T. 3.21 T4 3.34 T2 3.50 T5 3-28 did not enace old session. 73 3.59 × hi+b, AVE= 3.37 sec Q = 10.5/3.37= 3.12 LISEC - Bricket Q V= 21.01 · I picture 1/cg @ sTu / FCS\_ 3 T, 2.59 T4 3.18 FCS-4 Faro heet Conya Tz 2.56 T5 3.03 T3 3.10 × h1+10, AVE= 2.93 pH . 6.82 COND 4610 USICM · Q= 21,0/2.93 = 7.17 L/sec TPS 2020 mg/6 2120 Frula 1362.9 H2.5 7.23 LISEC TOL INC

8 Oct. 18 2005 FCS-5 (OLD X+1) · left FCS-6 to continue 4:00 pM -> Illing head pond, want pH 6.02 Fmla = 6.4 Usec to Garand house Creek COND 4350 (at 4:10 pm) where it flows into tailing, 1940 TDS directly accross from atte, To -0.1°C n-establish FCS.7. Est. QN 5450 · samples; 12, 250 mL, 100 mL · Weir H = 0. 117 m 410 pm FCS.7: chop a channel 150 cm wide (H= 0.121 m 6:22 pm (H= 0.115 m 14:45 oct.19) ave scm deep. Q phie bucket V= 10.5L Destimate: flowing section is 70 cm wide T, 1.78 T4 1.68 Tz 1.59 T5 1.69 (the rest is no V) T3 1.65 Xhitle AVE= 1.67 estimate V by floats to Q=10.5/1.67=6.28 travel 25 cm: FCS-6 T1 2.61 mark up FCG Tz 2.50 · instal wein box, T3 2.71 sec AVE= 2.61 sec Hover 90° wen = 0.06/m Q= (.096)(.028) pH 3.51 at 5:00 PM, 0.066 5:10 pm = 2.7 L/sec COND 5220 us/cm 0.081 5:10 pm 0.094 Gioppm TDS 23.60 mg/L (max.) pH 5.02 00 TOC Q= 3.71 LISEC (OND 4840 TDS 2200 myll · Samples; 12, 250 ml, 100 ml - 0.1% TP.

11 10 Oct. 18,2005 X-14 Oct. 19, 2005 FCS-7 CONTD · download X-14 main . mont wp FC7 0881533 logger 2:40 km. unant reading 0.1428 m . Saved file as 081 6913907 · sample ; 12, 250 ml, issue X-14-0ct, 19,2005 · sever 2 pirs liky across ita Fano Projets. truds GHC, Ikg us taks - Turnto X-14-2 backup old tailingo dam. (5:50 pm) to save ata due to balance of Q's burnt out aux. buttery not mough juice, recover susor + logger for uployment at FES-5 F.CS-1- 1.16 FCS.2 3.39 FCS.3 3.12 took quite a while to remove
 X 14-2 due to post, but FCS.4 7.26 FCS.5 6.4 (4:10 pm) FCS.6 3.71 came out OK F(S.7 2.70 - note Field piter blank with ALS Millio "OS/09/08" - Blind Dup Orss metals S FCS-6 15

16 R-7 5:00-> · Return to FCS-5, install chartpac + PS9800 sensor · download · unently reading . 3369 m behind weir current lepth on sensor = . 4073 m batteries good section à poran over unent It own Unotch = . 115 m may be artificially high (2:45 PM) . Stu installs Plag poles at stage due to re in pond DIS of section .. FCS-6 and FCS-7 "No way to do good Q - sunny, - 1°c calm Drop Strat RCDC\_4, go to FCS-4: · picked some Alpine rescue seeds · clean out blue green · no morement on sensor studge UIS wein, when monitoring H= 0.120 with set square 'saved file R-7 October, 19, 2005 nt 3:20 pm. - note time was off by ·NFRC-22/23 X-2 a jeu minutes so re-set 3:45-7 chop out shore jast + jazz. 1 ree to clean X-section, data logger dock to CPUL, ne-start logsing. mater with AA NFRC\_23 0.712 note stage will be artificially high due to icing



FCS\_1 Oct 18.05



FCS\_2 Oct.18.05



FCS\_3 Oct.18.05



FCS\_4 Oct.18.05. Note high water mark on bank.



FCS\_6 portable weir box



FCS\_6 looking across portable weir



FCS\_7 looking across towards Guardhouse Creek



FCS\_7 looking upstream towards old tailings dam Oct.18.05

ALS Environmental excellence in analytical testing

1988 Triumph Street, Vancouver, BC Canada V5L 1K5 Tel: 604-253-4188 Toll Free: 1-800-665-0243 Fax: 604-253-6700 #2 - 21 Highfield Circle SE, Calgary, AB Canada T2G 5N6 Tel: 403-214-5431 Toll Free: 1-866-722-6231 Fax: 403-214-5430 #2 - 8820 100th Street, Fort SL John, BC Canada V1J 3W9 Tel: 250-785-8281 Fax: 250-785-8286

#### CHAIN OF CUSTODY FORM

PAGE \_\_\_\_\_\_ OF \_\_\_\_\_

35175

SEND REPORT TO:

CLIENT: ROBERTSON GEDCONSULTANTS													-			
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	FCS-5	2005-10-18	HZO	V	V	~	~									
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Memorandum

To:Christoph Wels RGC, Dan Mackie SRKCopies:Deloitte, GLL, Faro Project OfficeFrom:Ken Nordin LES

January 5, 2006

#### Re: Additional Monitoring of Faro Creek Seepage Losses

This is a brief description of the results of the project *Additional Monitoring of Faro Creek Seepage Losses* for December 2005. All sites were accessed by skidoo – FCS\_2 and FCS\_7 were frozen solid and FCS\_5 was heavily affected by icing.

#### FCS\_1 WRD seepage in old Faro Creek channel (at X23)

FCS\_1 was visited at 12:30 PM. There was a thin ice cover over the notch and water was flowing freely underneath. H over the notch was 0.045m (0.59 L/sec). Bucket discharge was approximately 0.67 L/sec. We tried to download the PT2X sensor but the computer and battery were frozen (air temp.  $-20^{\circ}$  C).



## FCS\_2 Downstream of access road at suicide curve

The entire culvert barrel was filled with ice. There was still some wet areas downstream of the culvert so some seepage was likely still reporting trough the or under the ice although there was not a confined flow or any place to collect a sample.



FCS\_2 downstream of the road December 19, 2005



## FCS\_3 Subsurface seepage discharging at seepage face below road (at X7)

Again, conditions were favourable for sampling and reliable volumetric measurements at this site. Discharge was 2.3 L/sec.





FCS\_4 Combined seepage at the mouth of Faro Creek Canyon

Things looked frozen soild at this site, but proved otherwise. The V-notch was flowing freely uner 0.5 m of bright orange ice. H over the notch was hard to meaure, but was about 10 cm (4.3 L/sec). Discharge by bucket was 4.7 L/sec. The PT2X sensor was not downloaded due to dead computer.





FCS\_4 looking downstream Dec 19, 2005



0.5m ice cover; water flowing over notch freely



FCS\_4 at weir looking upstream

## FCS\_5 Old Tailings dam decant X\_1 at end of ditch

This site began to ice up in early December. By December 19 a thick accumulation had built up directly over the weir. Water was still flowing around the left hand side, some of which was spilling out onto the old tailings. A channel was cut to allow free flow around the left side. We chopped out the V-notch and found it frozen solid. The PT2X sensor could not be downloaded. The communication cable was re-set on a higher pole to make sure it would not be inundated by ice. The flow around the bulkhead was estimated with a bucket at 2.5 L/sec.



FCS\_5 glacier



Some flow onto old tailings







Looking upstream



Channel cut left side

new pole to suspend cable FCS\_5

**FCS\_6** Seepage flow approximately halfway towards Intermediate pond This site was heavily iced up but there was free water on the sides of the frozen channel. We were able to collect samples without difficulty but there was no way to estimate the flow. There was definitely some free flowing water.





FCS\_6 site Dec 19 2005

FCS\_7 seepage flow near pond (but u/s of inflow from Guardhouse Creek) This site was white, indicating that it was still frozen solid as it was in November.

### Results of measurements at recording stations

The following is a summary of discharge measurements taken to date at the recording sites

### FCS\_1 or X\_23

Date/Time	H (m)	Q (L/sec	Sensor (m)	Offset (m)	comment
Aug 30 11:00	0.045	0.585			fmla
Aug 30 11:00		0.6			Site bucket 26 L
Aug 30 11:00		0.59			LES bucket 9.6 L
Sept 15 11:32	0.059	1.15			fmla
Sept 15 11:32		1.14	.2478	.1888	21.5 L bucket
Oct 3 07:30?	0.061	1.25	.2450?		fmla
Oct 18 13:35	0.059	1.15	.2443	.1853	fmla
Oct 18 13:35		1.15			20.5 L in dedicated bucket
Nov 21 10:00		1.05			About 19.5 L in the
					dedicated bucket - can't
					make the bucket level
Nov 21 10:00	0.055	0.97	.2390	.1840	fmla
Dec 19 12:30	0.045	0.59			Fmla, computer frozen, not
					downloaded
Dec 19 12:30		0.67			Site bucket (19 L fill)

## FCS\_4 Faro Creek Canyon

Date/Time	H (m)	Q (L/sec	Sensor (m)	Offset (m)	comment
Sept 1 16:00	0.125	7.56			fmla
Sept 1 16:00		7.62			9.6 L bucket
Sept 15 15:30	0.130	8.30	.2394	.1091	fmla
Sept 15 15:30		8.6			9.6 L bucket
Oct 3 07:30?	0.115	6.11	.2219	.1069	fmla
Oct 18 15:23	0.123	7.23	.2291	.1061	fmla
Oct 18 15:30		7.17			21.0 L dedicated bucket
Nov 21 16:30		5.4			About 20 L in dedicated
					bucket – can't get bucket
					level
Nov 21 16:30	0.109	5.35	0.2155	.1065	fmla
Dec 19 13:30	0.1	4.3			fmla
Dec 19 13:30		4.7			Bucket, 19L. Not
					downloaded

	$1 \Lambda_1$ at the			Carryon	
Date/Time	H (m)	Q (L/sec	Sensor (m)	Offset (m)	comment
Oct 18 14:30	0.115	6.11	0.4028	.2878	fmla
Nov 21 16:30		4.5			11.75 L bucket, fills fast
Nov 21 16:30	0.094	3.7	0.359	.2650	fmla
Dec 19 16:00	0	2.5			Flow reporting around left side of bulkhead. V-notch frozen solid

FCS\_5 near old X\_1 at end of ditch from Faro Creek Canyon

Copies of field notes and chain of custody forms follow. If you have any questions contact the undersigned anytime.

#### Ken Nordin Laberge Environmental Services

Station ID	Location	December 2005 DD/Time	Trial 1	Trial 2	*Cond. <i>u</i> S/cm	рН	*TDS mg/L
Reach 3							
FCS1	WRD seepage in old Faro Creek channel (at X23)	19/12:30	0.67 <sup>v</sup>	0.59 <sup>w</sup>	4160	6.05	1910
FCS2	Culvert chock full, backing up upstream of road, still some wet spots below but no way to get samples or flow						
FCS3	subsurface seepage discharging at seepage face below road (at X7) (w/ organic smell)	19/14:30	2.3 <sup>v</sup>		5120	5.88	2540
FCS4	combined seepage below confluence of X7 and X23 (at mouth of Faro Creek canyon)	19/13:30	4.3 <sup>w</sup>	4.7 <sup>v</sup>	4100	5.85	1880
FCS5	Seepage flow at end of diversion ditch (prior to discharge into Interm. Impoundment) new weir installed Oct.18. Iced up Dec 19	19/16:00		2.5 <sup>∨</sup>	3590	5.84	1640
FCS6	Seepage flow appr. Halfway towards Interm. Pond. Station established and flagged Oct 18. Q by portable weir box. Iced up Dec 19, negligible flow	19/16:30	<0.1 <sup>e</sup>		4170	3.88	2080
FCS7	seepage flow near pond (but u/s of inflow from Guardhouse Creek) Station established and flagged Oct 18. still frozen Dec 19	19/16:4	nil		na	na	na
GHC	Guardhouse Creek before discharge into Intermediate Impoundment (at road)						

\*Orion conductivity meter 115A+

<sup>e</sup> estimated, judgmental <sup>v</sup> volumetric (average trials time to fill calibrated

bucket)

<sup>w</sup> weir

Note conductivity readings suspect - when checked at guesthouse later they were off-scale

(Corrected) Rating Table for Standard 90 degree V-notch Weir where Q (L/sec) =1362.9H^2.5

