

**Deloitte  
& Touche**

**Anvil Range Mining Complex  
2007 Waste Rock and Seepage  
Monitoring Report**

**2007/08 Task 17**

**DRAFT**



*Prepared for:*

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*On behalf of*  
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*Project Reference Number:*  
SRK 1CD003.090

**March 2008**

# **Anvil Range Mining Complex 2007 Waste Rock and Seepage Monitoring Report**

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Report prepared for  
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on behalf of  
**Faro Project Management Team**

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## Executive Summary

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### Objectives and Primary Findings:

This report summarizes the waste rock and seepage monitoring results from the 2007 monitoring program and compares the current year's results with results from previous monitoring.

Water quality at toe seepage and select routine monitoring stations indicates that water quality in drainage from the Faro and Vangorda waste rock dumps appears to have stabilized year-over-year, with considerable seasonal variation at several monitoring stations.

Water quality of drainage from the Grum Dump toe seeps also appears to be stable, with dissolved zinc concentrations typically in the range of 1.7 to 5.2 mg/L over the past several years, and isolated samples having dissolved zinc concentrations up to 60 mg/L. However, downgradient monitoring at routine monitoring station V15 demonstrated a sustained increase in zinc concentrations over the 2007 monitoring period to nearly 1.2 mg/L, and it is expected that this increase will continue in 2008.

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### Future Work Recommendations:

It is recommended that waste rock and seepage monitoring be continued at the same level as carried out since 2005.

## Table of Contents

Executive Summary .....	i
<b>1 Introduction .....</b>	<b>1</b>
<b>2 Methods .....</b>	<b>1</b>
2.1 Waste Rock Seepage Surveys .....	1
2.2 Pit Seepage Surveys.....	2
2.3 Routine Monitoring .....	3
2.4 Temperature and Oxygen Monitoring of Waste Rock .....	3
<b>3 Results.....</b>	<b>4</b>
3.1 Waste Rock Seepage Surveys .....	4
3.2 Pit Seepage Surveys.....	4
3.3 Routine Monitoring .....	4
3.4 Temperature and Oxygen Monitoring of Waste Rock .....	4
<b>4 Discussion.....</b>	<b>5</b>
4.1 2002 to 2007 Waste Rock Seepage Surveys .....	5
4.1.1 Faro Waste Rock Dumps .....	5
4.1.2 Grum Waste Rock Dumps.....	9
4.1.3 Vangorda Waste Rock Dumps.....	12
4.2 Routine Monitoring .....	13
4.2.1 Faro Site .....	13
4.2.2 Grum Site .....	16
4.2.3 Vangorda Site.....	18
4.3 Temperature and Oxygen Monitoring of Waste Rock .....	19
<b>5 Conclusions .....</b>	<b>20</b>
<b>6 References.....</b>	<b>22</b>

## List of Tables

Table 1: Characteristics of Faro Water Types.....	6
Table 2: Seepage Stations Classified By Water Type.....	8
Table 3: Characteristics of Grum Water Types .....	11
Table 4: Characteristics of Vangorda Water Types.....	12

## List of Figures

Figure 1: Waste Rock Seep Survey Locations - Faro Waste Rock Dumps	
Figure 2: Waste Rock Seep Survey Locations - Vangorda and Grum Waste Rock Dumps	
Figure 3: Pit Seep Survey Locations - Faro Pit	
Figure 4: Pit Seep Survey Locations - Grum Pit	
Figure 5: Pit Seep Survey Locations- Vangorda Pit	
Figure 6: Faro Seepage Station X23: Select Monitoring Results	
Figure 7: Faro Seepage Station X26: Select Monitoring Results	
Figure 8: Faro Seepage Stations FCO and A30: Select Monitoring Results	
Figure 9: Faro Seepage Station SP5-6: Select Monitoring Results	
Figure 10: Faro Seepage Stations NE1, NE2, and W5: Select Monitoring Results	
Figure 11: Faro Seepage Stations W8 and W10: Select Monitoring Results	
Figure 12: Grum Seepage Stations V15 and V2: Select Monitoring Results	
Figure 13: Grum Seepage Station V2A: Select Monitoring Results	
Figure 14: Vangorda Seepage Station V30: Select Monitoring Results	
Figure 15: Vangorda Seepage Station V32: Select Monitoring Results	
Figure 16: Vangorda Seepage Station V33: Select Monitoring Results	
Figure 17: Vangorda Seepage Station V27: Select Monitoring Results	

## List of Appendices

Appendix A: Waste Rock Seepage Monitoring Results	
Appendix B: Pit Seepage Monitoring Results	
Appendix C: Thermal and Pore Gas Monitoring Results	

# 1 Introduction

Waste rock and pit seepage surveys, and monitoring of waste rock thermal conditions and pore gas oxygen concentrations were continued in 2007. In addition, routine monitoring of selected waste rock and pit seepage stations was carried out by site environmental staff as a requirement of the water licence. This report presents results of the 2007 seep surveys, a review of the 2007 and historical routine seepage monitoring data, and results of the thermal and oxygen monitoring.

This report has been prepared as part of the ongoing technical evaluation for the closure planning of the Anvil Range Mining Complex.

# 2 Methods

## 2.1 Waste Rock Seepage Surveys

Sample locations were established in June 2002 by walking the toes of all waste rock dumps where the rock rests on original ground, and looking for any flowing seeps that emerged from these areas. Additional seeps were located by slowly driving along accessible roads and ramps in the Faro Pit complex that were below waste rock dumps or ore stockpiles. Most of the seeps were flowing, or were inferred to have been recently flowing based on observations of moisture along flow paths, or on ponded water surfaces at the elevation of pond spill points.

Subsequent seep surveys at each of the waste rock dumps were carried out in September 2002, and in both spring and fall of 2003 and 2004. Stations established in June 2002 were revisited and sampled as appropriate. Some seeps were found to flow intermittently, and new intermittent seeps were located during the follow-up surveys. At the end of 2004, there were approximately 80 seep locations that had been sampled at least once during the 2002-2004 waste rock seepage monitoring program (SRK 2006a).

As part of planning for 2005 monitoring, the waste rock seepage working group discussed the value of continuing to monitor waste rock seepage at the established stations. On the basis of these discussions, SRK carried out a review of previously-sampled seepage stations to assess which stations should continue to be monitored on a twice-yearly basis. Summaries of the rationale for maintaining or eliminating each station from the monitoring program were compiled in a draft memorandum that was circulated and subsequently revised to reflect feedback from the working group (see Appendix A in SRK 2006b). Sites identified for continued monitoring were visited in May and September of 2005 to 2007, and samples were collected as flow conditions permitted.

Several sites where waste rock seepage stations coincided with water licence sampling stations were identified for elimination pending review of sampling practices employed by site staff. These sites were reviewed at the beginning of May during the 2005, 2006 and 2007 field visits, but it was found

that field filtration of dissolved metal samples was not being carried out as part of water licence sampling. As the protocol for the waste rock seepage survey includes field filtration of dissolved metal samples, it was decided that the water licence stations would continue to be sampled on an interim basis as part of the waste rock seepage monitoring program.

The review of monitoring stations also recommended allowance for a limited number of supplementary samples, to be collected at the investigator's discretion, to characterize intermittent surface flows that daylight only under ideal flow conditions.

In past years, laboratory analysis of seepage samples has been carried out by ICP-OES. During planning meetings for 2007 monitoring, a need was identified to obtain lower detection limits for several trace parameters. It was decided by the various technical working groups that the added cost was justified, and all 2007 samples were analyzed using a combination of ICP-OES and ICP-MS.

Samples were collected for analyses of routine parameters (pH, conductivity, acidity, alkalinity, chloride and sulphate), and dissolved metals (dissolved metals by ICP-MS/OES). The samples were filtered and preserved in the field according to standard methods for collection of environmental samples. Field pH, conductivity, redox, and temperature measurements were taken at each station using a WTW meter. Flow estimates were made using the bucket and stopwatch method, by estimating the velocity and cross sectional area of the seep, or by visual estimation.

The sampling locations were marked for later reference with flagging tape and were surveyed using a hand-held GPS. The locations are shown in Figures 1 and 2. Photographs were taken to document the general appearance of the station, as well as any precipitates along the flow paths.

Duplicates and field blanks were collected as a check on the quality of the field methods and laboratory results.

## 2.2 Pit Seepage Surveys

Monitoring of seepage and wall rock runoff in the Faro, Grum, and Vangorda pits was carried out in 2003 and 2004 as part of studies relating to future pit lake water quality and treatment requirements. Beginning in 2005, seepage monitoring has also been carried out in the pits, as recommended by the technical working group. It was felt that permanent monitoring stations should be established to allow pit seepage water quality trends to be tracked over time.

Pit seepage surveys were carried out in Faro and Grum pits in May 2007 and in all three pits (Faro, Grum and Vangorda) in September 2007. Pit seep locations are shown in Figures 3 through 5.

Pit seep samples were collected and analysed in the manner described for waste rock seeps in Section 2.1.

## 2.3 Routine Monitoring

Waste rock seep surveys have been completed both by the operator, as prescribed in water licences, and by various regulatory authorities since approximately 1986 at the routine sample locations shown in Figures 1 and 2. Deloitte & Touche has continued to collect seep samples on a regular basis as part of water licence routine monitoring programs.

The historical water quality database was transferred to a Microsoft Access platform, to allow distribution of results to other users. The development of the Access database was coordinated by Gartner Lee Limited (GLL), who report the compiled results in the annual monitoring reports (e.g. GLL 2007) and who continue to maintain the database. Quality assurance and validation of reported historical monitoring results is undertaken as issues are raised.

## 2.4 Temperature and Oxygen Monitoring of Waste Rock

Waste rock temperature and oxygen monitoring has been carried out intermittently beginning in February 2003 at seven instrumented drillholes on the Faro, Grum, and Vangorda waste rock dumps (SRK, 2004a). In each drillhole, a thermistor string was installed that measures temperature at discrete locations down hole. In addition, each hole has several plastic tubes installed that permit pore gas to be extracted from discrete locations. The installation of these monitoring facilities is described in detail in SRK 2004a.

In the past, erratic results from some individual monitoring ports have raised the question of whether the ports were blocked or pinched. Prior to sampling, a hand-operated vacuum pump was used to test whether pore gas was able to move freely through each monitoring tube. Those tubes which allowed a vacuum to develop were inferred as being blocked in some fashion, and the previous results from the blocked ports are considered questionable.

Pore gas oxygen content was measured directly using a Servomex Oxygen Analyser. This instrument was calibrated on a daily basis using ambient air (20.9% oxygen) and pure nitrogen gas (0% oxygen). The instrument was connected to each monitoring tube by means of silicone tubing, and pore gas was drawn through the instrument by means of an integrated pump. The oxygen analyser produced a direct real-time readout of oxygen content which was then manually recorded.

Temperature measurements were collected by connecting an Omega thermistor reader to a switch box at the terminal end of each installed RST Instruments thermistor string. Temperature at individual monitoring points was collected by selecting the appropriate channel on the switch box. The Omega thermistor reader measured resistance across each monitoring point and converted this measured resistance to temperature. The calculated temperature was then displayed on a digital readout, and was manually recorded.

In the 2007/08 field program, monitoring of pore gas oxygen content and waste rock temperature was carried out in May and September 2007, and in March 2008.

## 3 Results

### 3.1 Waste Rock Seepage Surveys

The results of the 2002 through 2007 waste rock seepage surveys are provided in Appendix A.

Select parameters (ranges of pH, conductivity, flow, sulphate and zinc concentrations for the period of record) are presented in Figures 1 and 2.

### 3.2 Pit Seepage Surveys

Locations of the pit seepage monitoring stations are shown in Figures 3 to 5. Results of the 2003 to 2007 pit seepage surveys are included in Appendix B. The purpose of these surveys has been to provide inputs for updates to water quality predictions (such as those presented in SRK 2004b) and to monitor changes in pit seepage water quality over time. Pit seep locations are less consistent than waste rock seep locations, necessitating the establishment of several unique stations to sample available flows. However, there are some stations that have been sampled over several monitoring rounds at locations where seepage has been consistently available. In some areas, for comparison of year-over-year seepage chemistry results, it is necessary to group stations where seepage water interacts with similar pit wall rock types.

### 3.3 Routine Monitoring

Locations of the routine seepage monitoring stations are shown in Figures 1 and 2. Analytical data are stored on the Anvil Range water quality database and are not reproduced here. The results of the historical seepage monitoring are provided in Figures 6 through 17. The time series shown in these figures provide a graphical summary of the data that allows trends to be easily identified.

### 3.4 Temperature and Oxygen Monitoring of Waste Rock

The results of temperature and oxygen monitoring are provided in Appendix C. Waste rock temperatures recorded in 2007/8 were generally within the range of those recorded previously.

Measured 2007/8 pore gas oxygen concentrations were also similar to those previously measured.

Vacuum testing of all individual sampling tubes indicated that several monitoring ports were blocked. Many of the blocked ports had a paired port at the same elevation that remained functional at the time of testing; however, three monitoring elevations (1.4, 5.6 and 10 m below surface) at installation 10M3 (Grum Dump) had pairs of monitoring tubes blocked at each elevation.

Installation 30M4 (Vangorda Dump) was found to have blockages in both monitoring ports at 1.4 m depth, however ports sampling both higher and lower elevations appeared to be functioning properly.

Installation 30M1 (Faro Dump) was found to have a blockage in the 30 m monitoring port. This port was not installed in duplicate and may not provide any further data. However, ports sampling both higher and lower elevations appeared to be functioning properly.

## 4 Discussion

### 4.1 2002 to 2007 Waste Rock Seepage Surveys

The 2007 freshet occurred in late May, which is typical timing based on monitoring since 2002. The spring waste rock survey took place during the same week as it did in 2006. Judging from the small remnant snow drifts on the rock dumps, and the lack of other snow cover, the waste rock runoff likely peaked at the end of the week prior to the waste rock seepage survey.

Fall runoff conditions appeared to be average to slightly dry, based on the number of flowing seeps and the size of previously observed flows. There was a significant rainfall event during the first day of the fall sampling period. This allowed a number of stations that flow only shortly after precipitation events to be sampled.

#### 4.1.1 Faro Waste Rock Dumps

No new seeps were identified at Faro during the 2007 surveys. The previously-developed grouping of seeps into water types continues to be a useful tool for summarizing the monitoring results.

##### Water Types

Seepage from the Faro Waste Rock Dumps was divided into three distinct types on the basis of pH and zinc concentrations (Table 1):

- Type 1 seeps typically had pHs greater than 6.5, and zinc concentrations less than 6 mg/L. Other trace metals (e.g. aluminum, iron, manganese) were low or below detection limits.
- Type 2 seeps had pHs typically between 6 and 7, and zinc concentrations from 4 to 595 mg/L. Cadmium, cobalt, iron, manganese, and nickel were also elevated in several of the samples.
- Type 3 seeps had pHs of less than 6, and zinc concentrations typically greater than 10 mg/L and as high as 21,100 mg/L. Aluminum, cadmium, cobalt, copper, iron, manganese, and nickel concentrations were also high in several of these samples. Samples with zinc concentrations greater than 900 mg/L are sourced from oxide fines, ore stockpiles, or the immediate area around the mill.

The summary of Faro water type characteristics in Table 1 was prepared from a modified data set. Where values were reported as less than detection, the detection limit was inserted as the analytical value for the purposes of the statistical calculations. Method detection limits are listed in Table 1; these limits were taken from the non-detect results of blank submissions. Any non-detect result that specified a detection limit more than 10X the minimum method detection limit was excluded from the statistical calculations. This arose in cases where samples had high ionic strength. The variation in the number of samples used in the statistical summary is a reflection of this exclusion.

**Table 1: Characteristics of Faro Water Types**

Parameter	Detection Limit	Type 1					Type 2					Type 3				
		Average	Median	Min	Max	N	Average	Median	Min	Max	N	Average	Median	Min	Max	N
Field pH (s.u.)	0.01	7.11	7.10	6.02	8.14	63	6.62	6.62	5.72	7.51	90	3.89	3.50	0.9	7.59	61
Acidity pH 8.3	1	10	8	1	30.7	63	210	74	14	2160	90	5478	352	27	49500	62
Alkalinity Total	1	139	147	19	322	63	164	108	2	445	90	14	2	0.5	310	62
Chloride	0.2	1.0	0.8	0.28	2.7	54	5.5	2.0	0.5	25	61	35.4	0.7	0.46	1050	41
Sulphate	0.5	445	382	3.1	2470	63	2607	2280	334	8180	90	7353	1935	69	59800	62
Calcium	0.05	97	95	7.5	263	63	355	388	49	675	90	218	214	6.5	508	62
Magnesium	0.1	77	69	1.09	378	63	375	269	37	1750	90	326	207	3.8	3210	62
Potassium	0.02	6.0	4.0	1.1	24	43	9.7	9.1	2.8	28.8	86	7.6	6.8	1.17	19.8	34
Sodium	0.01	21	7.0	1.5	150	47	26	19	3	122	89	14	5.4	1.62	114	46
Aluminum	0.001	0.014	0.0048	0.0020	0.070	8	0.53	0.24	0.0020	3.6	23	79	9.2	0.23	986	57
Cadmium	0.00001	0.0026	0.0010	0.00030	0.010	9	0.070	0.041	0.0073	0.62	77	4.2	0.16	0.018	47	60
Cobalt	0.0001	0.0045	0.0022	0.0001	0.016	11	0.22	0.065	0.010	1.1	86	2.0	0.28	0.030	20	61
Copper	0.0002	0.0077	0.0063	0.0016	0.021	11	0.12	0.038	0.0069	0.78	67	44	1.5	0.030	559	61
Iron	0.01	0.033	0.030	0.01	0.10	63	17	2.47	0.010	135	89	897	39	0.030	15100	62
Lead	0.0002	0.018	0.0003	0.00017	0.15	9	0.054	0.056	0.00020	0.23	25	0.88	0.41	0.069	4.92	40
Manganese	0.005	0.097	0.010	0.00015	0.52	63	20	4.6	0.037	85	90	110	9.1	0.16	2360	62
Nickel	0.0002	0.049	0.055	0.00077	0.121	19	0.38	0.21	0.050	1.8	87	1.9	0.36	0.050	15	60
Zinc	0.005	1.8	1.4	0.010	14	63	112	42	3.9	595	90	2069	104	2.2	21100	62

Note: Units in mg/L except for acidity/ alkalinity in mg CaCO<sub>3</sub> eq/L

Note 2: Detection limits were used for statistical purposes when values were less than detection.

Where detection limits were greater than 10x the minimum detection limit due to high ionic strength, non-detect results were excluded from statistical calculations.