H (m)	L/sec												
0.030	0.21	0.073	1.96	0.108	5.22	0.143	10.54	0.178	18.22	0.223	32.01	0.258	46.08
0.035	0.31	0.074	2.03	0.109	5.35	0.144	10.72	0.179	18.48	0.224	32.37	0.259	46.53
0.040	0.44	0.075	2.10	0.110	5.47	0.145	10.91	0.180	18.73	0.225	32.73	0.260	46.98
0.041	0.46	0.076	2.17	0.111	5.59	0.146	11.10	0.181	19.00	0.226	33.09	0.270	51.63
0.042	0.49	0.077	2.24	0.112	5.72	0.147	11.29	0.182	19.26	0.227	33.46	0.280	56.54
0.043	0.52	0.078	2.32	0.113	5.85	0.148	11.48	0.183	19.53	0.228	33.83	0.290	61.72
0.044	0.55	0.079	2.39	0.114	5.98	0.149	11.68	0.184	19.79	0.229	34.20	0.300	67.18
0.045	0.59	0.080	2.47	0.115	6.11	0.150	11.88	0.185	20.06	0.230	34.58	0.310	72.92
0.046	0.62	0.081	2.54	0.116	6.25	0.151	12.08	0.186	20.34	0.231	34.95	0.320	78.95
0.047	0.65	0.082	2.62	0.117	6.38	0.152	12.28	0.187	20.61	0.232	35.33	0.330	85.26
0.048	0.69	0.083	2.70	0.118	6.52	0.153	12.48	0.188	20.89	0.233	35.72	0.340	91.87
0.049	0.72	0.084	2.79	0.119	6.66	0.154	12.68	0.189	21.17	0.234	36.10	0.350	98.77
0.050	0.76	0.085	2.87	0.120	6.80	0.155	12.89	0.190	21.45	0.235	36.49	0.360	105.98
0.051	0.80	0.086	2.96	0.121	6.94	0.156	13.10	0.191	21.73	0.236	36.88	0.370	113.49
0.052	0.84	0.087	3.04	0.122	7.09	0.157	13.31	0.192	22.01	0.237	37.27	0.380	121.32
0.053	0.88	0.088	3.13	0.123	7.23	0.158	13.52	0.193	22.30	0.238	37.66	0.390	129.46
0.054	0.92	0.089	3.22	0.124	7.38	0.159	13.74	0.194	22.59	0.239	38.06	0.400	137.92
0.055	0.97	0.090	3.31	0.125	7.53	0.160	13.96	0.195	22.88	0.240	38.46	0.410	146.70
0.056	1.01	0.091	3.40	0.126	7.68	0.161	14.18	0.196	23.18	0.241	38.86	0.420	155.81
0.057	1.06	0.092	3.50	0.127	7.83	0.162	14.40	0.197	23.48	0.242	39.26	0.430	165.25
0.058	1.10	0.093	3.59	0.128	7.99	0.163	14.62	0.198	23.78	0.243	39.67	0.440	175.02
0.059	1.15	0.094	3.69	0.129	8.15	0.164	14.84	0.199	24.08	0.244	40.08	0.450	185.14
0.060	1.20	0.095	3.79	0.130	8.30	0.165	15.07	0.200	24.38	0.245	40.49	0.500	240.93
0.061	1.25	0.096	3.89	0.131	8.47	0.166	15.30	0.201	24.69	0.246	40.91		
0.062	1.30	0.097	3.99	0.132	8.63	0.167	15.53	0.202	24.99	0.247	41.32		
0.063	1.36	0.098	4.10	0.133	8.79	0.168	15.77	0.203	25.30	0.248	41.74		
0.064	1.41	0.099	4.20	0.134	8.96	0.169	16.00	0.204	25.62	0.249	42.17		
0.065	1.47	0.100	4.31	0.135	9.13	0.170	16.24	0.205	25.93	0.250	42.59		
0.066	1.53	0.101	4.42	0.136	9.30	0.171	16.48	0.206	26.25	0.251	43.02		
0.067	1.58	0.102	4.53	0.137	9.47	0.172	16.72	0.207	26.57	0.252	43.45		
0.068	1.64	0.103	4.64	0.138	9.64	0.173	16.97	0.208	26.89	0.253	43.88		
0.069	1.70	0.104	4.75	0.139	9.82	0.174	17.21	0.209	27.22	0.254	44.31		
0.070	1.77	0.105	4.87	0.140	10.00	0.175	17.46	0.210	27.54	0.255	44.75		
0.071	1.83	0.106	4.99	0.141	10.17	0.176	17.71	0.211	27.87	0.256	45.19		
0.072	1.90	0.107	5.10	0.142	10.36	0.177	17.96	0.222	31.65	0.257	45.63		

Additional Monitoring Faro Creek Seepage Losses December 2005

26 Dec. 19 2005 FCS-1 · unable to download -12:30 pm PH 736 6.05 pattery and computer and 4160 us/cm pozen (-20 °C) TDS 1910 mg/L HN.045m · It measured it noth a wire -02 samples; 12, 2× 250 mc ~ 8cm . Trred to D/L but bettery and computer poren -T, 29.47 21.02 backst, Tz 26.26 N 20 0r 19.5 L · bucket (20 L Pail brought with us) T1 5.12 t4 4.61 Tz 28.70 due to stant T2 4.95 T5 4.55 T3 4.59 · FCS-2, 1000 1:00 pm. Pipe frozen full on Dis FCS\_3 pH 5,88 2130pm end, no discharge through 1010 5120 us/cm TOS 2540 ms/cm culvert. TO 0.900 · FCS \_ 4 PH- 5,85 1:30pm · samples; 12, 2×250 GND 4100 1- 10 · bucket 20L TDS 1880 TI 7.81 T4 8.15 Tº ( 1.80 Tz 8.01 T5 8.20 · chop out, Smile, T3 8.29 gree from and white

·28 Dec. 19,2005 PH 5.84 COND 3590 FCS-6 4:30.7 pH 3.88 COUP 4170 US/CM TDS 1640 TR -0.5° TDS 2080 mg/L TC -0.10 , weir bullchead to frozen · site is heavily glacrated, solid aufre all over The place - onto old tailings. main flow free water on the edges, barely moving but rejentely some flow, <. 5 Lisec now channelized to LB near the top of the berm. · samples; 1L, 28250 mc . re-set the PT2x cable on a new pole- danger of being ited over. ' chop a channel to get samples and to bucket flow (small 91 bucket, white) T, 3.05 (-9.5 cm from Tz 3.04 brim) Tz 3.05 - note FCS-5 is Blind Duplicate ·samples; ZXIL, 4×250mL



1988 Triumph Street, Vancouver, BC Canada V5L 1K5 Tel: 604-253-4188 Toll Free: 1-800-665-0243 Fax: 604-253-6700 #2 - 21 Highfield Circle SE, Calgary, AB Canada T2G 5N6 Tel: 403-214-5431 Toil Free: 1-866-722-6231 Fax: 403-214-5430 #2 - 8820 100th Street, Fort St. John, BC Canada V1J 3W9 Tel: 250-785-8281 Fax: 250-785-8286

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Appendix C ETA Flood Hydrology Estimation of a Design Hydrograph for the Proposed ARD Collection System at the Emergency Tailings Area - DRAFT

# 1 Introduction

Mine drainage in the old Faro Creek channel below the Main Waste Dump contains elevated metal and sulphate concentrations. Owing to a steady increase in the sulphate concentrations since about 1998, it has been recognized that this drainage may pose a significant risk to the environment long before permanent closure measures can be implemented at the mine. To minimize the impact of this drainage over the short term, a plan has been proposed to collect the drainage and treat it at new water treatment plant. Implementation of this plan will require the development of a collection system, comprising a dam constructed in the old Faro Creek channel, pumps and a pipeline to the new water treatment plant. This memorandum summarizes the hydrological information used as the basis for sizing of the collection system.

Figure 1 shows key information that was assembled to help estimate design flows for the proposed collection system. It is a map of the pre-mining topography with the following pieces of information superimposed on it:

- 1) the locations of the two water quality monitoring sites within the old Faro Creek channel, one at the toe of the Main Dump (X23) and the other near the outlet of the Faro Creek canyon (X7);
- 2) outlines of the waste dumps around the perimeter of the Main Pit (excluding the Faro Valley Waste Dump);
- 3) outline of the lake within the Main Pit;
- 4) the course of the Faro Creek prior to development of the mine;
- 5) location of the Emergency Tailings Area (ETA); and,
- 6) the outline of the approximate area contributing flows to the old Faro Creek channel under present-day conditions.

The dam for the proposed collection system would be located near the downstream end of the ETA, or just above Station X7. The deposited tailings within the ETA would probably have to be removed to develop an adequate volume of storage for the collection system.

The catchment boundary outlined on Figure 1 was based on pre-mining topography rather than the surface topography of the waste dumps. Pre-mining topography generally provides an accurate basis for predicting where seepage from a waste dump will emerge (i.e., at points where the toes of waste dumps intersect stream channels). Following from this observation, the pre-mining topography was judged to also be useful for outlining the catchment area contributing to the proposed dam (after allowance was made for the effect of excavating the open pit). The tacit assumption in using pre-mining topography for this purpose is that percolation through a waste dump is primarily a vertical process and the underlying original ground is where the percolated water is forced to move laterally. The catchment outlined on Figure 1 suggests that Station X7, and hence the proposed dam, controls an area of about 1.8 km<sup>2</sup>.

The adopted design event for the proposed collection system was an extremely wet year with a return period of 100 years. The estimation of such an event was not a straightforward exercise for the following reasons:

- 1) the records of flow within the old Faro Creek channel are sparse, with continuous monitoring of flow only being established in September 2005;
- 2) owing to a large drainable porosity, the waste dumps have considerable, but unquantified, capacity to attenuate the runoff response from the catchment of the proposed dam;
- 3) there is uncertainty as to the true size of the drainage area that will be controlled by the proposed dam (e.g., it is unknown whether pre-mining or existing surface topography provides a better indication of the location of drainage divides); and,
- 4) there is uncertainty in the true long-term average yield generated by the waste dumps. Evaporation from the waste dumps may be enhanced relative to undisturbed catchments because of heat generated by oxidation within the waste dumps. In addition, the lack of vegetation on the waste dumps may allow a loss of snowpack due to wind re-distribution.

Taking the complications outlined above into consideration, a procedure was developed to estimate the 100-year hydrograph for the proposed collection system. The procedure was broken down into two broad tasks. The first involved using regional streamflow records to infer what the 100-year hydrograph would look like at Station X7 if no waste dumps existed within the station's catchment (i.e., a condition representative of the largely undisturbed catchments that are measured by the regional streamflow gauging stations). The second task entailed modifying the hydrograph developed in the first task so that it reflected the significant storage attenuation caused by the waste dumps. This second task relied heavily on the available flow measurements made thus far at X23 and X7. The two tasks are described below under separate headings.

## 2 Flood Hydrology of Typical Streams

The proposed collection system will include a reservoir to temporarily store a portion of incoming flood flows for subsequent pumping to the mill for treatment during periods of low flow. To determine an adequate size for the reservoir, an understanding must be developed of the volumes of water associated with the flood. With this in mind, this section uses regional streamflow gauging data to infer what the 100-year hydrograph would look like at the proposed dam, but without the attenuation benefits caused by storage within the waste dumps. As mentioned above, the influence of the waste dumps on the flood hydrology will be examined in the next section of the memorandum. The developed hydrograph spans a full calendar year and is based on a daily time step.

The volume characteristics of local floods were estimated using a technique known as Regional Analysis. This technique involved developing empirical relationships that could be used to transpose the flood data from regional streamflow gauging stations to the site of the proposed dam. Application of the Regional Analysis entailed six steps. The first step was the assembly of regional data. Emphasis was placed on finding streamflow gauging stations that had long periods of record and that were located on small drainage areas. To maximize the amount of data available from which to choose, a search was made of the networks of streamflow gauging stations operated by three government agencies: Water Survey of Canada (WSC), Environment Yukon (EY) and United States Geological Survey (USGS). The search for data in the WSC and EY networks extended over the entire Yukon Territory south of latitude 65°. The search within the USGS network was limited to the eastern central region of Alaska. Examination of the three networks revealed a total of 15 stations that could potentially be useful in characterizing the flood hydrology of the collection system. Table 1 provides details of these stations, including length of record, drainage area, mean annual runoff and the name of the authority that operated the station.

The second step entailed a statistical analysis of the assembled records. From each streamflow record, a total of 12 annual series were extracted. All of the series had one characteristic in common: they contained a list of the highest discharge in each year. The differences in the 12 annual series related to the period over which the highest discharge was defined. These periods were 1, 2, 3, 5, 7, 10, 15, 30, 60, 90, 183 and 365 consecutive days. Each of these annual series was fitted to a theoretical frequency distribution to estimate the average flood flow rate for a return period of 100 years. This meant a total of 180 fittings were undertaken (i.e., 15 stations x 12 annual series per station). Table 1 summarizes the results obtained from performing this step. To facilitate comparison of the floods generated by the widely different catchment sizes, the flood values in Table 1 are expressed as unit discharges in units of L/s/km<sup>2</sup> (i.e., the absolute flood discharge was divided by the contributing catchment area). The extraction of the annual series and the fitting of frequency distributions were performed using a suite of hydrological programs developed by the USGS (viz., ANNIE4.1, IOWDM4.1 and SWSTAT4.1).