Table 2 lists the seepage stations by each of the three water types. The results boxes in Figure 1 also indicate these groupings by colour.

The Type 1 seeps included samples from below the Upper Parking Lot dump (FD02), along the toe of the Northeast Dump (FD05, 06, and 07), a seep entering the pit below the Northeast Dumps FD 26), the Zone II East Dump (FD50), the Ramp Zone Dump (FD14), and the Upper Northwest Dump (FD16 and 18). According to the inventory of rock types presented in the 1996 ICAP report (RGI, 1996), these dumps generally contained relatively low proportions of sulphide waste rock, and higher proportions of calc-silicate or intrusive rock compared to other parts of the Faro Dump. The Intermediate Dump contains higher proportions of sulphide material. The seepage chemistry reflects some buffering by reactive carbonate minerals, which help to maintain neutral pH conditions. The sample collected from station FD26 in May 2007 was classified as a Type 2 seep due to an elevated zinc concentration of 20 mg/L; however, during the September 2007 sampling period the zinc concentration was much lower, nearly 2 mg/L, and so this seep was reclassified as a Type 1 seep.

The Type 2 seeps have included samples from several different areas, including ore and low grade ore stockpiles (FD01, 09, 10, 12, 31 and 38), the Main Dump West (FD30), the Main Dump East (FD08), the Intermediate Dump (FD44 , 48 and 49), the Lower Northwest Dump (FD19), seeps entering the pit below the Northeast Dumps (FD21, 22, 23, 24, 26 and 27) the Ramp Zone Dump (FD14), seeps in the mill area (FD32 and 35), and seeps from below the Faro Valley Dumps (FD40). A common element of all these areas is the presence of sulphides or oxidized schist. Although the pH is typically in the pH 6 to 7 range, it is clear that this drainage is strongly influenced by oxidation of sulphide minerals. However, many of these seeps contain high levels of calcium and magnesium, suggesting that there are still some carbonates present in the source materials. Samples from below the Main Dump West (FD12, Zn = 331 mg/L, FD31, Zn = 329 mg/L) contained the highest zinc concentrations of Type 2 seeps sampled in 2007, and were likely influenced in part by ore stockpiles upgradient of this location. Water quality trends at FD31 are included in Section 4.2.1 under X23 (equivalent station). Type 2 samples outside of the influence of the ore stockpiles and mill area had zinc concentrations below 90 mg/L in 2007.

The Type 3 seeps have included samples from the Oxide Fines Stockpile (FD04 and 46), low grade ore stockpiles (FD12 and 38), the Medium Grade Stockpile (FD37), the mill area (FD33, 34 and 35), the West Main Dump (FD30 and 36), the Intermediate Dump (FD13, 47 and 49), the Faro Creek Diversion dyke (FD20), and, on occasion, seeps entering the pit below the Faro Valley Dump (FD40),or the Northeast Dumps (FD21, 22, 23, 24 and 27). Portions of the waste rock or pit benches in all of the above areas contained sulphides or oxidized schist. The seepage quality indicates very little if any neutralizing minerals are available to control the pH and metal concentrations in these seeps. Samples from the Oxide Fines Stockpile (FD04, Zn = 1230 to 10,900 mg/L), the Medium Grade Stockpile (FD37, Zn = 6130 to 21,100 mg/L), and the mill area (FD33, Zn = 1110 to 6770 mg/L) had the highest zinc concentrations. However, zinc concentrations in the remaining acidic seeps ranged from 2.2 to 877 mg/L (overall median of 53 mg/L) indicating that seeps with high zinc concentrations occur in association with the sulphide waste rock cells and other sulphidic waste rock.

**Table 2: Seepage Stations Classified By Water Type**

Type 1		Type 2		Type 3	
(pH typically >6.5, Zn <6 mg/L)		(pH typically between 6 and 7, Zn concentrations ranging from 4 to 595 mg/L)		(pH <6, Zn typically >10mg/L)	
ID	Location	ID	Location	ID	Location
SRK-FD02	Upper Parking Lot Dump	SRK-FD1	Ore and Low Grade Ore Stockpiles	SRK-FD04	Oxide Fines Stockpile
SRK-FD05	Toe of Northeast Dump	SRK-FD8	East Main Dump	SRK-FD12 (1 of 8)	Ore and Low Grade Ore Stockpiles; West Main Dump
SRK-FD06	Toe of Northeast Dump	SRK-FD9	Ore and Low Grade Ore Stockpiles; West Main Dump	SRK-FD13	Intermediate Dump
SRK-FD07	Toe of Northeast Dump	SRK-FD10	Ore and Low Grade Ore Stockpiles; West Main Dump	SRK-FD20	Faro Creek Diversion
SRK-FD14 (7 of 8)	Ramp Zone Dump	SRK-FD12 (7 of 8)	Ore and Low Grade Ore Stockpiles; West Main Dump	SRK-FD21 (5 of 10)	Northeast Dumps towards Pit
SRK-FD16	Upper Northwest Dump	SRK-FD14 (1 of 8)	Ramp Zone Dump	SRK-FD22 (1 of 3)	Northeast Dumps towards Pit
SRK-FD17	Upper Northwest Dump	SRK-FD19	Lower Northwest Dump	SRK-FD23 (5 of 8)	Northeast Dumps towards Pit
SRK-FD18	Upper Northwest Dump	SRK-FD21 (5 of 10)	Northeast Dumps towards Pit	SRK-FD24 (1 of 12)	Northeast Dumps towards Pit
SRK-FD26 (11 of 12)	Northeast Dumps towards Pit	SRK-FD22 (2 of 3)	Northeast Dumps towards Pit	SRK-FD27 (1 of 4)	Northeast Dumps towards Pit
SRK-FD44 (1 of 4)	Intermediate Dump	SRK-FD23 (3 of 8)	Northeast Dumps towards Pit	SRK-FD30 (2 of 7)	West Main Dump
SRK-FD50	Zone II East	SRK-FD24 (11 of 12)	Northeast Dumps towards Pit	SRK-FD33	Mill
		SRK-FD26 (1 of 12)	Northeast Dumps Towards Pit	SRK-FD34	Mill
		SRK-FD27 (3 of 4)	Northeast Dumps towards Pit	SRK-FD35 (2 of 4)	Mill
		SRK-FD30 (5 of 7)	West Main Dump	SRK-FD36	West Main Dump
		SRK-FD31	Ore + Low Grade Ore Stockpiles, West Main Dump	SRK-FD37	Medium Grade Stockpile
		SRK-FD32	Mill	SRK-FD38 (1 of 2)	Ore and Low Grade Ore Stockpiles
		SRK-FD35 (3 of 4)	Mill	SRK-FD40 (6 of 11)	Faro Valley Dump
		SRK-FD38 (1 of 2)	Ore + Low Grade Ore Stockpiles	SRK-FD46	Oxide Fines Stockpile, Mill
		SRK-FD40 (5 of 11)	Faro Valley Dump	SRK-FD47	Intermediate Dump
		SRK-FD44 (2 of 4)	Intermediate Dump	SRK-FD49 (1 of 2)	Intermediate Dump
		SRK-FD48	Intermediate Dump		
		SRK-FD49 (1 of 2)	Intermediate Dump		

Note: Where water type at a sampling station was variable over time, the number of samples of each type is shown, along with the total number of samples collected at a particular station. For example, eight samples have been collected at SRK-FD14, with seven samples having Type 1 water quality, and one sample having Type 2 water quality.

Management activities upgradient of FD40 during Summer 2006 may have altered the chemical conditions at this station. In an effort to neutralize acid seepage and mitigate elevated zinc concentrations, a quantity of coarse crushed limestone was used to line an upgradient seepage pathway above the Faro Valley Dump. A review of the pH, alkalinity, Ca, Mg, and Zn concentrations over time indicates that the influence of the limestone on the water chemistry at FD40 is negligible.

#### **4.1.2 Grum Waste Rock Dumps**

The 2007 Grum Dump seepage surveys were conducted by walking the external toe of the dump and selected internal dump toes and looking for toe seeps, as previously described. No new seeps were identified in 2007.

##### **Water Types**

Most Grum toe seeps sampled in previous years had neutral to slightly alkaline pHs, and would be classified as Type I seeps under the system described for Faro. However, further division is possible on the basis of sulphate and zinc concentrations.

- Type 1a seeps had low to intermediate sulphate (7.0 to 840 mg/L) and low zinc concentrations (<0.005 to 0.4 mg/L). These seeps reflect drainage from calcareous phyllites and till in the northwest draining portion of the dump. Surface mapping in this drainage indicated some sulphides were present in this area, but they were typically in small isolated pockets, and were surrounded by extensive areas of calcareous phyllites.
  - One station (SRK-GD18) returned a field pH of 6.7 and a zinc concentration of 0.39 mg/L in May 2005. These May 2005 results represent the lowest pH and the highest zinc concentrations observed to date in seepage from the northwest portion of the dump. In May 2006 and 2007, SRK-GD18 returned higher pH values (7.1 and 7.7, respectively) and lower zinc concentrations (0.13 and 0.03 mg/L, respectively).
  - An adjacent seep (SRK-GD13) had a field pH (6.9) lower than previously observed and a sulphate concentration (730 mg/L) higher than previously observed during May 2006 monitoring, along with a dissolved zinc concentration near the upper limit of the previously observed range. This station represents the only major flow from the Southwest Dump. In September 2006, the field pH was slightly higher than in May, and sulphate concentrations were slightly lower. During May and September 2007 monitoring, the field pH at this station remained relatively high (7.7 and 7.4, respectively), and zinc concentrations relatively low (0.011 and 0.016 mg/L); however, the sulphate concentration recorded in September 2007 was higher than previously observed (840 mg/L). There is a weak trend of increasing sulphate concentrations over the 2002 (313 to 386 mg/L) through 2007 (660 to 840 mg/L) period; magnesium and calcium concentrations show a similar pattern, with magnesium concentrations doubling over the 2002-2007 monitoring period. Zinc concentrations do not show this same pattern at this time. Nickel concentrations up to

0.13 mg/L have been observed at SRK-GD13, which is the maximum nickel concentration observed from Type 1a seeps at Grum.

- Type 1b seeps generally had zinc concentrations in the range of 2 to 7 mg/L, and sulphate concentrations greater than 500 mg/L. Most of these seeps were towards the southeast, and were below the sulphide cell. However, SRK-GD11, which was theoretically upgradient of the sulphide cell, also fell into this group until 2005. Waste rock mapping completed in September 2002 indicated that significant amounts of sulphide were present above this location, and that sulphidic waste rock was not limited to the sulphide cell. In May 2006, SRK-GD11 returned a dissolved zinc concentration of 17.4 mg/L and a field pH of 6.45 and was reclassified as Type 2 using the Faro system. Conditions had improved by May 2007, with a zinc concentration of 8 mg/L and the field pH of 6.9; however water quality at this station appears to be transitional between Type 1b and Type 2 water classification.
- Station SRK-GD17, sampled at the toe of an internal dump lift near the sulphide cell, is also classified as Type 2. The zinc concentration at this station in May 2007 was considerably lower than observed in 2004 and 2006, but remained elevated (12 mg/L).
- Station SRK-GD16, which flows intermittently, was sampled in May of 2004, 2005 and 2007. The May 2005 sample returned the highest zinc concentration (59.9 mg/L) observed from the southern toe of the Grum Dump to date. The zinc concentration at this station in May 2007 was considerably lower than observed in 2005, but remained elevated at 16.5 mg/L.
- One sample of Type 3 water (SRK-GD19) was collected in Spring 2004 from the toe of an internal lift. Under the Faro system, this sample would be considered transitional between Type 2 and Type 3, based on a dissolved zinc concentration of 107 mg/L and a field pH of 5.7. This station had been dry during subsequent surveys until Spring 2007 when the reported zinc concentration was significantly lower (9.23 mg/L) and was reclassified as Type 2 for that sampling round.
- Station SRK-GD01 returned dissolved zinc concentrations of 17 mg/L and 11 mg/L in Spring 2004 and Spring 2005 respectively. These spring results are elevated over previously observed zinc concentrations at this station, although sulphate concentrations and field pH were within the range of previous results. However, in 2006 and 2007, SRK-GD01 returned zinc concentrations between 2.6 and 3.1 mg/L during both spring and fall sampling periods. These are typical zinc concentrations at SRK-GD01 for the 2002-2007 monitoring period, and appear to indicate that higher concentrations recorded in Spring 2004 and 2005 were not indicative of a trend. As the flow at SRK-GD01 is the single largest source of seepage from Grum Dump, water quality at this station will continue to be closely monitored.

Figure 2 shows the location of seep sampling stations, and provides a graphical summary of the distribution of the different seepage types. Table 3 provides a summary of key characteristics for each of the above seepage types. Full seepage results are included in Appendix A.





## 4.2 Routine Monitoring

### 4.2.1 Faro Site

Locations of the routine seepage monitoring stations are shown in Figure 1. Data for these stations are available in the Anvil Range master water quality database maintained by GLL. Graphs of key parameters are provided in Figures 6 through 11. Plotted values typically represent dissolved concentrations; where only total concentration was reported, this value was used for plotting purposes. Where reported concentrations were below method detection levels, the detection level values were plotted; this causes plotted values for some trace metals to decline over time as lower detection levels were achieved, and is not necessarily indicative of declining concentrations of these parameters.

#### *Station X23*

Station X23 is located east of the mill, in the original Faro Creek channel, below the East Main Dump, the Oxide Fines Stockpile, and the Medium Grade Stockpile. Interpretation of trends in this data is complicated by the deposition and removal of some of the stockpiled ore, and by leakage of pit water into the old Faro Creek channel from a drainage ditch during operations (RGI 1996).

This station has been monitored for select parameters since 1986, and for a full suite of parameters since 1989. The data are summarized in Figure 6.

Sulphate concentrations increased in stages, from less than 2000 mg/L in 1986 to approximately 6000 mg/L in 2000/2001, followed by a slight decrease to approximately 4000 mg/L in 2002 and 2003. From mid-2004 through to the beginning of 2007, sulphate concentrations at X23 varied between approximately 5000 and 7000 mg/L, and have levelled out at approximately 5000 mg/L during 2007 (Figure 6). The higher sulphate concentrations observed since 2004 correspond to a slight decrease in pH from an average of approximately 7 to an average of approximately 6.5, and changes in the major ion chemistry from calcium-dominated to magnesium-dominated. Peak concentrations of iron (200 mg/L), zinc (1000 mg/L), and several other metals (e.g. aluminum, cadmium, cobalt, copper, and nickel) were also observed during this period. These observations indicate that the seepage chemistry continues to be in a state of transition, and that concentrations could continue to increase over time. However, concentrations of most parameters exhibited similar seasonal patterns and magnitudes in both 2004 and 2005, and have generally demonstrated decreasing trends since that time.

#### *Station X26*

Station X26 is the discharge from the dewatering sump for the Zone II pit, which is completely filled in and buried by waste rock. The sump is operated over an approximately 3 month period during the summer, and water levels are allowed to fluctuate between depths of 60 and 52 metres below the current ground surface. The variable water table in the backfilled Zone II pit causes repeated wetting and submergence of backfill and former pit walls, and sulphides in the former pit walls and backfill may contribute oxidation products to the water sampled at X26. Results from this station are shown in Figure 7.

Monitoring data showed an initial decrease in sulphate concentrations over the first few years of monitoring (1991 to 1995), followed by a gradual and relatively steady increase with concentrations reaching a maximum of 3400 mg/L in 2005, and then beginning to decrease to approximately 2600 mg/L in 2007. pHs have been relatively stable, in the range of 6 to 6.5. Zinc concentrations also peaked in 2005 at 130 mg/L and have since decreased slightly. Iron concentrations followed a similar trend.

There is a weak seasonal pattern in metal concentrations, with generally higher concentrations occurring later in the summer, when dewatering levels are at a minimum. Based on this variation, it is possible that there is some stratification of concentrations in the flooded zone of the pit, or that oxidation products leached from backfill and pit walls are responsible for this increase.

#### *Station FCO/A30*

Station A30 (SRK-FD40) is located along the north wall of the pit, below the Faro Valley Dump. The current location is accessed by hiking down from the Faro Valley Dump. In earlier years, the station was a sump, which may have received drainage from other seepage along the pit walls. Station FCO is located immediately upstream of the Faro Valley Dump, and maybe somewhat influenced by leakage through mineralized rock in the road and berm that comprise the Faro Creek Diversion. Both stations have relatively high flows, and these flows contact waste rock along the flowpath between A30 and FCO. Results for these two stations are provided in Figure 8.

Sulphate concentrations at A30 increased from approximately 200 mg/L in 1989 to 600 mg/L in 1997, and ranged from 100 to 600 mg/L for the subsequent period through 2003. Average and peak sulphate concentrations at A30 increased significantly beginning in 2004, with a maximum concentration of 1340 mg/L recorded in May 2005 before demonstrating a decreasing trend. Sulphate concentrations in 2007 ranged from 270 to 430 mg/L. pHs decreased from pH 7.5 in 1989 to pH 3 in 1998. Monitoring from 2001 through 2006 recorded fluctuating pHs, with several pH values near 3 and several in the range of 5.6 to 7.5. 2007 pH values were all approximately 3.3. Zinc concentrations followed similar trends to sulphate, reaching peaks of approximately 100 mg/L in 1998, and then varying from approximately 10 to 100 mg/L since that time. The wide variations in seepage chemistry may reflect differences in the amount of vertical versus lateral flow through the pile. In addition, interpretation of water quality results since 2006 may be confounded by alkaline additions from placement of limestone upgradient during summer 2006, as discussed in Section 4.1.1.

#### *Station SP5/6*

Station SP5/6 (equivalent to SRK-FD26) is located below the Upper North East Dump. Flows at this station are substantial, and may be derived from leakage from the Faro Creek Diversion. Results for this station are provided in Figure 9.

Historically, sulphate concentrations have ranged from 200 mg/L to 1000 mg/L, and have been quite variable. Samples from 2006 and 2007 had sulphate concentrations ranging from 360 to 540 mg/L. Zinc concentrations ranged from approximately 2 mg/L to 5 mg/L, and calcium, magnesium, and zinc concentrations were strongly correlated with sulphate. Variations in concentrations appear to be weakly correlated to flows, with lower concentrations occurring in the June samples (high flow), and higher concentrations occurring in the fall samples (low flow).