The third step involved examining the data for trends that could form the basis for transposing the regional data to the minesite. For peak instantaneous floods, there is a tendency for the unit flood discharge to exhibit an inverse relationship with catchment area (i.e., unit flood discharge increases with decreasing catchment area). It was suspected that this inverse relationship may also apply to peak daily average flows and perhaps even longer durations. Figures 2 and 3 were prepared to test the flood data for such scale effects. These figures show a total of 12 plots of unit flood discharge vs. catchment area, one for each of the durations given in Table 1. Examination of these 12 plots revealed that unit discharge is virtually independent of catchment area for all 12 durations, at least over the range of catchment areas (13.7 km<sup>2</sup> to 7250 km<sup>2</sup>) represented by the data in Table 1. This observation suggested that there was no need to adjust the unit flood values from large catchments to make them representative of the floods on small minesite catchments.

Having discovered that scale effects are minimal for durations of a day and longer, a search was undertaken to identify another independent variable that might help explain the variation in unit flood values. This search identified mean annual runoff (MAR) as a potential variable, particularly for the longer durations. Figures 4 and 5 were developed

to explore the relationship between unit flood discharge and MAR. Again, a total of 12 plots are presented on these figures, one for each duration. Examination of these plots revealed that MAR is a poor predictor of flood values for short durations up to 5 days, a fair predictor for durations between 7 and 15 days, and a good predictor for durations of 30 days and longer. On the basis of this observation, it was decided to use the observed relationships between flood discharge and MAR to help estimate flood magnitudes at the minesite. To facilitate application of the relationships, a linear regression was fitted to the data set of each duration (see red lines on the plots). A text box on each plot expresses the relationship as an equation. To provide a means of making conservative flood estimates, envelope curves were also drawn on the plots that encompass all the data points (see blue lines).

The fourth step in developing the Regional Analysis was to estimate the MAR of the catchment of the proposed collection dam. This was accomplished using an empirical relationship developed during preparation of the 1996 Integrated Comprehensive Abandonment Plan for the Anvil Range Mine Complex. Figure 6 is a reproduction of this relationship. It examines the relationship between MAR and the elevational characteristics of the catchment that generated the runoff. The variable used as a measure of catchment elevation was the median elevation, which is the contour that divides a catchment into halves. Given a median elevation of about 1160 m, the catchment of the proposed collection dam has an estimated MAR of 205 mm. For a catchment area of 1.8 km<sup>2</sup>, this is equivalent to a long-term average flow of 11.7 L/s.

The fifth step involved estimating short and long duration flood flows for the proposed collection dam on old Faro Creek, but without accounting for the benefit of storage attenuation caused by the dumps. Using the estimated MAR of 205 mm, each of the plots on Figures 4 and 5 were entered to provide estimates of the 100-year flood magnitude for durations from 1 to 365 days. In recognition of the scatter in the plots, two values were extracted from each plot, one based on the best-fit linear regression and the other on the envelope curve. Table 2 summarizes all the flood values extracted from the plots, all expressed in normalized units of L/s/km<sup>2</sup>.

The sixth and final step entailed constructing 100-year hydrographs from the flood values assembled in Table 2. This was primarily a process of splitting the long duration flood flows presented in Table 2 into daily values. For example, Table 2 indicates the flood flows for one day and two consecutive days are 189 and 169 L/s/km<sup>2</sup>, respectively. This means the flood would comprise one day with an average flow of 189 L/s/km<sup>2</sup> and a second day with an average flow of 149 L/s/km<sup>2</sup> (i.e., 2 x 169 – 189 = 149). By repeating this process for the remaining long duration flows, a set of 365 daily values were computed from the flood information presented in Table 2. In most cases, this resulted in multiple days with the same discharge rate. For example, Table 2 provides flow values for periods of 5 and 7 consecutive days of 123 and 107 L/s/km<sup>2</sup>, respectively. As a result, the 6<sup>th</sup> and 7<sup>th</sup> highest flow days during the year were assumed to experience the same flow rate of 67 L/s/km<sup>2</sup> (i.e., (7 x 107 – 5 x 123) / 2 = 67).

Two hydrographs were constructed, one based on the "best estimate" values of unit discharge and the other on the "conservative estimate" values. Figure 7 graphically portrays the resulting two hydrographs. Significant features of the hydrographs are as follows:

- the hydrographs possess a daily time step and cover the period from January 1 to December 31;
- the peak daily discharge occurs on June 1 (based on the approximate average date on which the peak is observed to occur at streamflow gauging stations in the region);
- the unit flood values presented in Table 2 were multiplied by 1.8 km<sup>2</sup> to obtain absolute discharge rates representative of a typical stream with the same drainage area as the proposed collection dam; and,
- the hydrographs were given a "symmetrical" shape.

To create the "symmetrical" shape, the flow rate was made to progressively increase towards the peak and then progressively decrease away from it. This was accomplished by placing the daily flows, from largest to smallest, in an alternating pattern about the peak. Accordingly, the largest daily flow was assigned to June 1, second largest to May 31, the third largest to June 2, the fourth largest on May 30, etc. Beyond the 60<sup>th</sup> largest flow, this placement pattern was modified to create a skewed appearance to the hydrograph to approximate the shape of natural hydrographs in the region.

The 100-year "best estimate" hydrograph has a daily average peak of 341 L/s, which corresponds to an equivalent depth of 16 mm/d spread uniformly over the 1.8 km<sup>2</sup> catchment. The similar numbers for the 100-year "conservative estimate" hydrograph are 585 L/s and 28 mm/d. The 100-year "best estimate" and "conservative estimate" hydrographs have annual average flows of 22 L/s and 30 L/s, respectively. As a comparison, the long-term average yield from the collection dam catchment is estimated to be about 11.7 L/s (as computed above in Step 4).

The hydrographs presented in Figure 7 are approximations of what a 100-year wet year might look like on a "typical", natural stream with a 1.8 km<sup>2</sup> catchment. Flows measured at X23 and X7 suggest that the hydrology of the old Faro Creek is more subdued than a typical stream in the region. The next section describes how the estimated hydrograph for a typical stream was modified to represent conditions within the old Faro Creek channel.

# 3 Effect of Waste Dumps on Flood Hydrology

Figure 8 shows a plot of all known flow measurements made at the toe of the Main Waste Dump (Station X23) since 1987 (excluding the data collected at the recently established Station FCS-1). To provide a comparison with a "typical" stream in the region, the corresponding flow record for Vangorda Creek (EY Station 29BC003) has been superimposed on this plot. As the Vangorda Creek station is operated only during the open water season, missing data were patched using a correlation with another regional

streamflow gauging station (WSC Station 09BA001). Comparison of the X23 and Vangorda Creek records demonstrates that the runoff response in the old Faro Creek channel is very subdued. Baseflows in the winter are high and peaks during the spring freshet are comparatively small.

The subdued nature of the runoff response in the old Faro Creek channel can almost certainly be attributed to the waste dumps that occupy much of the catchment. As described in the introduction, the mouth of Faro Creek canyon (near Station X7) controls a drainage area of about  $1.8 \text{ km}^2$  (see Figure 1). The volume of waste rock overlying this catchment area is approximately 43 million m<sup>3</sup>, corresponding to an average dump height of 24 m. If the average moisture content of this dump was, say, 10% by mass, then the water contained in the dump would represent an equivalent depth of 3 m. With an estimated average yield from the dump of 0.2 m per year, the average retention time in the dump works out to be about 15 years. A retention time of this order is adequate to explain the subdued nature of the runoff response in the old Faro Creek channel.

A partial validation check of the X23 flow record was made by creating a scatter diagram with the coincidental Vangorda Creek flows (see bottom plot on Figure 8). The main aim of this diagram was to check the reasonableness of the largest flow on record (29 L/s on June 9, 1992). The scatter diagram indicates that this value is indeed plausible, as it corresponds with a high flow in Vangorda Creek. Examination of other regional data provides additional support for this conclusion. At the WSC station on Ross River, the average flow in June 1992 was exceptionally high; only one other June in the station's 42 year record has experienced a higher flow.

Besides serving as a consistency check, the scatter diagram provided some insight into the nature of the runoff response in the old Faro Creek channel. During average and dry years, the runoff response has a very strong groundwater character. During wetter years, such as 1992, a faster runoff component is also evident, possibly due to the temporary initiation of shallow groundwater flow or saturated overland flow.

The subdued nature of the X23 flow record suggests that flood hydrographs created for typical streams (as shown on Figure 7) probably overestimate the magnitude of flood discharges in the old Faro Creek channel, particularly for shorter durations of up to say 15 days. Because of this, a method was sought to modify the flood hydrographs estimated in Section 2 so they would be more representative of actual conditions below the waste dumps. The basic requirement of the method was it had to account for the storage attenuation within the waste dumps. One option would be to employ a common technique used in hydrological models to simulate lag and attenuation of runoff within a catchment, namely: cascading linear reservoirs. However, this method was rejected because the available flow data at the proposed dam site was judged to be insufficient to calibrate the technique. In recognition of the limited database, a simplified method was adopted. The basic premise of the method was that the 100-year flood hydrograph in the old Faro Creek could be represented as a mix of a typical stream response and a perfectly regulated system. The former response was estimated in Section 2. The latter response

assumes that the system has so much internal storage that the outflows from the system are nearly constant year-round.

The adopted method for accounting for storage attenuation within the waste dumps required calibration. Figure 9 graphically illustrates the calibration process. The data used for the calibration comprised flow measurements collected at a new gauging station (FCS-4) located near Station X7 and the proposed collection dam. The first flow measurement at this site was made in the fall of 2004. In September 2005, a triangular weir was established and an automated water level recorder installed. Blue box symbols on Figure 9 represent the 7 direct discharge measurements made at FCS-4 during 2005. The solid blue line represents the hourly flows computed from the water level measurements made at FCS-4. For the purpose of the calibration, the flow measurements made at FCS-4 during 2005 were used to roughly represent the long-term average flows from the waste dump. However, there is some evidence that these flows may actually represent higher-than-average conditions. For instance, the total precipitation measured at the Whitehorse Airport during the 2005 water year (October 2004 to September 2005) was 139% of normal.

The red line on Figure 9 shows the first attempt at reproducing the average flows at the proposed dam site. The key assumptions behind the construction of this line are as follows:

- the pre-mining topography provides a reliable basis for determining the drainage area controlled by the proposed collection dam (1.8 km<sup>2</sup>);
- the regional relationship between MAR and catchment median elevation provides a reliable estimate of the average yield generated by the waste dumps (205 mm);
- the Main and Northwest Waste Dumps have reached a steady-state moisture content so that there is no longer any net storage of water within the dumps; and,
- the shape of the seasonal runoff distribution can be approximated as being 50% of a typical streamflow distribution and 50% of a perfectly regulated distribution.

Figure 10 graphically illustrates the mechanics of constructing the red line. The average monthly flows measured at a WSC station on Tay River were used to represent the average distribution of a "typical" stream.

The flows represented by the red line on Figure 9 lie considerably above the measured flows at Station FCS-4 during 2005, and this is despite that fact that 2005 may have experienced greater-than-average flows from the dumps. After examining the underlying assumptions associated with development of the red line, the following potential reasons were hypothesized for the overestimation:

- 1) the true drainage area controlled by site FCS-4 could be less than determined using the pre-mining topography;
- the true yield of Station FCS-4's catchment could be less than estimated by the regional relationship between MAR and median elevation, particularly if waste dumps act to enhance evaporation losses above what is observed within natural catchments;

- 3) the moisture storage within the waste dump may not have reached a steady state; or,
- 4) some portion of the catchment yield may be flowing in the ground below Station FCS-4.

Much of the uncertainties outlined above can be addressed after a longer flow record has been measured at FCS-4. In the meantime, it was decided to recognize these uncertainties by defining ranges in which the true average monthly flows at FCS-4 would fall. The red line on Figure 9 was adopted as the upper limit on the true average flows. For a lower limit, the following assumptions were made:

- the drainage area remains the same as used for the red line, or 1.8 km<sup>2</sup>;
- the catchment yield is 100 mm, or roughly half the value estimated by the regional relationship between MAR and elevation; and,
- the distribution of flows can be approximated as being 30% of a typical stream distribution and 70% of a perfectly regulated distribution.

The orange line on Figure 9 represents this lower limit.

The analysis presented on Figure 9 explored ways of accounting for the storage attenuation caused by the waste dumps during average (or near average) flow conditions. The understanding of the waste dump hydrology gained from preparing Figure 9 was used in developing 100-year hydrographs for the proposed collection dam. Figure 11 presents the results. As was done for average flows, upper and lower limits are presented that are estimated to contain the true 100-year hydrograph at the site. The upper limit is based on the same assumptions used in developing the red line on Figure 9 (i.e., MAR = 205 mm, drainage area =  $1.8 \text{ km}^2$  and a 50:50 weighting of a typical streamflow hydrograph and a perfectly regulated hydrograph). Similarly, the lower limit is based on the same assumptions used to define the orange line (MAR = 100 mm, drainage area =  $1.8 \text{ km}^2$  and a 30:70 weighting). In both cases, the 100-year hydrograph for a typical stream is represented by the "best estimate" hydrograph developed in Section 2 (see Figure 7).

The "high estimate" 100-year hydrograph has a daily peak of 182 L/s and an annual average flow of 22 L/s. The similar values for the "low estimate" hydrograph are 114 L/s and 16 L/s.

#### Table 1 Estimated 100-Year Floods at Regional Streamflow Gauging Stations

Strear	nflow Gauging Station	Length of Record	Drainage Area	Mean Annual Runoff	Authority <sup>c</sup>	y <sup>c</sup> Average discharge in L/s/km <sup>2</sup> for the following number of consecutive days <sup>d</sup> :											
ID No.	Name	(years)	(km²)	(mm)		1	2	3	5	7	10	15	30	60	90	183	365
10AB003	King Creek at km 20.9 Nahanni Range Road	12	13.7	290	WSC	150	126	112	101	93	86	77	72	63	52	30	16
29AB006	Upper Wolf Creek <sup>a</sup>	9	14.5	179	EY	325	257	197	150	122	103	85	68	56	47	31	16
15344000	King Creek near Dome Creek	7	15.2	100	USGS	114	112	110	100	91	73	65	41	21	13	9	5
15535000	Caribou Creek near Chatanika	15	23.8	200	USGS	169	140	115	89	72	69	65	58	42	34	22	13
15439800	Boulder Creek near Central	20	81.0	131	USGS	257	246	199	151	132	112	89	60	43	31	15	8
29BC003	Vangorda Creek at Faro Townsite Road <sup>a</sup>	22	91.2	235	EY	149	134	113	96	85	72	71	61	43	32	20	11
09AD002	Sidney Creek at km 46 South Canol Road	11	372	350	WSC	231	200	173	141	126	116	103	86	62	50	29	18
10AA002	Tom Creek at km 34.9 Robert Campbell Highway	18	435	218	WSC	105	103	102	100	97	90	82	76	57	51	29	16
09AG003	South Big Salmon River below Livingstone Creek	14	515	246	WSC	198	175	157	132	112	100	84	72	53	43	25	15
09AA012	Wheaton River near Carcross <sup>a</sup>	49	875	285	WSC	107	105	101	92	88	81	70	56	44	37	24	13
15511000	Little Chena River near Fairbanks <sup>b</sup>	37	963	199	USGS	256	242	226	189	154	123	94	58	38	29	21	11
09BB001	South MacMillan River at km 407 Canol Road	22	997	624	WSC	216	216	205	187	173	169	151	126	103	79	47	25
09EA004	North Klondike River near the mouth	29	1100	379	WSC	159	145	135	124	113	104	97	82	64	51	32	18
15484000	Salcha River near Salchaket <sup>b</sup>	56	5618	261	USGS	285	239	217	179	152	131	107	77	52	44	29	15
09BA001	Ross River at Ross River <sup>a</sup>	41	7250	293	WSC	113	111	109	103	97	88	81	72	52	39	22	13

Notes: a) The daily streamflow records of these four stations were patched prior to them being fitted to frequency distributions. Estimates were only made for short gaps and/or missing periods during winter. Years with substantial periods of missing data during the open water season were not included in the flood frequency analysis.

b) For each station and each duration, the annual series of flood data were fitted to a theoretical frequency distribution (Log-Pearson Type III or 3-parameter lognormal) to estimate the magnitude of the 100-year flood discharge. A visual inspection revealed a good fit to the data sets for all but a few of the stations. For the Salcha River and Little Chena River, the fit was only fair for durations from 1 to 30 days because of the existence of a high outlier.

c) WSC = Water Survey of Canada; EY = Environment Yukon; USGS = United States Geological Survey

d) To facilitate comparisons of the flood values for the widely differing catchment areas, the flood values in this table have been expressed as unit discharges in units of L/s/km<sup>2</sup> (i.e., the absolute flood discharges have been divided by the contributing catchment areas).
## Table 2 Estimated 100-Year Flood Hydrograph for a Typical Stream at the Minesite

Mothod of Estimation <sup>a</sup>		Average discharge in L/s/km <sup>2</sup> for the following number of consecutive days <sup>b</sup> :											
Method of Estimation	1	2	3	5	7	10	15	30	60	90	183	365	
Linear regression fitted to trend between flood discharge and MAR	189	169	148	123	107	93	80	62	45	36	22	12	
Envelope curve on trend between flood discharge and MAR	325	258	226	190	155	124	99	74	59	50	33	17	

Notes: a) The estimate based on linear regression is designated as being the "best estimate" hydrograph. The other is designated as being the "conservative estimate" hydrograph.

b) These unit flood estimates are for a stream with a mean annual runoff (MAR) of 205 mm.





Figure 2: Magnitude of 100-year Flood vs. Catchment Area (1 to 10 Day Durations)



Figure 3: Magnitude of 100-year Flood vs. Catchment Area (15 to 365 Day Durations)





Figure 5: Magnitude of 100-Year Flood vs. Mean Annual Runoff (15 to 365 Day Durations)



Figure 6: Regional Relationship Between Mean Annual Runoff and Catchment Median Elevation



Figure 7: Estimated 100-Year Flood Hydrograph at Proposed Collection Dam (without allowance for effect of waste dumps)



Figure 9: Comparison of Estimated Average Flows and Observed Flows at FCS4





Figure 11: Estimated 100-Year Flood Hydrograph at Proposed Collection Dam (with allowance for effect of waste dumps)

Appendix D Hydraulic Testing Results

Appendix D-1 Packer Testing Data Sheets











Appendix D-2 Pumping Test Discharge Data

Well No:	SRK 04-04 (Pumping Well)		
Static WL: 9.05h	6.904m		
Date: October 2, 2005	Start at 9.22h		
Elapsed Time (minutes)	Manometer Reading	Caculated Flow (USgpm)	
0.25	7.155	52	
0.45	7.915	51	
1.5	7925	51	
1.75	7.865	50	
2	7.94	52	
3.5	0.00	96	??
4	8.025	53	
4.5	8.04	53	
6	7.98	50	
7	7.98	50	
8	7.984	50	
9	7.99	50	
12	8.006	50	
14	8.016	50	
16	8.024	49	
20	8.042	49	
25	8.06	49	
30	8.077	49	
40	[8.145]	[50]	
50	8.16	50	
60	8.195	50	
70 80	8.222	50	
90	8.251	50	
100	8.278	50	
120	8.322	50	
160	8.41	50	
180	8.443	50	some uncertainty
210	8.502	50	
240	8.546	50	Flow oscillating between 50-51 gpm
300	8.635	50	· · · · · · · · · · · · · · · ·
334	8.672	50	
360	8.733	50	
420	8.762	50	
450	8.785	50	
480	8.81	50	
540	8.858	50	
570	8.877	51	
600	8.897	51	
660	8.933	51	
690	8.94	51	
720	8.97	51	
750	8.976	51	
810	8.981	51	
840	8.944	50	
870	9.005	50	
930	9.027	50	
960	9.04	50	
990	9.045	50	
1050	9.054	50	
1080	9.076	50	
1110	9.085	50	
1140	9.097	50 50	
1200	9.12	50	
1230	9.124	50	
1260	9.133	50	
1320	9.153	50	
1350	9.163	50	
1380	9.172	50	<u> </u>
1410	9.172	49	
AVERAGE		50.96	
1	1	1	1