#### *Stations NE1, NE2 and W5*

Stations W5, NE1 and NE2 are located along the toe of the Northeast Dump. W5 was monitored from 1989 to 1991, and then again since 2004, and was probably sampled close to the toe of the dump. Therefore, it is likely that this station is equivalent to SRK-FD05 and -FD06. NE1 and NE2 are monitored in the regular seepage program, but are collected approximately 100 metres downstream of the toe, and may be influenced by interaction with the overburden in that area. Results for all three stations are provided in Figure 10. Data from SRK-FD05 and FD06 are shown in the W5 graphs.

Sulphate concentrations in W5 increased from approximately 300 mg/L in 1989 to approximately 400 mg/L in the more recent samples. One sample collected from SRK-FD06B had a peak value of 1160 mg/L sulphate in spring 2006. Sulphate concentrations in 2007 samples ranged from 200 to 700 mg/L. Zinc concentrations increased from much less than 1 mg/L in 1989/90 to seasonably-variable concentrations of 0.5 to 5 mg/L in the last few years with a peak Zn value of 14.4 mg/L collected from SRK-FD06B in spring 2006. Field pHs during this period were neutral, and concentrations of other metals were generally low.

Sulphate concentrations at NE1 were stable within the range of 70 to 130 mg/L from 1997 through 2003, and then increased to range between 100 and 290 mg/L since 2004. Metal concentrations at this station were low throughout the monitoring period.

Sulphate concentrations in NE2 increased from approximately 100 mg/L in 1997 to variable concentrations between 400 and 900 mg/L since mid-2000. Metal concentrations at this station were low throughout the monitoring period.

#### *Station W8/W10*

Station W8 is located in Upper Guardhouse Creek, which flows under and alongside an approximately 50 metre section of the Northwest Dump. This station is equivalent to SRK-FD16. Station W10 is located approximately 100 metres upstream from this station and is unaffected by mining activities. Results for W8 (including data from SRK-FD16), and W10 are provided in Figure 11.

Results for both stations indicate consistently neutral pHs, low sulphate and low metal concentrations. There is little if any increase in concentration between W10 and W8.

#### 4.2.2 Grum Site

The routine monitoring stations at Grum are shown in Figure 2. Station V2 has been monitored on a regular basis since 1988, with monitoring at V2A since 1997 and at V15 since 1995. The routine stations are located along the road access and are between 200 and 800 metres below the toe of the dumps, where dilution of toe seepage by surface water and interaction of toe seepage with soils along the flow-paths could be expected. As such, results from these stations are not directly comparable to seepage at the toes of the dumps. The routine seepage monitoring data were available from the Anvil Range master water quality database maintained by GLL. Graphs of key parameters are provided in Figures 12 and 13. Plotted values represent dissolved concentrations; where only total concentration was reported, this value was used for plotting purposes. Where reported concentrations were below method detection levels, the detection level values were plotted, causing plotted values for some trace metals to decline over time as lower detection levels were achieved. As such, these changes are not necessarily indicative of declining concentrations of these parameters. Results from these stations have also been discussed as part of the 2007 Adaptive Management Plan (AMP) Event #4 report (SRK 2008).

##### *Stations V15 and V2*

Station V15 represents surface runoff and groundwater seepage discharging from a sedimentation pond below the toe of the waste dump at the upper limit of Tributary A. The area immediately upgradient is a groundwater discharge zone for one of the Grum Dump catchments, and toe seeps at stations SRK-GD04 and SRK-GD21 are intermittently present as surface discharges. Water at station V15 has been in close contact with soils and sediments, and chemical interaction with these materials is likely significant. Results for these stations are provided in Figure 12.

Sulphate concentrations at Station V15 increased gradually between 1996 and 2000 (from 100 mg/L to 300 mg/L), and then more rapidly in 2000 and 2001, reaching levels in the range of 1000 mg/L by June 2001. Sulphate concentrations remained at approximately 1000 mg/L until 2004, and then increased to approximately 2000 mg/L by fall of 2005, with a March 2005 spike to approximately 3000 mg/L. Sulphate concentrations remained at approximately 2000 mg/L through 2006 and ranged from 1400 to 1900 mg/L through 2007. The increase in sulphate concentrations corresponded to increases in calcium and magnesium concentrations, with calcium concentrations generally exceeding magnesium concentrations. pHs had been stable in the range of 7.5 to 8 until the beginning of 2006 when they began to decrease to as low as 6.3 in the fall of 2007. There has been a corresponding increase in zinc concentrations as pH has dropped from concentrations generally much less than 0.1 mg/L prior to 2006 to a high of 1.2 mg/L recorded in October 2007. Over this same time period, nickel concentrations have increased from 0.01 mg/L to 0.07 mg/L and cadmium concentrations have increased from <0.0002 mg/L to 0.001 mg/L.

Station V2 is located upstream of Vangorda Creek in the original Grum Creek channel, and is fed by discharge from the sedimentation pond at V15, as well as by seepage losses from the Grum Creek

diversion and local runoff. During 2007, discharge from V15 was diverted by a ditch to Moose Pond via Grum Creek. Minor seepage losses from the ditch continue to report to V2.

Sulphate concentrations at V2 increased from less than 50 mg/L in the late 1980s (i.e. prior to dump construction) to approximately 150 mg/L in 1998. In 1998, concentrations started to increase rapidly, reaching 600 to 700 mg/L by 2000. Sulphate concentrations remained between 600 and 700 mg/L from 2000 through mid-2004 then increased to approximately 1000 mg/L since 2005. pHs were in the range of 7 to 8.5 throughout this period. The increase in sulphate concentrations was accompanied by an increase in both calcium and magnesium concentrations. Calcium remained the dominant cation. Zinc concentrations at station V2 have been highly variable, with typical concentrations ranging from less than 0.01 to 1 mg/L prior to 2002, and from 0.005 to 0.05 mg/L since 2002. Concentrations of other metals (e.g. cadmium, iron, cobalt, copper, and nickel) were variable, but generally low, and have remained constant or declined in recent years.

#### *Station V2A*

Station V2A represents the downstream extent of the largest seepage flows that originate at the toe of the Grum dump, at stations SRK-GD01 and SRK-GD02. In addition, diverted water from Station V15 began reporting to Grum Creek above the sampling point in mid-2007. Station V2A is sampled where water diverted from Grum Creek discharges into Moose Pond via a culvert. Sample frequency was irregular at this station prior to 2002. Results for this station are provided in Figure 13.

Sulphate concentrations at Station V2A increased erratically between 1996 and 1999 (from 100 mg/L to 400 mg/L). No monitoring was conducted at V2A during 2000, and sulphate concentrations had jumped to 800 mg/L by the next monitoring event in June 2001. Sulphate concentrations at V2A then varied seasonally between 300 and 1000 mg/L from 2001 to late 2007, demonstrating a slightly increasing trend over time (2007 samples ranged from 750 to 1000 mg/L, with the highest sulphate concentrations likely influenced by the diversion of water from V15). The seasonally-fluctuating sulphate concentrations were mirrored by calcium and magnesium concentrations over the 2001-2007 period. pHs were stable in the range of 7.5 to 8.5 throughout the monitoring period.

Zinc concentrations at Station V2A were highly variable, with typical concentrations ranging from less than 0.01 to 0.1 mg/L prior to 2000. Zinc concentrations from 2001 to present appeared to follow a pattern of annual peaks during June freshet (0.2 to 3.4 mg/L) and lows associated with late winter flows (0.05 to 0.08 mg/L). Zinc concentrations at V2A peaked during summer 2005, ranging from 0.7 to 1.3 mg/L. 2007 concentrations show a similar seasonal pattern with somewhat lower absolute concentrations and a smaller range of values (0.03 to 0.23 mg/L). Other metal concentrations (e.g. cadmium, iron, cobalt, copper, and nickel) were variable, but generally low, and did not change significantly over time.

#### 4.2.3 Vangorda Site

The routine seepage monitoring stations at Vangorda are the three drains shown in Figure 2, as well as a Vangorda Creek monitoring station below all mine inputs (not shown on Figure 2). Results for these stations were available from the Anvil Range master water quality database maintained by GLL. Graphs of key parameters are provided in Figures 14 to 17. Plotted values represent dissolved concentrations; where only total concentration was reported, this value was used for plotting purposes. Where reported concentrations were below method detection levels, the detection level values were plotted; this causes plotted values for some trace metals to decline over time as lower detection levels were achieved, and is not necessarily indicative of declining concentrations of these parameters.

##### *Drains 3, 5 and 6 (Stations V30, V32 and V33)*

Three of the drains (Drain 3, 5 and 6) at Vangorda have been monitored as part of the routine monitoring programs since 1994. The other drains (Drains 1, 2 and 4) have historically been consistently dry. Results are shown in Figures 14 through 16.

Results for station V30 (Drain 3, SRK-VD03) are provided in Figure 14. Seepage from this station has had pHs close to 6 throughout the monitoring period. Sulphate concentrations increased from 2500 mg/L in 1994 to a peak concentration of 6200 mg/L in fall 2004 and had decreased to approximately 4200 mg/L in 2007. Zinc concentrations were highly variable in the range of 150 to 600 mg/L from 1994 through spring 2000. Station V30 was not monitored again until spring 2002 and zinc concentrations have been more stable since that time, in the range of 200 to 600 mg/L.