Appendix D-3 24-Hour Pumping Test Drawdown Data and Interpretations













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Appendix E Water Quality

Sample ID							X2	3							
Date Sampled	1/12/2004	2/16/2004	3/15/2004	4/14/2004	5/14/2004	6/14/2004	7/12/2004	8/9/2004	9/13/2004	10/12/2004	11/14/2004	5/9/2005	6/20/2005	7/25/2005	8/22/2005
Nature															
Physical Tests															
Field Conductivity			coool	4190	6100	9100	7150	0000	0200	0.460	8040	7550	11500	6200	7150
Hardness CaCO3		Г	4210	4150	0100	0190	7130	3980	9290	4023	4290	3840	4660	4300	4970
pH		-													
pH F		7	7	6.9	6	6.5	6.5	6.7	6.4		[	6	6.4	6.6	6.7
Disselved Aniens															
Acidity (to pH 8.3) CaCO3															
Alkalinity-Total CaCO3															
Bromide Br															
Chloride Cl															
Sulphate SO4	4190	3163	4440	4390	5610	5590	6370	6490	6500	5630	5380	5910	6510	5500	5800
		L.													
Total Metals	0.001	0.005	0.000		0.00	0.000	0.010	0.010	0.017	0.000	0.000	0.15	0.04	0.000	
Aluminum I-Al Antimony T-Sh	<0.001	0.005	0.009	0.14	9.38	0.063	<0.012	<0.012	<0.017	<0.009	<0.029	8.15	0.04	0.009	<0.005
Arsenic T-As	0.012	0.013	0.002	<0.03	0.022	0.004	0.004	0.002	0.002	0.002	0.003	0.019	0.004	0.002	0.005
Barium T-Ba	0.015	0.014	0.015	0.005	0.023	0.021	0.02	0.016	0.016	0.015	0.019	0.021	0.017	0.016	0.016
Beryllium T-Be	0.0007	0.0005	< 0.001	<0.003	0.008	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	0.004	< 0.001	< 0.001	< 0.001
Bismuth I-Bi Boron T-B	<0.01	<0.05	<0.001	<0.01	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001
Cadmium T-Cd	0.0172	0.0164	0.015	0.02	1.2	0.16	0.104	0.083	0.058	0.06	0.064	0.847	0.211	0.134	0.132
Calcium T-Ca	655.6	611.8	608	564	382	550	508	490	460	472	483	444	495	469	578
Chromium T-Cr	< 0.001	0.005	0.001	< 0.01	0.004	<0.001	<0.001	<0.001	<0.001	< 0.001	< 0.001	0.004	< 0.001	< 0.001	< 0.001
Copper T-Cu	0.396	0.365	0.021	0.36	1.43	0.26	0.91	0.085	0.075	0.046	0.86	1.16	0.14	0.87	0.063
Iron T-Fe	9.026	12.849	15.3	30.2	58.4	123	169	151	139	137	110	86.7	190	140	167
Lanthanum T-La	<0.001	<0.001		?											
Lead T-Pb	0.019	0.011	<0.001	<0.03	0.19	0.008	0.003	0.003	0.005	0.003	0.002	0.11	0.005	0.003	0.002
Magnesium T-Mg	745.7	707.4	0.011	665	577	0.21 681	647	0.15	0.15	0.14	749	662	0.21	760	0.22
Manganese T-Mn	46.209	42.319	57.6	49.2	65.3	75.5	73.8	70.8	65.8	66.9	64.1	73.4	83.8	69.9	89
Mercury T-Hg	<0.0001	<0.0001	< 0.00002	< 0.00002	0.00004	< 0.00002		<0.00002	< 0.00002	<0.00002	< 0.00002	0.00004	< 0.00002	< 0.00002	< 0.00002
Molybdenum T-Mo	0.03	0.034	< 0.0005	< 0.02	< 0.0005	< 0.0005	< 0.0005	< 0.0005	<0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005
Phosphorus T-P	0.400	0.45	0.34	0.04	1.72	1.47	1.15	1	0.99	0.94	1.04	1.47	1.31	1.00	1.21
Potassium T-K	27.4	25.1	16.1	14.3	12.9	16	16.2	13.9	14.5	15.3	22.7	12.8	14.7	13.9	16.8
Selenium T-Se	< 0.005	<0.005	0.009	0.0077	0.053	0.017	0.014	0.011	0.008	0.007	0.015	0.027	0.017	0.01	0.018
Silicon T-Si	0.0040	0.0004	16.2	7.49	20.2	17.9	16.4	16.7	18.4	16.4	30.2	19.5	18.7	17	17.4
Silver T-Ag Sodium T-Na	63.5	55.7	<0.00025	<0.01	<0.00025	<0.00025	<0.00025	<0.00025	<0.00025	<0.00025	0.0009	<0.00025	<0.00025	<0.00025	<0.00025
Strontium T-Sr	4.017	3.542	3.69	3.37	2.33	3.35	3.58	3.46	3.6	3.35	3.32	2.45	3.55	3.37	4.09
Sulfur T-S	1472.9	1319.4													
Tellurium T-Te	0.000	0.000	< 0.001		< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001
Thailium T-Th	<0.002	<0.002	0.0011		0.0019	<0.0005	<0.0016	<0.0013	<0.0014	0.0015	<0.0016	0.0019	< 0.0005	<0.0015	<0.0017
Tin T-Sn	< 0.002	<0.002	<0.001	< 0.03	< 0.001	< 0.001	< 0.001	0.001	< 0.001	< 0.001	0.001	< 0.001	< 0.001	< 0.001	< 0.001
Titanium T-Ti	<0.001	0.003	0.002	<0.005	0.002	0.001	0.001	0.001	< 0.001	<0.001	0.005	0.001	<0.001	<0.001	0.001
Tungsten T-W	<0.03	<0.03	0.017	r	0.028	0.000	0.012	0.011	0.012	0.016	0.019	0.022	0.0001	0.01	0.012
Vanadium T-V	0.004	<0.001	<0.001	<0.01	<0.038	<0.009	<0.012	<0.001	<0.001	< 0.016	<0.018	<0.001	<0.0091	<0.001	<0.001
Zinc T-Zn	155.508	152.852	141	158	964	498	512	438	375	372	327	808	655	480	449
											-				
Discolved Metals															
Aluminum D-Al	<0.001	0.005	<0.005	0.13	0.4	<0.005	<0.005	<0.005	< 0.005	<0.005	<0.005	0.58	0.006	<0.005	<0.005
Antimony D-Sb	< 0.002	<0.002	<0.001	< 0.05	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001	< 0.001
Arsenic D-As	0.013	0.013	0.002	< 0.03	0.011	0.003	0.003	0.002	0.002	0.001	0.003	0.005	0.003	0.002	0.004
Barium D-Ba	0.013	0.013	0.013	<0.001	0.021	<0.016	<0.015	0.013	<0.013	<0.014	<0.018	0.018	0.014	0.015	<0.012
Bismuth D-Bi	<0.01	<0.01	<0.001	<0.003	< 0.001	<0.001	<0.001	< 0.001	< 0.001	<0.001	<0.001	< 0.001	<0.001	<0.001	< 0.001
Boron D-B	0.1	<0.05	<0.05	<0.01	<0.05	< 0.05	< 0.05	<0.05	<0.05	< 0.05	<0.05	< 0.05	<0.05	< 0.05	< 0.05
Cadmium D-Cd	0.0156	0.0155	0.013	< 0.01	1.16	0.131	0.083	0.07	0.05	0.054	0.052	0.734	0.177	0.121	0.104
Calcium D+Ca Chromium D+Cr	0.002	0.003	532 c0.001	499	243 <0.001	408	433	<0.001	<0.001	465	400	<0.001	433	434	408
Cobalt D-Co	0.387	0.389	0.35	0.320	1.340	0.95	0.77	0.67	0.75	0.73	0.82	1.06	0.91	0.8	0.85
Copper D-Cu	0.009	0.027	0.005	<0.02	6.56	0.052	0.03	0.014	0.01	0.006	0.006	4.08	0.027	0.034	0.012
Iron D-Fe	1.341	0.339	<0.01	<0.01	13.1	6.6	31.9	45	55.2	14.8	38.9	28.9	69.9	9.02	25.2
Lead D-Pb	0.013	0.015	<0.001	<0.03	0.005	<0.001	<0.001	<0.001	<0.001	< 0.001	<0.001	0.003	<0.001	< 0.001	< 0.001
Lithium D-Li	0.0.0		0.1		0.18	0.15	0.16	0.12	0.13	0.14	0.18	0.16	0.17	0.16	0.18
Magnesium D-Mg	717	758.3	579	614	512	587	560	530	638	695	711	572	692	689	687
Manganese D-Mn	44.714	43.355	49.9	-0.00002	57.6	62.3	-0.00002	59	59.9	65.5	63.4	64.5	70.1	64.3	69.2
Molybdenum D-Mo	0.03	0.035	<0.00002	<0.0002	< 0.0005	< 0.0005	< 0.00002	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005
Nickel D-Ni	0.403	0.478	0.47	0.46	1.62	1.21	0.94	0.83	0.9	0.87	1.03	1.34	1.1	0.96	0.95
Phosphorus D-P															
Potassium D-K Selenium D-Se	26	27.5	14	12.8	12.3	11.6	12.3	10.2	12.8	14.5	18.1	11.4	11.5	0.000	13.1
Selenium D-Se Silicon D-Si	<0.005	<0.005	12.9	5.7	15.7	11.1	11.3	11.7	12.2	11.6	16.8	0.025	12.3	12.8	10.2
Silver D-Ag	0.0049	0.0029	<0.00025	<0.01	<0.00025	< 0.00025	<0.00025	<0.00025	<0.00025	<0.00025	0.0005	<0.00025	<0.00025	<0.00025	<0.00025
Sodium D-Na	60.4	60.8	52.8	50	33	42.6	38.5	37.5	47.4	53.5	68.4	33.7	42.3	43	47.8
Strontium D-Sr Sulfur D-S	3.771	3.814	3.2	3.01	2.32	2.72	3.03	2.64	3.23	3.19	3.16	2.36	3.02	3.11	3.25
Tellurium D-Te	14/1.8	1406.1	<0.001	1	<0.001	<0.001	<0.001	< 0.001	<0.001	< 0.001	<0.001	<0.001	<0.001	<0.001	< 0.001
Thallium D-TI	<0.002	<0.002	0.001		0.0017	0.0012	0.0011	0.0009	0.0012	0.0013	0.0012	0.0016	0.0012	0.0012	0.0013
Thorium D-Th			0.0007		0.0006	< 0.0005	< 0.0005	<0.0005	<0.0005	<0.0005	< 0.0005	<0.0005	<0.0005	<0.0005	<0.0005
Tin D-Sn Titanium D-Ti	0.002	<0.002	<0.001	<0.03	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	< 0.001
Tungsten D-W	<0.001	<0.03	<0.001	~0.005	0.002	0.001	~0.001	<0.001	<0.001	L0.001	~0.001	<0.001	L0.001	<0.001	<0.001
Uranium D-U			0.014	í	0.0087	0.0011	0.0018	0.0016	0.0048	0.0019	0.0055	0.0084	0.0014	0.0039	0.0014
Vanadium D-V	0.003	0.002	<0.001	<0.01	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	<0.001	< 0.001	<0.001	< 0.001	< 0.001
Zinc D-Zn	148.051	157.429	121	134	857	417	446	338	340	351	326	/32	562	440	391

Sample ID	SRK04- 04	SRK04-04 (1HF	4-04 (1HR, 10HR, 24HR, 36HR) during Ptest (36-hr is dup)			SRK04-03A	SRK04-03B	P96-8	BA	P96-8B		
Date Sampled Time Sampled	5/5/2005 10:45	10/2/2005	10/2/2005 19:20	10/3/2005 9:20	10/3/2005	5/5/2005	5/5/2005	5/3/2005 16:35	9/10/2005	5/3/2005	9/10/2005	
Nature	Water	Water	Water	Water	Water	Water	Water	Water	Water	Water	Water	
Physical Tests Field Conductivity												
Lab Conductivity (uS/cm)		8390	7780	7610	7630			197	6370	5540	6620	
Hardness CaCO3 pH	3810 5.23	4.84	5.34	5.39	5.44	3970 5.87	3890 3.72	94.4 6.76	4240	3770	4080	
pH F												
Dissolved Anions												
Acidity (to pH 8.3) CaCO3		4640	4400	4890	4280							
Alkalinity-Total CaCO3 Bromide Br	63.6	67.3	95.4	93.5	108	91	61.7	12.4	108	214	131	
Chloride Cl		<50	<50	<50	<50							
Sulphate SO4	7080	8100	7460	7460	7370	5480	16700	71.2	5040	4520	4980	
Total Mateix												
Aluminum T-Al		2.32	1.84	1.62	1.67							
Antimony T-Sb		< 0.010	<0.010	< 0.010	< 0.010							
Barium T-Ba		0.010	<0.010	<0.010	<0.010							
Beryllium T-Be Bismuth T-Bi		<0.050	<0.050	<0.050	<0.050							
Boron T-B		<1.0	<1.0	<1.0	<1.0							
Cadmium T-Cd Calcium T-Ca		0.0278	0.0243	0.0279	0.0248							
Chromium T-Cr		<0.050	<0.050	<0.050	<0.050							
Cobalt T-Co Copper T-Cu		0.651	0.643	0.644	0.647							
Iron T-Fe		2410	2020	1980	1940							
Lantnanum T-La Lead T-Pb		0.0689	0.0737	0.0517	0.0531							
Lithium T-Li		<0.50	<0.50	<0.50	<0.50							
Magnesium I-Mg Manganese T-Mn		562 66.3	540 67.0	558 66.1	544 65.6							
Mercury T-Hg		0.0050	0.0050	0.0050	0.0050							
Nickel T-Ni		<0.0050	<0.0050	<0.0050	<0.0050							
Phosphorus T-P		<0.90	<0.90	< 0.90	<0.90							
Selenium T-Se		<0.10	<0.10	<0.10	<0.10							
Silicon T-Si Silver T-Ag		17.1	16.4	16.6	16.3							
Sodium T-Na		71.0	76.8	88.0	86.2							
Strontium T-Sr Sulfur T-S		3.93	4.05	4.21	4.16							
Tellurium T-Te												
Thallium T-TI Thorium T-Th		<0.010	<0.010	<0.010	<0.010							
Tin T-Sn		<0.010	<0.010	<0.010	<0.010							
Tungsten T-W		<0.030	<0.030	<0.030	<0.030							
Uranium T-U		0.0074	0.0072	0.0070	0.0074							
Zinc T-Zn		<0.10 469	<0.10 451	<0.10 450	<0.10 444							
Dissolved Metals												
Aluminum D-Al Antimony D-Sh	1.4	2.06	1.72	1.60	2.70	<1.0	4.3	0.029	<1.0	<0.50	<1.0	
Arsenic D-As	<0.10	<0.010	<0.010	<0.010	<0.010	<0.10	<0.10	<0.0010	<0.10	<0.050	<0.10	
Barium D-Ba Bervllium D-Be	<0.040 <0.010	0.0115 <0.050	0.0113 <0.050	0.0118 <0.050	0.0118 <0.050	<0.040 <0.010	<0.40 <0.10	0.042	<0.040 <0.010	<0.10 <0.025	<0.040 <0.010	
Bismuth D-Bi		<0.050	<0.050	<0.050	<0.050							
Boron D-B Cadmium D-Cd	<0.20 0.0151	<1.0 0.0242	<1.0 0.0246	<1.0 0.0260	<1.0 0.0246	<0.20 0.0066	<2.0 <0.0050	<0.10 0.000901	<0.20 0.220	<0.50 0.0410	<0.20 0.0970	
Calcium D-Ca	457	516	517	526	532	454	459	25.6	456	424	431	
Cobalt D-Co	0.500	0.631	<0.050	0.642	0.627	0.349	0.225	<0.00050	0.312	<0.025	<0.050	
Copper D-Cu Iron D-Fe	<0.10	0.022	0.025	0.024	0.023	<0.10	<0.10	0.0044	<0.10	<0.050	<0.10	
Lanthanum D-La	1030	2000	2020	1000	1000	093	0010	0.004	0.001	0.22	3.03	
Lead D-Pb Lithium D-Li	<0.10	0.0058 <0.50	0.0235 <0.50	0.0509 <0.50	0.0510	<0.10	0.16 <1 0	<0.0010	<0.10	<0.050	<0.10	
Magnesium D-Mg	649	558	546	546	551	690	666	7.39	752	659	730	
Manganese D-Mn Mercury D-Hg	56.0 <0.00020	64.6	65.5	65.2	64.1	49.7 <0.00020	72.9 <0.00020	0.014	98.7 <0.00020	49.0 <0.00020	90.7 <0.00020	
Molybdenum D-Mo	<0.10	< 0.0050	< 0.0050	< 0.0050	<0.0050	<0.10	<0.10	<0.0010	<0.10	<0.050	<0.10	
Phosphorus D-P	0.72	<0.90	<0.90	<0.90	<0.90	0.60	<0.50	0.0108	1.70	0.69	1.40	
Potassium D-K	-0.40	11.5	11.4	11.2	11.6	-0.40	-0.40	-0.0010	-0.40	-0.050	-0.10	
Silicon D-Si	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.0010	<0.10	<0.000	<0.10	
Silver D-Ag Sodium D-Na	<0.0050	<0.0010	<0.0010	<0.0010 86.2	<0.0010 86 5	<0.0050	<0.0050	<0.000050	<0.0050	<0.0025	<0.0050	
Strontium D-Sr	00.5	3.83	4.06	4.16	4.05	50.0	00	~z.0	54.9		50.0	
Sulfur D-S Tellurium D-Te												
Thallium D-TI	<0.020	<0.010	<0.010	<0.010	<0.010	<0.020	<0.020	<0.00020	<0.020	<0.010	<0.020	
Tin D-Sn		<0.010	<0.010	<0.010	<0.010							
Titanium D-Ti	<0.10	<0.030	<0.030	<0.030	<0.030	<0.10	<1.0	<0.050	<0.10	<0.25	<0.10	
Uranium D-U	<0.020	0.0061	0.0066	0.0074	0.0067	<0.020	<0.020	<0.00020	<0.020	<0.010	<0.020	
Vanadium D-V Zinc D-Zn	1.07	<0.10	<0.10	<0.10	<0.10	0.507	3.81	< 0.030	0.078	0.17	< 0.20	
2.117 0-211	350	461	447	444	438	∠33	/49	1.67	604	173	306	
Nutrients												

Sample ID		FC	S-1		FCS-2			FCS-3			FCS-4			
Date Sampled	10/20/2004	4/17/2005	7/1/2005	10/18/2005	10/20/2004	7/1/2005	10/18/2005	10/20/2004	4/17/2005	7/1/2005	10/18/2005	10/20/2004	4/17/2005	10/18/2005
Time Sampled				Water										
Physical Tests				Trator										
Field Conductivity	6470	5240	4610	4230	7810	3850	4890	7410	5710	5120	5010	7080	5190	4610
Lab Conductivity (uS/cm)	6320	6110		6990			6990	7470	6920		8400	6900	5770	6780
nH				6 37			6 55				5.17			5.45
pH F	6.71		6.7	6.21	6.5	5.78	7.47	5.96		5.58	6.07	6.69		6.82
-			•											
Dissolved Anions														
Acidity (to pH 8.3) CaCO3														
Alkalinity-Total CaCO3 Bromide Br				111			120				60.0			57.0
Chloride Cl				13			14				<10			15
Fluoride F				0.43			<0.40				0.55			<0.40
Sulphate SO4	5580	5030.0		6200			6210	5790	5550.0		6570	5490	4170	4860
Total Matala														
Aluminum T-Al				~0.40			0.58				1.52			7.00
Antimony T-Sb				<0.40			<0.40				<0.60			<0.40
Arsenic T-As				<0.40			<0.40				<0.60			<0.40
Barium T-Ba				<0.020			0.026				<0.030			0.049
Beryllium T-Be				<0.010			< 0.010				<0.015			< 0.010
Bismuth I-Bi				<0.40			<0.40				<0.60			<0.40
Cadmium T-Cd				0.102			0.112				<0.030			0.047
Calcium T-Ca				496			493				279			477
Chromium T-Cr				<0.020			<0.020				<0.030			<0.020
Cobalt T-Co				0.972			0.748				0.290			0.471
Iron T-Fe	123	88.0		0.036			0.030	1650	1210		<0.060	572	801	<0.050
Lanthanum T-La	123	00.9		137			01.0	1030	1210		1120	5/2	001	,13
Lead T-Pb				<0.10			0.30				<0.15			0.26
Lithium T-Li				0.199			0.216				0.078			0.145
Magnesium T-Mg				881			928				309			626
Manganese T-Min				90.4			104				31.1			69.3
Molybdenum T-Mo				<0.060			<0.060				<0.090			<0.060
Nickel T-Ni				1.12			1.27				0.33			0.69
Phosphorus T-P				<0.60			<0.60				<0.90			<0.60
Potassium T-K				16.7			18.2				6.9			14.2
Selenium I-Se Silicon T-Si				<0.40			<0.40				<0.60			<0.40
Silver T-Ag				<0.020			<0.020				< 0.030			<0.020
Sodium T-Na				60.6			64.4				87.4			98.5
Strontium T-Sr				3.67			3.88				2.12			3.31
Sulfur T-S														
Tellurium T-Te				~0.40			~0.40				<0.60			<0.40
Thorium T-Th				<0.40			<0.40				<0.00			<0.40
Tin T-Sn				<0.060			< 0.060				< 0.090			<0.060
Titanium T-Ti				<0.020			<0.020				< 0.030			<0.020
Tungsten T-W														
Uranium T-U				-0.060			-0.060				-0.000			-0.060
Zinc T-Zn	371	295		<0.000			<0.060	309	309		<0.090	319	174	<0.000
Lino i Lin	5/1	205					-100	303	303			515	174	010
Dissolved Metals														
Aluminum D-Al				<0.40			<0.40				<0.60			<0.40
Areopic D-As				<0.40			<0.40				<0.60			<0.40
Barium D-Ba				<0.020			<0.020				<0.030			<0.020
Beryllium D-Be				<0.010			<0.010				< 0.015			<0.010
Bismuth D-Bi				<0.40			<0.40				<0.60			<0.40
Boron D-B				< 0.20			<0.20				< 0.30	1		< 0.20
Caumium D-Cd				0.096 493			U.111 400				<0.030			0.045 482
Chromium D-Cr				<0.020			<0.020				<0.030			<0.020
Cobalt D-Co				0.965			0.755				0.292			0.478
Copper D-Cu				<0.020			<0.020				<0.060	1		<0.050
Iron D-He		34.4		119			44.7		1090		1090		604	670
Lead D-Pb				<0.10			<0.10				<0.15	1		<0.10
Lithium D-Li				0.199			0.219				0.073			0.153
Magnesium D-Mg				876			936				314	1		635
Manganese D-Mn				90.4			105				37.3			70.6
Mercury D-Hg Molybdenum D Mo				0.000							-0.000			-0.000
Nickel D-Ni				<0.060			<0.060				<0.090			0.000
Phosphorus D-P				<0.60			<0.60				<0.90			<0.60
Potassium D-K				16.8			18.4				6.8			14.5
Selenium D-Se				<0.40			<0.40				<0.60			<0.40
Silicon D-Si Silicor D-Ag				6.74			9.60				7.97			9.53
Sodium D-Na				<0.020			<0.020				<0.030			<0.020
Strontium D-Sr				3.51			3.94				2.09			3.33
Sulfur D-S														
Tellurium D-Te				-			-							
Thailium D-TI				<0.40			<0.40				<0.60			<0.40
Tin D-Sn				<0.060			<0.060				<0.000			<0.060
Titanium D-Ti				<0.020			<0.020				<0.030			<0.020
Tungsten D-W														
Uranium D-U														
Vanadium D-V Zinc D-Zn		070		<0.060			< 0.060		001		<0.090		400	< 0.060
Zinc D-Zn		278		470			462		291		219	1	150	309
Nutrients												l		

Appendix F Scoping Level Numerical Model Results


Steffen, Robertson and Kirsten (Canada) Inc. Suite 800 – 1066 West Hastings Street Vancouver, B.C. V6E 3X2 Canada

vancouver@srk.com www.srk.com

Tel: 604.681.4196 Fax: 604.687.5532

# **Revised Technical Memo**

То:	Faro – Groundwater Group	Date:	February 28, 2006
cc:		From:	Quinn Jordan-Knox, SRK
			Christoph Wels, RGC
Subject:	Results of ETA Scoping Simulations	Project #:	1CD003.073 Task 20e

A review of water quality data from the Emergency Tailings Area (ETA) by RGC and subsequent investigations by RGC and SRK indicated that highly contaminated groundwater and surface water are discharging to Rose Creek Valley (Preliminary Seepage Collection Options, SRK, 2005). This water is likely a combination of seepage from the main Faro waste rock dump and the mill area, and precipitation infiltrating through tailings present in the ETA. Initial results suggested that the ETA tailings likely contribute to the overall contaminant loading, in particular with respect to iron.

The 2005 report recommended a seepage collection system consisting of a line of pumping wells adjacent to the mine access road, combined with surface water collection sumps. Further works were undertaken in 2005 to increase the understanding of groundwater flow in the ETA to improve design and costing of the collection system. If further work indicates that groundwater collection above the road is difficult, an alternative seepage and surface water collection system will be necessary below the road.

A conference call was held on November 10<sup>th</sup>, 2005, to provide an update on the status of work on various 2005 studies at Faro, including the Emergency Tailings Area (ETA). Preliminary analyses were discussed and next steps identified to complete the project, focussing on information that will be critical for identifying and evaluating closure alternatives in the next few months. Major preliminary results of 2005 ETA investigations were:

- Results of a 24 hr pumping test of SRK04-4 indicated that hydraulic barriers were not intersected; all monitoring wells showed drawdown.
- The seepage face flow on the downstream side of the access road did not show a noticeable change (based on visual assessments and continuous flow monitoring at the mouth of the Faro Creek canyon) during the pumping test. ETA groundwater flow is thought to be connected to the seepage face but reasons for the lack of reduction in seepage during pumping were unclear.
- Flow and loading estimates along Faro Creek appeared to balance with the exception of iron. Zinc load estimates at FCS4 in Faro Creek were approximately 71 tonnes/yr

During the November 10 conference call, a number of issues were identified for further consideration:

- 1. Do numerical simulations support the conclusion that it is feasible to collect contaminated groundwater in the ETA with a pumping well system?
- 2. Do simulations indicate that the 24 hr pumping test should have cut-off seepage below the access road and if not, why?
- 3. What is the estimated capture efficiency for groundwater collection within the ETA?
- 4. Should the tailings be removed to reduce the contaminant load to the aquifer (particularly iron) and improve the operation of pumping system (reduce operation costs).

This memo describes the construction of a groundwater flow model for the ETA area and summarizes results of numerical scoping simulations of groundwater flow in the ETA and potential groundwater capture

scenarios for the collection of groundwaters at the mine access road. The numerical model was constructed to provide a tool for testing conceptual model assumptions and potential capture systems. Scoping simulation results are discussed with respect to the above identified issues.

# BACKGROUND

The geologic conceptual model for the ETA has been updated using geologic and geophysical data from the 2004 and 2005 field programs. Figures 1 and 2 are location maps for the site. Figures 3 and 4 are cross-sections showing the primary aquifer unit, overlying tailings deposit, and underlying bedrock surface.

- The aquifer is primarily comprised of coarse (sand and gravel sized) alluvium deposited by the old Faro Creek near the main access road but is also interpreted to include some Till of sandy or gravelly silt texture further up-valley (towards the Faro waste dumps) and towards the valley sides.
- The distribution/thickness of alluvial and till deposits up-valley from the ETA is uncertain. The alluvial unit is interpreted to continue under the access road, sloping downwards with topography.
- The tailings vary in texture from gravel and sand to silt with lenses of visible pyrite-rich sand/gravel observed in drill-core. The tailings deposit is approximately 6.5m thick near the access road and thins up-valley and towards the valley sides.
- Coarser materials within the tailings deposit may represent mine wastes (waste rock) being placed onto the tailings surfaces after intermittent discharges to the ETA. These coarser waste layers may represent preferential pathways in the tailings unit but cannot be accurately delineated with the available data.
- The tailings are interpreted to represent a partially confining layer to the alluvial aquifer (where tailings are fine sand/silt-size) which may desaturate during pumping in the alluvial aquifer.
- The narrow bedrock channel observed below the main access road is interpreted to extend up-valley under the mine access road into the ETA to approximately SRK04-4. The bedrock surface is interpreted to rise up-valley from the access road to the waste rock dumps and valley-sides.
- Fine grained (silt) Till is assumed to blanket the adjacent hillsides bordering the ETA

Significant uncertainty exists regarding the distribution and thickness of alluvial and till deposits up-valley where there is little geologic data. Preferential pathways in coarser till units or thicker alluvial sediments may exist but cannot be accurately defined with the available data.

Hydraulic properties for the aquifer were estimated by analyses of the 2005 24-hr pumping test using analytical techniques:

- Aquifer transmissivity (T) was estimated to be approximately 66 m<sup>2</sup>/d, with an average hydraulic conductivity (K) of 2 x 10<sup>-4</sup> m/s. These estimates are based on a new analysis of the 24hr pump test data.
- Aquifer storativity was estimated to be approximately 0.02; this estimate is significantly greater than typical values for confined aquifers (<0.001) suggesting the influence of leakage to the alluvial aquifer (semi-confined conditions).
- Drawdown curves for several wells screened in the aquifer and the overlying tailings suggest leakage from the tailings to the aquifer, indicating the potential for complete desaturation of the tailings
- Seepage from below the mine access road was not observed to change significantly during the 24 hour pumping test
- The bedrock monitoring well (SRK05-ETA-BR2) showed delayed but similar drawdown response to shallower wells screened within the aquifer (SRK05-ETA-BR1).