Results for station V32 (Drain 5, SRK-VD04) are provided in Figure 15. pHs at this station have decreased from approximately 5 in 1994 to approximately 3 over the 2003 through 2006 period, with one value of 2.2 reported. 2007 pH values in 2007 ranged from 2.9 to 3.8. Sulphate concentrations increased substantially during this period from approximately 5000 mg/L in 1994 to greater than 65,000 mg/L in 2006/2007, with a peak of 93,000 mg/L reported in 2004. The high sulphate concentrations were supported by a similar magnitude of increase in magnesium concentrations (from approximately 500 mg/L in 1994 to 5500 mg/L in 2007) as well as substantial increases in iron, manganese and zinc concentrations. Recent iron and manganese concentrations have been in the range of 3000 to 4000 mg/L, and zinc concentrations have been approximately 11000 mg/L. Concentrations of aluminum increased until 2005 and have fluctuated between 150 and 350 mg/L since then. Cadmium increased to a peak concentration of 18 mg/L in 2004 and ranged between 5 and 10 mg/L in 2007. Cobalt (30 mg/L) and nickel (20 mg/L) concentrations were relatively stable over the 2002 through 2006 monitoring period. Concentrations of copper have generally decreased from a maximum of 0.5 mg/L in 2003; however a peak value of 2 mg/L was recorded in October 2007. Lead concentrations were highest between 1997 and 2000 (10 mg/L) and have decreased to approximately 1 mg/L since 2005 with one value of 3 mg/L recorded in October 2007.

Results from Station V33 (Drain 6, SRK-FD05) are provided in Figure 16. pHs were in the range of 6 to 7 from 1994 to 1999, dropped to approximately 6 in the 2000 to 2002 samples, and further decreased to between 5 and 6 in 2004 and 2005, with a single low value of 3.2 reported in June 2005. 2006 and 2007 pH values were slightly higher, ranging from 5.7 to 6.6. Sulphate concentrations were in the range of 2000 to 5000 mg/L from 1994 to 2000, and then increased to a peak of approximately 90,000 mg/L in 2004, and concentrations of approximately 50,000 mg/L in 2007. Magnesium concentrations increased over time, beginning in 2001, and were correlated with sulphate concentrations. Concentrations of iron, zinc and several other metals have increased slightly since 2000, with zinc concentrations generally in the range of 3000 to 10,000 mg/L in recent years.

#### *Station V27*

Station V27 is the first routine monitoring station in Vangorda Creek downstream of the Vangorda and Grum Waste Rock dumps and other mine-related sources. Although interpretation of the surface water sampling stations is not directly relevant to the seepage monitoring program, the low flows observed in seeps from the Vangorda Dump suggest that some water could be leaving the system via the subsurface. Station V27 is sufficiently lower in elevation that it could be reasonably expected to reflect any contaminants that are transported via the groundwater system. However, Vangorda Creek flows are very large, so significant changes in loading would be needed before any changes from the dumps could be observed. Also, arrival of Vangorda Dump contaminants at Vangorda Creek may not yet have occurred, as groundwater movement in the soils underlying the dump is expected to be slow. In addition, any loading entering the creek from the Grum dump would contribute to loads at this location. Results for station V27 are presented in Figure 17.

Results for this station indicated consistently neutral pHs, and generally low sulphate concentrations from 1991 to 2002. A sulphate concentration of 125 mg/L was recorded in spring 2002 but concentrations have been decreasing since that time to approximately 50 mg/L since 2005. Metal concentrations were generally very low, and have exhibited a slightly decreasing trend since 2002. Copper and zinc concentrations in Vangorda Creek were near or above CCME guidelines for aquatic life, and detection limits for several other parameters were too high to evaluate whether they meet CCME criteria.

### **4.3 Temperature and Oxygen Monitoring of Waste Rock**

There were no significant changes in the temperature and oxygen monitoring data compared to data from previous years (see SRK 2007). A discussion of the thermal and oxygen data will be included in the report being prepared to summarize the findings of Task 29 – Supplemental Geochemical Studies.

## 5 Conclusions

Seeps associated with the Faro Waste Rock Dumps showed a wide range of pH and zinc concentrations. The highest zinc concentrations (>200 mg/L, as high as 21,100 mg/L) were associated with the ore stockpiles and mill area. High zinc concentrations were also associated with the sulphide cells on the Main Dumps (up to 900 mg/L). A large number of seeps associated with the waste rock were acidic or partially buffered, and had zinc concentrations in the range of 5 to 600 mg/L. A moderate number of seeps at Faro had alkaline pHs, and zinc concentrations of less than 5 mg/L. These were associated with dumps that contained relatively little sulphide waste rock. Cadmium, cobalt, copper, iron, manganese and nickel concentrations in some of the seeps are at concentrations that significantly exceed receiving water quality criteria.

Seeps associated with the Grum Waste Rock Dumps had consistently neutral to alkaline pHs. Seeps draining to the southeast had zinc concentrations generally in the range of 2 to 7 mg/L and elevated sulphate concentrations. Seeps draining to the northwest had zinc concentrations ranging from <0.005 to 0.4 mg/L, dissolved nickel concentrations up to 0.13 mg/L, and generally lower, but increasing, sulphate concentrations. The seeps to the Southeast were located below the sulphide cell, or below sulphidic waste identified in the SRK September 2002 surface mapping programs. Zinc concentrations remained within stable ranges at all southeast draining seeps in 2007. No other trace metals of concern have been identified in the Grum seepage to date.

Seeps associated with the Vangorda Waste Rock Dumps were acidic to partially buffered, and contained high to very high zinc concentrations (from 10 to 13,100 mg/L). Sulphate and zinc concentrations have increased since the routine seepage monitoring programs were initiated in 1994, but concentrations appear to have stabilized within broad ranges of seasonal variation since 2004. Other trace metals significantly exceeding receiving water quality criteria in the Vangorda seepage include aluminum cadmium, cobalt, copper, iron, manganese and nickel. Cobalt and nickel are notably higher compared to acidic seeps at Faro.

Routine monitoring station X23 below the Faro Main Dump showed similar seepage chemistry to that observed in 2006. Sulphate, major ion and trace metal concentrations have generally demonstrated decreasing trends since 2004/2005.

Routine monitoring of stations V2 and V2A (southeast of Grum Dump) in 2007 showed similar pH conditions and major and trace element concentrations as observed in 2006. However, the dramatic increase in zinc concentration at station V15 (upgradient of V2) has continued, with dissolved zinc concentration steadily increasing from the previously-observed maximum of about 0.7 mg/L in December 2006 to nearly 1.2 mg/L in October 2007. The trend in dissolved zinc concentrations at station V15 appears to show a classic case of breakthrough of an attenuated chemical species, and the results reviewed to date do not suggest that the breakthrough process has run its course. For a

detailed discussion of the evolution of Grum Dump seepage chemistry, see the 2007 AMP Event #4 Response: Status Report (SRK, 2008).

Pit seepage monitoring was carried out during 2007 to supplement the existing pit seep monitoring record. Detailed evaluation of the 2007 pit seep monitoring results was not carried out.

There were no significant changes in the temperature and oxygen monitoring data compared to data previous years (see SRK 2007). Therefore, further interpretation of the data is not warranted at this time.

This report, "**1CD003.090 – Anvil Range Mining Complex- 2007 Waste Rock and Seepage Monitoring Report: 2007/08 Task 17 – DRAFT**", was prepared by SRK Consulting (Canada) Inc.