Table 1 summarizes selected groundwater quality data collected in 2005 from existing and newly installed monitoring wells in the ETA area. The following conclusions can be drawn about the groundwater quality in the ETA area:

• All groundwater samples exceed the CCME limit of 0.05 mg/l for Zinc

- Groundwater quality in the two up-gradient wells (P96-6A&B) showed significant variations over time suggesting variable contributions of (more dilute) surface water recharge;
- Concentrations of sulphate, zinc, and in particular iron were generally higher in alluvial wells located in the ETA area compared to those wells located up-gradient; furthermore, the highest concentrations of sulphate, zinc and iron were observed in the single well screened in tailings (SRK04-03B); these observations suggest that the ETA tailings represent a significant source of contaminant loading (in particular iron) to the Faro Creek seepage;
- The sample from the bedrock well, SRK05-ETA-BR2, had the lowest contaminant concentrations of all wells in the ETA. This information suggests that bedrock groundwater is less affected by ARD seepage from the Faro mine site than alluvial groundwater;
- Lower concentrations in the bedrock sample may indicate that deeper groundwaters are less contaminated than shallow groundwaters. Bedrock hydraulic conductivities estimated from packer tests in SRK05-BR2 were low (10<sup>-7</sup> 10<sup>-8</sup> m/s) and suggest that bedrock groundwater fluxes are also low. Consequently, bedrock groundwaters may not be critical for collection (i.e. bypass of deep groundwaters may have lower impact to receiving environment than shallower alluvial groundwaters).

ID	Date	Lab pH	Lab Conductivity (µS/cm)	SO4 (mg/L)	Zn (mg/L)	Fe (mg/L)
	N	lovembe	er 2005 Sampling	g		
SRK05-ETA-BR1	11/2005	5.42	9750	9250	681	3100
SRK05-ETA-BR2	11/2005	6.8	2040	1200	7.41	22.5
	October	2005 S	RK04-04 Pumpi	ng Test		
SRK04-4	10/2/2005	5.44	7630	7370	438	1950
	Octob	per 2005	Sampling (Pre-	Test)		
P96-8A	9/10/2005	6.50	6370	5040	604	0.061
P96-8B	9/10/2005	6.35	6620	4980	368	9.85
	÷	May 2	005 Sampling			
SRK04- 04-04	5/5/2005	5.23		7080	350	1630
SRK04-03A	5/5/2005	5.87		5480	233	693
SRK04-03B	5/5/2005	3.72		16700	749	6610
P96- 8A	5/3/2005	6.76	197	71.2	1.67	0.064
P96- 8B	5/3/2005	7.01	5540	4520	173	0.22

Table 1 Monitoring Well Water Quality

Surface water samples were taken repeatedly in 2005 at four stations along the Faro Creek channel (from X23 at the toe of the WRDs to the mouth of the Faro Creek Canyon). Table 2 summarizes the concentrations of selected constituents (SO4, Zn-T and Fe-T) observed in these surveys. The measured streamflows at each station are also shown for reference. The following conclusions can be drawn with respect to streamflows and surface water quality in the ETA area:

- Toe seepage from the WRDs (FCS-1 at X23) occurs year-round with some decrease in seepage during winter baseflow; concentrations of sulphate, zinc and iron are highly elevated but have remained relatively steady during the period of observation;
- Surface runoff from the ETA area (FCS-2 at culvert) shows more variable flow than toe seepage upgradient of the ETA area but concentrations of sulphate, zinc and iron are very similar to those observed in toe seepage up-gradient of the ETA area (except for a small decrease in total iron concentrations);

- Subsurface seepage from the ETA area (FCS-3 at downstream seepage face) flows year-round with flow estimates ranging from 2.3 4.8 L/s; water quality of this seepage is generally similar to that observed in the alluvial wells in the ETA area; iron concentrations in this seepage are consistently about one order of magnitude higher than in WRD seepage entering (and leaving) the ETA area;
- Surface runoff at the mouth of the Faro Creek Canyon (FCS-4) also flows year-round with flow rates ranging from 4.5 to 11.7 L/s (peak flows during snowmelt and/or heavy precipitation events may be higher but have not yet been measured); the water quality represents a mixture of surface runoff and subsurface seepage from the ETA area with intermediate concentrations of total iron;
- The two most reliable flow surveys (May and October 2005) suggest that incremental gains in streamflow along the Faro Creek Canyon (between FCS-2/3 and FCS-4) are very small (0.1 to 0.7 L/s) suggesting only a very small, if any, groundwater discharge along the Faro Creek canyon; this hypothesis is supported by loading calculations for October 2005 (see below).

		Flow	S04	Zn-T	Fe-T
Station ID	Date	L/s	mg/L	mg/L	mg/L
	10-Apr-05	1.3	5,030	295	88.9
(RD	13-May-05	4.6		no sample	
S-1 S_1	18-Oct-05	1.2	6,200	477	157
S E C	21-Nov-05	1.0	6,370	458	131
Att	19-Dec-05	0.6	5,920	516	180
	Average	1.7	5,880	437	139
pad	10-Apr-05	frozen		no sample	
N IC	13-May-05	9.0		no sample	
S-2 belo	18-Oct-05	3.4	6,210	459	61.6
ert b	21-Nov-05	>1.0	5,890	437	36.8
NIN:	19-Dec-05	frozen		no sample	
At 6	Average	n/a	6,050	448	49
MO	10-Apr-05	4.8	5,550	309	1,210
bel	13-May-05	2.6		no sample	
S-3 ace ad	18-Oct-05	3.1	6,570	222	1,120
ge f ro	21-Nov-05	3.4	7,460	371	1,790
epa	19-Dec-05	2.3	7,030	430	1,990
Se	Average	3.2	6,653	333	1,528
ч	10-Apr-05	6.6	4,170	174	801
oyn	13-May-05	11.7	No sample		
S-4 Ca	18-Oct-05	7.2	5,750	310	773
h of FC	21-Nov-05	5.4	5,610	266	940
lout	19-Dec-05	4.5	5,540	278	1,220
2	Average	7.1	5,268	257	934

Table 3	2: 5	Surface	Water	Quality	,
	L. \	Juillace	<b>T</b> ator	Quanty	,

Table 3 summarises the observed surface water flows and calculated loads of sulphate, total zinc and total iron at the four sampling locations along the Faro Creek channel. The following conclusions can be drawn with respect to contaminant loading in surface water upstream and downstream of the ETA area:

- The contaminant load associated with surface runoff entering the ETA area is generally much smaller (<25%) than the combined contaminant load discharging from the ETA area (in surface runoff and seepage flow combined);
- The contaminant load associated with surface runoff from the ETA area varies significantly (in relative and absolute terms), mainly due to the large variability in flow;
- Subsurface seepage discharging downstream of the ETA area represents the primary source of contaminant loading during winter baseflow and a significant source of loading during the remainder of the year;
- The total sulphate and zinc load discharging from the ETA area (at FCS-4) varies with flow conditions; for example, zinc loading ranged from 70 t/yr during the (wet) fall to 36 t/yr during the winter baseflow; the total iron load at FCS-4 remained surprisingly constant over time (~170 t/yr);
- During the October 2005 survey, the combined sulphate and zinc loads from FCS-2 and FCS-3 agreed very well with the observed total loads at FCS-4. The load estimates for the other surveys are considered too uncertain (in particular for FCS-2 due to freezing of the culvert) to allow similar mixing calculations.

Date	Station	Flow L/s	S04 Load t/yr	Zn Load t/yr	Fe Load t/yr
	FCS-1	1.3	206	12.1	3.6
05	FCS-2	frozen	0	0	0
Apr-	FCS-3	4.8	840	46.8	183
10-/	FCS-4	6.6	868	36.2	167
	FCS-1	1.15	225	17.3	5.7
05	FCS-2	3.4	666	49.2	6.6
Oct-	FCS-3	3.1	642	21.7	109
18-0	FCS-4	7.2	1306	70.4	176
	FCS-1	1.01	203	14.6	4.2
-05	FCS-2	1.0 to 2.0	196 – 392	14 – 29	1.9 – 3.9
Vov	FCS-3	3.4	800	39.8	192
21-1	FCS-4	5.4	955	45.3	160
	FCS-1	0.63	118	10.3	3.6
-05	FCS-2	frozen	0	0	0
Dec	FCS-3	2.3	510	31.2	144
19-1	FCS-4	4.5	786	39.5	173
			Estimated rand	pe	

Table3:	<b>ETA Surface</b>	Water and	Seepage	-low and Zin	c Loading,	October	2004
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A summary of conceptual groundwater flow in the ETA is as follows:

- Under static conditions groundwater generally flows from the north-east to the south-west with converging flow towards the seepage face below the access road.
- The converging flow potentially indicates the presence of a bedrock paleochannel extending beneath the road into the ETA. The bedrock channel is assumed to have been filled with coarse waste rock during the construction of the access road.
- Groundwater is discharging into the current Faro Creek alignment along much of its length in the ETA, resulting in the increase of flow between stations (Table 3).
- The observed groundwater discharge from the ETA area, as measured by FCS-3 at the seepage face below the access road, ranges from 2.3 to 4.8 l/s. The October survey suggests that this station captures most, if not all, groundwater moving through the alluvial aquifer in the ETA area

- Groundwater seepage from ETA is year-round as indicated by measurements made as recently as mid-December 2005 by Laberge Environment, with flow decreasing during winter;
- After pumping SRK04-4 for 24-hrs, groundwater flows is generally towards the pumping well, but has not yet captured flow discharging at the seepage face.
- Uncertainty exists whether the aquifer will act as a confined aquifer with significant leakage from the tailings, or as an unconfined aquifer with the tailings draining completely during extended pumping.

# CONCEPTUAL GROUNDWATER COLLECTION SYSTEM

A conceptual groundwater collection system was presented in *Preliminary Seepage Collection Options*, *SRK*, *2004*. This design was used to provide the basis for scoping simulations of ETA groundwater capture. A brief summary of the concept used for simulations is as follows:

- Three pumping wells installed adjacent to the main access road at the south-western edge of the ETA
- Each well to have variable speed pumps, with a potential capacity between 0.93 and 2.84 l/s (15-45 USGPM or 55-275 m<sup>3</sup>/d).
- Well are to be screened within aquifer sediments only and potentially 1-2 meters below the top of the aquifer. The water levels within the wells are not to be drawn to the screens to limit oxidation on the screens.
- For simulations purposes, the top of the well screens has been set to 1097 masl, above the estimated elevation of the seepage face at approximately 1096.5 masl. This configuration assumes that the aquifer and bedrock surface are dipping to the south-west.

# SCOPING SIMULATIONS

Two numerical groundwater models were constructed for the ETA to conduct scoping level simulations for (i) confined conditions and (ii) unconfined conditions. The purpose of these scoping simulations was to test conceptual model assumptions and provide further information on the issues of:

- potential hydraulic barriers to collecting groundwater,
- lack of seepage flow cut-off during 24hr pumping test, and
- scoping level estimates of capture efficiency.

# **Model Construction**

The model extents, mesh, and observation wells were the same for both the confined and unconfined models and are shown in Figures 1 and 5. The models were 2-D simulations due to the limited information available regarding surficial geology and bedrock topography away from the ETA wells.

The choice of 2-D simulations allows for a simple geologic conceptual model commensurate with geologic uncertainty but has drawbacks for calculating transient model heads from pumping. The 2-D confined model simulations assume the aquifer sediments remain saturated during pumping and do not receive significant leakage from the overlying tailings. The 2-D unconfined simulations allow for desaturation of the aquifer but do not allow for additional leakage from the overlying tailings.

It is recognized that neither model may represent actual aquifer conditions (i.e. a semi-confined aquifer with leakage from the overlying tailings). Nevertheless, these model simulations should be sufficient to provide a scoping level test of assumptions and insight on identified issues. It was assumed that the access road was constructed of coarse waste rock placed and thus not a barrier to flow in the model.

# **Model Calibration**

The models were first run steady-state and adjusted to static conditions (pre-pumping) observed on October 1<sup>st</sup>, 2005. Next, the 24 hr pump test at SRK04-4 was simulated using the initial model and the resulting calculated head/drawdown compared to observed end-of-pumping conditions. The simulated water levels

were compared to observed levels in ETA monitoring wells screened in aquifer sediments (SRK04-4, ETA05-2, ETA05-04, ETA05-3, SRK04-3A). ETA monitoring wells screened in tailings (ETA05-1, ETA05-5, SRK04-3B) or bedrock (SRK05-ETA-BR2) were not used for comparison. The distant monitoring well below the waste rock dump, P968A/B, was also not used.

Calibrated confined model transmissivities and recharge are shown in Tables 5 and 6.

Parameter	Description	Transmissivity
$T_1$	Fine grained till blanket on hillsides	$1.15 \text{ x } 10^{-6} \text{ m}^2/\text{s}$
$T_2$	Till, sand, and gravel sediments	$4.63 \text{ x } 10^{-4} \text{ m}^2/\text{s}$
T <sub>3</sub>	Sand and gravel sediments in ETA	$6.94 \text{ x } 10^{-4} \text{ m}^2/\text{s}$
$T_4$	Bedrock paleochannel filled with coarse waste rock	$185 \text{ x } 10^{-4} \text{ m}^2/\text{s}$

### Table 4 Confined Model Transmissivities

# Table 5 Confined Model Recharge Distribution

Parameter	Description	Recharge
<b>R</b> <sub>1</sub>	Hillsides, low infiltration	30 mm
<b>R</b> <sub>2</sub>	Valley sediments	450 mm
<b>R</b> <sub>3</sub>	Valley sediments in ETA, including potential	900 mm
	leakage from pipeline	

#### Table 6 Unconfined Model Transmissivities

Parameter	Description	Transmissivity
<b>K</b> <sub>1</sub>	Fine grained till blanket on hillsides	1.0 x 10 <sup>-5</sup> m/s
K <sub>2</sub>	Till, sand, and gravel sediments	$1.0 \ge 10^{-4} \text{ m}^2/\text{s}$
<b>K</b> <sub>3</sub>	Sand and gravel sediments in ETA	$2.5 \text{ x } 10^{-4} \text{ m}^2/\text{s}$
$K_4$	Bedrock paleochannel filled with coarse waste rock	$50 \text{ x } 10^{-4} \text{ m}^2/\text{s}$

#### Table 7 Unconfined Model Recharge Distribution

Parameter	Description	Recharge
<b>R</b> <sub>1</sub>	Hillsides, low infiltration	30 mm
R <sub>2</sub>	Valley and ETA sediments	450 mm

Model storage storativity was as follows:

- 0.002 for all areas for confined simulations
- 0.05 for ETA and 0.02 for surrounding areas for unconfined simulations

Model boundary conditions were as follows:

- Constant heads were set at ~2m below topography along most of model boundary
  - Heads were constrained by flux to only allow outflow from the south-west boundary below the access road, heads were set at the estimated elevation of the seepage face of 1096.5 masl.
- Transfer boundary along Faro Creek with water levels set approximately at current elevation and maximum inflow/outflow rate 1/s per unit length of creek.
- The calibrated stead-state model was run transiently for 24 hours while pumping SRK04-4 at 244  $m^3/d$  (45 USGPM).

# **Model Predictions**

Once calibrated, the numerical models were used to simulate various pumping scenarios over a year of pumping. The number of wells, pumping rates, and locations were varied to increase capture efficiency (reduce ETA groundwater bypassing wells).

Capture efficiency was generally defined as the reduction in groundwater flow through the ETA from static conditions to pumping conditions. This flow was measured in simulations as the net flux past the boundary line indicated in Figure 5. This net flux reduction was then compared for the various pumping scenarios.

#### **Results of Scoping-Level Simulations**

#### Steady-State

Both the confined and unconfined models reasonably simulated steady-state conditions in the ETA (Figure 6).

Results of confined steady-state simulations were as follows:

- Calculated heads were within approximately 0.4 m of observed heads with the exception of the upvalley well, ETA05-4. The root-mean squared (RMS) error was 0.25 m, excluding ETA05-4 and 0.89 m with the 05-4 well.
- Calculated flux from the model was  $\sim 3.5 \text{ l/s}$  ( $\sim 304 \text{ m}^3/\text{d}$ ), with most of that coming from flux through the ETA (3 l/s). The total flux is similar to observed seepage at FCS3 (Table 3).
- The observed pattern of converging heads could not be simulated without the presence of a high Transmissivity paleochannel extending into the ETA.

ETA Monitoring Well	Calculated Head (masl)	Observed Head (masl)	Residual Head (masl)
SRK04-4	1098.15	1098.18	-0.02
ETA05-2	1098.95	1098.77	0.18
SRK05-ETA-BR1	1098.03	1098.31	-0.28
ETA05-3	1098.32	1098.55	-0.23
SRK04-3A	1098.06	1098.45	-0.38
ETA05-4	1100.87	1102.97	-2.10

Table 8 Static Conditions: Calculated Heads and Residuals

Results of unconfined steady-state simulations were as follows:

- Calculated heads were within approximately 0.5 m of observed heads. The RMS error was 0.36 m for all wells.
- Calculated flux from the model was 6 l/s (518 m<sup>3</sup>/d), with most of that coming from flux through the ETA (5.65 l/s). The total flux was significantly greater than the observed seepage at FCS3 during moderate to low flow conditions (typical for the October 1 survey) but similar to previous broad estimates of groundwater flux (*Faro Seepage Collection Options Report, SRK, 2004* and *Discussion Memo: Preliminary ETA Results, SRK, 2005*).
- The observed pattern of converging heads could not be simulated without the presence of a high hydraulic conductivity paleochannel extending into the ETA.

Transient

The confined model reasonably simulated overall drawdown after 24-hrs of pumping SRK04-4, and was considered suitable for scoping level analyses of groundwater capture (Figure 7). The confined model over-predicted the drawdown in the pumping well and under-predicted the observed drawdown in the observations

wells. Leakage from the overlying tailings (not simulated) is the likely cause for this inconsistency between predicted and observed drawdown with distance from the pumping well.

The unconfined model was less successful in simulating drawdown, resulting in higher RMS (0.7) due to insufficient drawdown at most observation wells (Figure 7).

Results of confined transient simulations were as follows:

- The overall RMS of calculated vs. observed drawdown was 0.41 m, sufficient for scoping simulations of capture.
- The model overpredicted the actual pumping level by 0.7 m and underpredicted the drawdown in the surrounding aquifer levels (between 0.03 0.3 m)
- The calculated flux leaving the model (at the downstream seepage face) was reduced to 1.9 l/s but was not cut-off. A stagnation point occurred between the pumping well and the seepage face limiting the capture of seepage.
- Simulated storativity (0.002) was less than determined from pumping tests (0.02) as the model could not simulate vertical leakage.

ETA Monitoring Well	Calculated Drawdown (m)	Observed Drawdown (m)	Residual Drawdown (m)
SRK04-4	3.30	2.572	0.73
ETA05-2	1.28	1.339	-0.06
SRK05-ETA-BR1	1.31	1.618	-0.31
ETA05-3	0.82	0.731	0.09
SRK04-3A	1.14	1.173	-0.03
ETA05-4	0.80	0.178	0.62

 Table 8 24-Hr Pumping Test Heads Comparison to Confined Simulations

# Capture Simulations

The confined model was used to simulate capture from several extraction wells installed above the access road. Results of the "steady-state" (year-long) capture simulations were as follows (Figures 8-10):

- The use of 3 pumping well (at a pumping rate of 2.84 l/s each) with the top of screen set 0.5m above the seepage face elevation (1097 vs. 1096.5 masl) captured only about ~1/3 of groundwater flow through the ETA.
- A stagnation point occurred between the pumping wells and the seepage face due to the limited drawdown.
- Capture efficiency did not increase with increasing pumping rate as pumping durations at low rates were still limited by screen elevation (wells turned off in model when maximum allowable drawdown reached).
- Adding incrementally more extraction wells with similar screen elevations only moderately improved capture efficiency (2/3 of flow, Figure 9). Simulations of 7 and 9 wells (2.84 l/s each) still had significant bypass (~1/3 of flow). The capture efficiency was primarily controlled by the limited allowable drawdown in the pumping wells and the position of the pumping level and seepage face elevation (Figure 10).
- Effective capture (>90%) occurred only when pumping levels were able to drop to the same or below the elevation of the downstream seepage face. Due to the sloping bedrock surface and aquifer it is expected to be difficult to drop the aquifer water levels to/below the seepage face elevation (Figure 11).
- The assumed presence of a paleochannel extending beneath the access road focussed bypassing groundwater to the seepage face by acting as a drain. Thus the existence of a paleochannel would

excluded from calculation of RMS

reduce capture efficiency of any upstream collection system, but could improve efficiency of downstream collection below the access road.

# DISCUSSION

Scoping simulations have been carried out to test conceptual model assumptions and provide further information on:

- potential hydraulic barriers to collecting groundwater,
- lack of seepage flow cut-off during 24hr pumping test, and
- scoping level estimates of capture efficiency.

The confined model reasonably simulated observed conditions in the ETA and was used to simulate capture of groundwater above the access road. The following conclusions are drawn from the simulations:

- The limited available drawdown (<2.5m) for any potential ETA groundwater pumping system above the access road may result in poor capture (i.e. bypass in the order of 10% or greater). The pumping wells will likely be unable to drawdown aquifer levels to/below the seepage face on a continuous basis and thus may not cut off groundwater discharge to Faro Creek below the access road.
- There is a significant potential for continued "on-off" cycling of pumps (due to the limited available drawdown) which further increases the potential for groundwater bypass to Faro Creek.
- Attempts to draw groundwater levels further into the aquifer may expose a portion of the well screens to the atmosphere, likely resulting in iron oxidation ("well fouling") of the exposed screens, these conditions can be expected to cause difficulties with well maintenance (well efficiency, pump failure, etc.).
- Simulations also suggest that the presence of a paleochannel (or similar high permeability zone) beneath the access road may provide a means to concentrate seepage to a small discharge zone (seepage face). Such a paleochannel would further complicate upstream groundwater collection and facilitate groundwater bypass of the wells. Contrastingly, this focussed seepage would improve capture efficiency for a seepage collection system installed below the access road.
- Simulations suggest that groundwater discharge at the seepage face should have decreased but not cut-off completely during the 24-hr pumping. This simulated reduction in flux is likely because the model does not allow for leakage from tailings, which would supply additional water to the pumping well. It is interpreted that a stagnation point developed between the pumping well and the seepage face.

In summary, scoping simulations indicate that the position of the seepage face below expected pumping levels due to limited available drawdown represents a significant impediment to an effective groundwater capture by active pumping. The balance of flows and load estimates indicates that most seepage is accounted for in FCS2 and FSC3. Based on the available information, it therefore appears that seepage and surface collection immediately below the road in heated sumps is likely more effective than a groundwater pumping system.

# RECOMMENDATIONS

The following preliminary recommendations are made regarding the capture of contaminated groundwater from the ETA:

- A combined seepage and surface water collection system of sumps should be designed below the main access road near FCS2/3 (Figure 12). This system would likely capture the majority of groundwater and base flows from the ETA and allow for easier maintenance of installations.
- Further detailed flow and load surveys should be conducted along Faro Creek to determine where groundwater is discharging to/receiving from Faro Creek. This data will assist determination of potential variability (and inconsistency) of flows/loads.

- Further modelling to simulate groundwater conditions in 3-dimensions is not recommended due to limited geologic data up-valley and similar difficulty of seepage face elevation control.
- A simple test of the sump concept could be conducted by collecting all flows at FCS2/3 and pumping/bypassing them downstream of FCS4. Flows would then be monitored at FCS4 to estimate potential bypass.
- Additional drainage ditches and sumps could be installed within the ETA (e.g. base of WRD) to control surface drainage and thereby reduce downstream seepage collection and prevent contaminated flows from entering Faro Creek.









SECTION A-A' 1:750





	00.4			+.38		6.87	
	1104			1104		1105	
				0+24	-0		
<u>LE</u>	GEND						
	ROAD FI	ILL					
	COARSE SAND &	TAILI GRA	NGS VEL ALL	.UVIUM			
	TILL, CC SAND & PHYLLITI	MPAC GRA	T COAR VEL TO	SE SILT			
	GROUND	WATE	R FLOW	ARRO	V		
			FAR	D MIN	e site	ΕΤΑ	
				Sectio	on A-A	λ'	
	PROJECT NO.	073	DATE Feb.	2006	APPROVED	IK	FIGURE 3



SECTION B-B'

SCALE: 1:1000



1107.01		 					
1107.25							
1107.67							
1108.28							



+	200

0+240

<u>LEGEND</u>

$\sim$
$\sim$
$\geq$

ROAD FILL COARSE TAILINGS SAND & GRAVEL ALLUVIUM TILL, COMPACT COARSE SAND & GRAVEL TO SILT PHYLLITIC BEDROCK

	FARO MIN	E SITE ETA	
	Sectio	n B-B'	
PROJECT NO.	DATE	APPROVED	FIGURE
1CD003.073	Feb., 2006	QJK	4





Confined Simulation: RMS = 0.89m RMS=0.25 without ETA-05-4 Unconfined Simulation: RMS = 0.36m









APPROVED:

DATE:

Jan 2006

PROJECT: 1CD003.73

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In this scenario all well screens are set at 1097 masl, 0.5 m above seepage face.

N

While capture zones appear to cover most of the ETA, significant bypass occurs when wells are forced to cycle on/off with fluctuating water levels.