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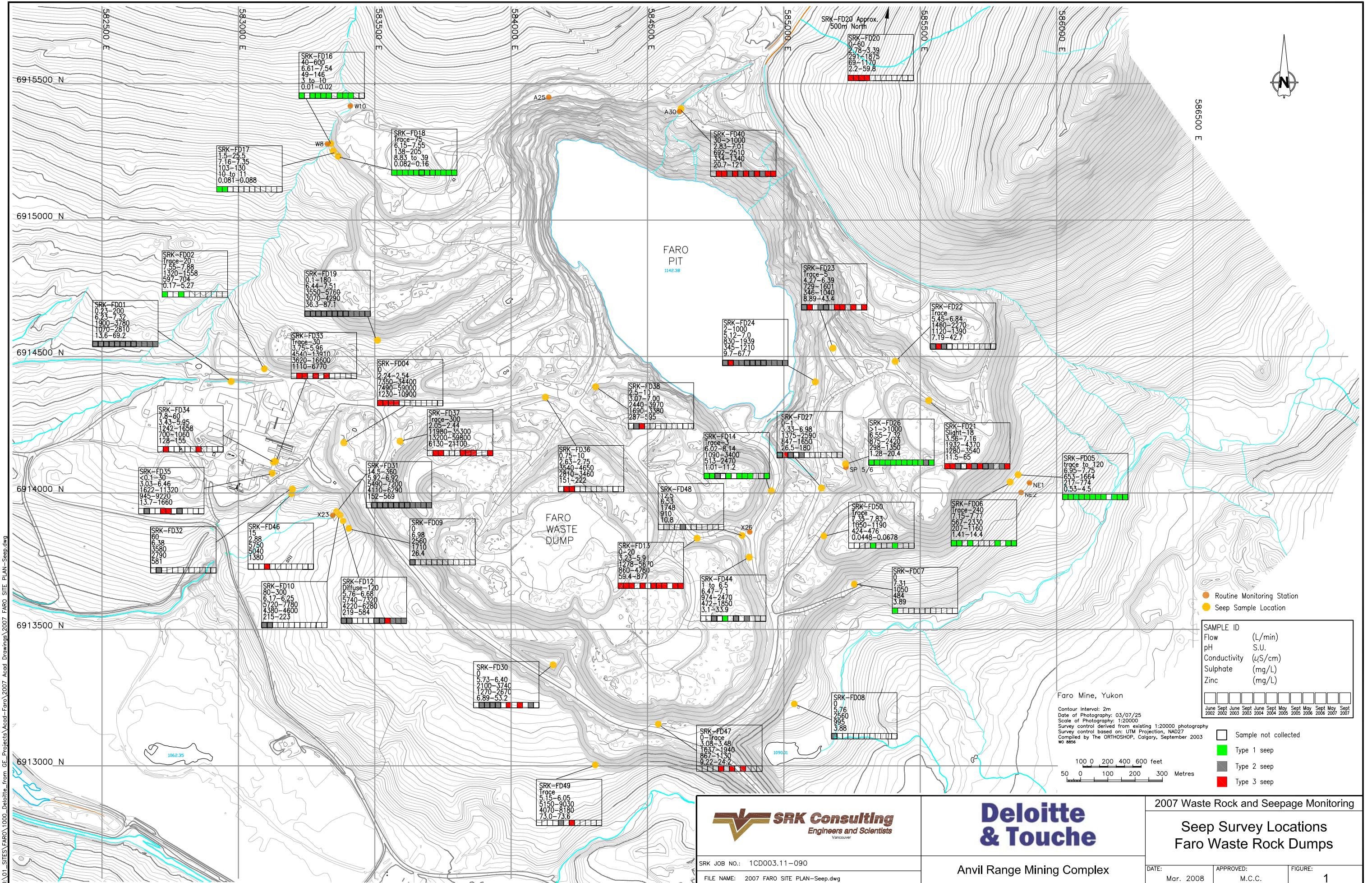
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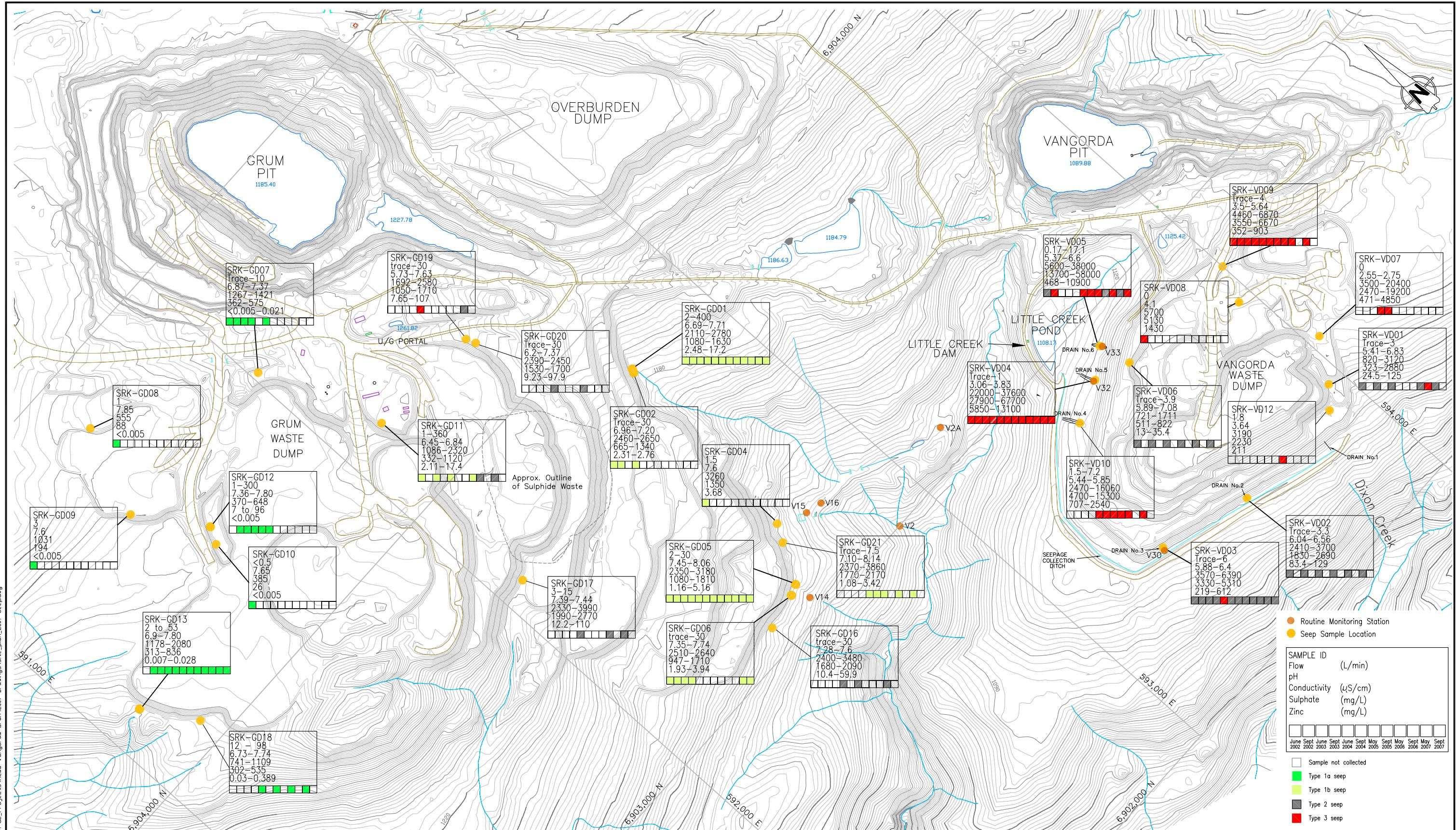
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Principal Environmental Geochemist

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## **Figures**





J:\OU\STEST\ARD\1000\SeepSite\Drawings\Acad\Vangorda\Acad\Site\_Plan\_2007\_Seep.dwg

Contour Interval: 2m  
Date of Photography: 03/07/25  
Scale of Photography: 1:20000  
Survey control derived from existing 1:20000 photography  
Survey control based on UTM Projection, NAD27  
Compiled by The ORTHOSHOP, Calgary, September 2003  
W0 8856

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SRK JOB NO.: 1CD003.11-090

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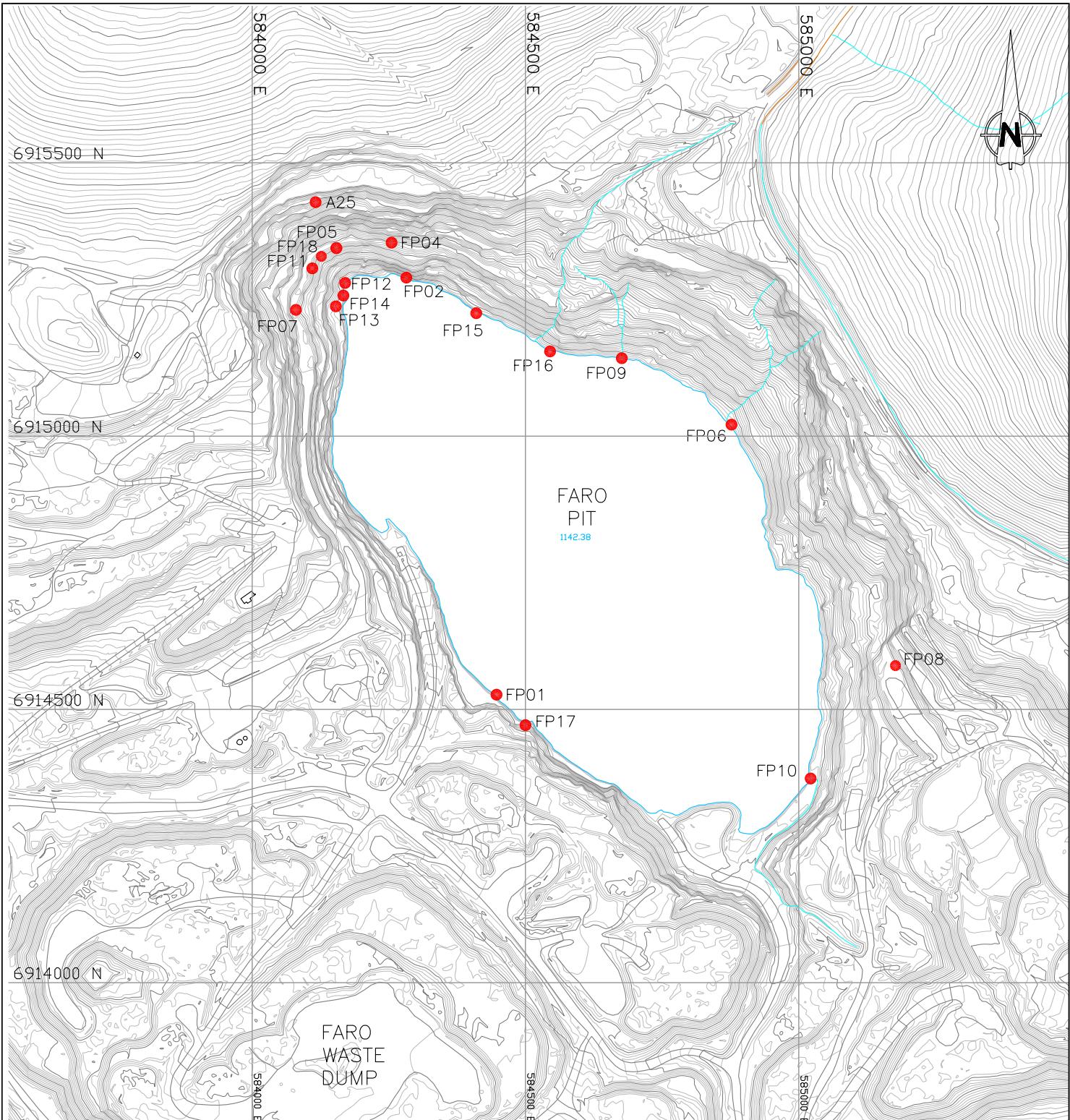
**Deloitte & Touche**

Anvil Range Mining Complex

2007 Waste Rock and Seepage Monitoring

Seep Survey Locations  
Vangorda & Grum  
Waste Rock Dumps

DATE: Mar. 2008 APPROVED: M.C.C. FIGURE: 2

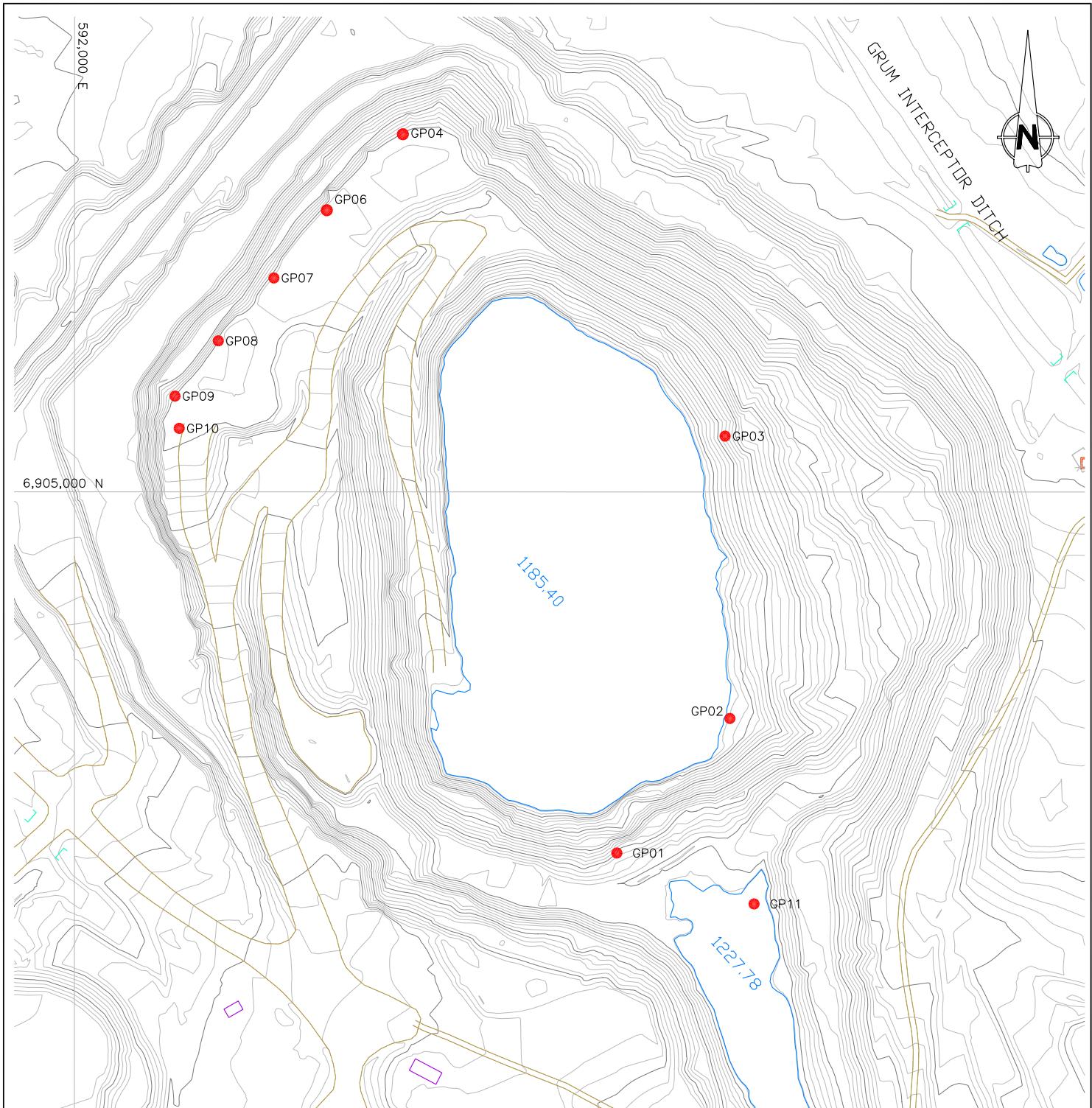


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 Survey control based on: UTM Projection, NAD27  
 Compiled by The ORTHOSHOP, Calgary, September 2003  
 WO 6856

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● Seep Sample Location

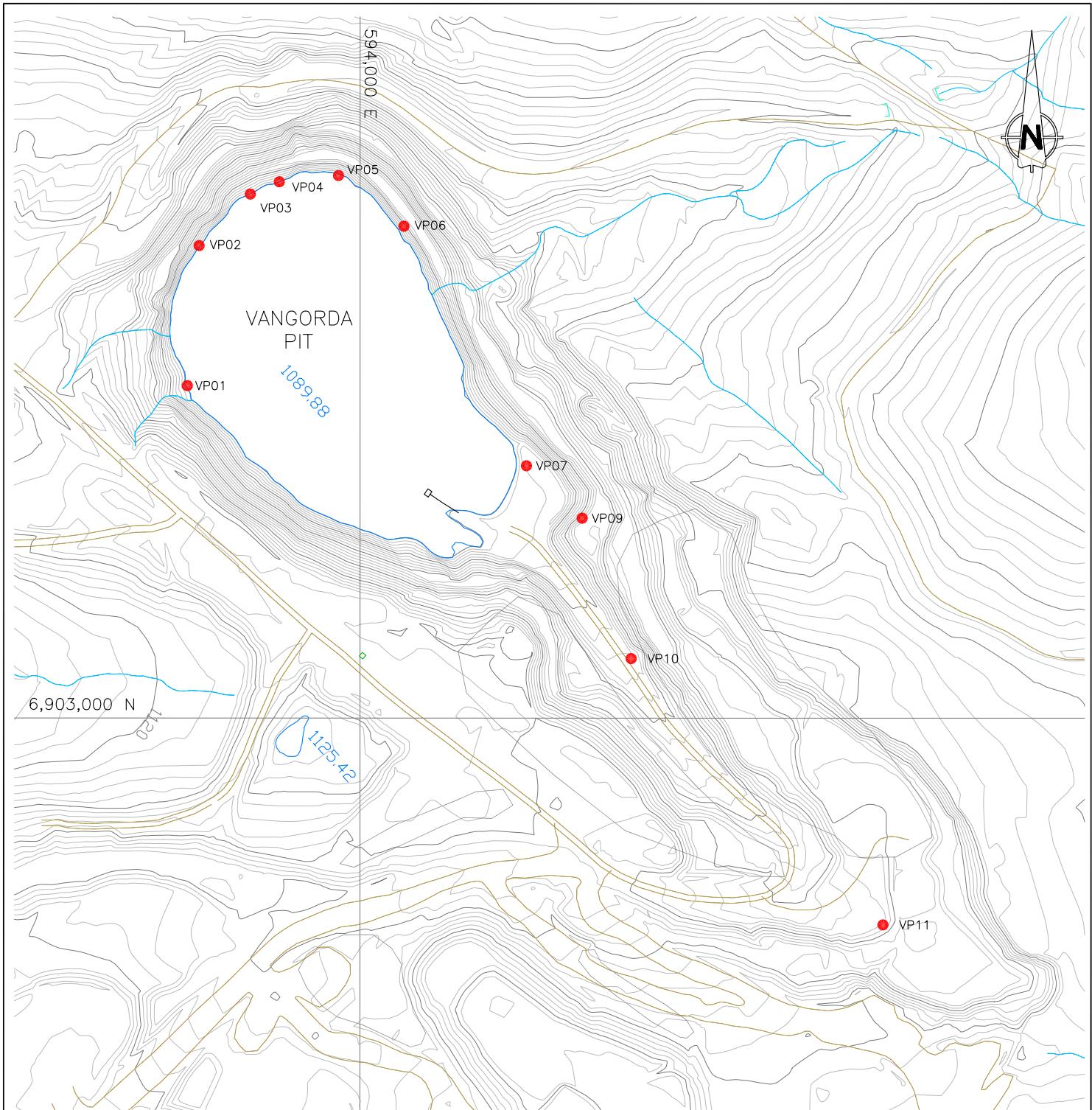
 <p>SRK Consulting Engineers and Scientists Vancouver</p>	 <p>Deloitte &amp; Touche</p>	2007 Waste Rock and Seepage Monitoring
		Seep Survey Locations Faro Pit
SRK JOB NO.: 1CD003.11-090	Anvil Range Mining Complex	DATE: Mar. 2008 APPROVED: M.C.C. FIGURE: 3
FILE NAME: 2007 FARO SITE PLAN-Seep.dwg		



● Seep Sample Location

Contour Interval: 2m  
Date of Photography: 03/07/05  
Scale of Photography: 1:20000  
Survey control derived from existing 1:20000 photography  
Survey control based on: UTM Projection, NAD27  
Compiled by The ORTHOSHOP, Calgary, September 2003  
WO 8866

0 50 100 150 200 250  
Scale 1:5000

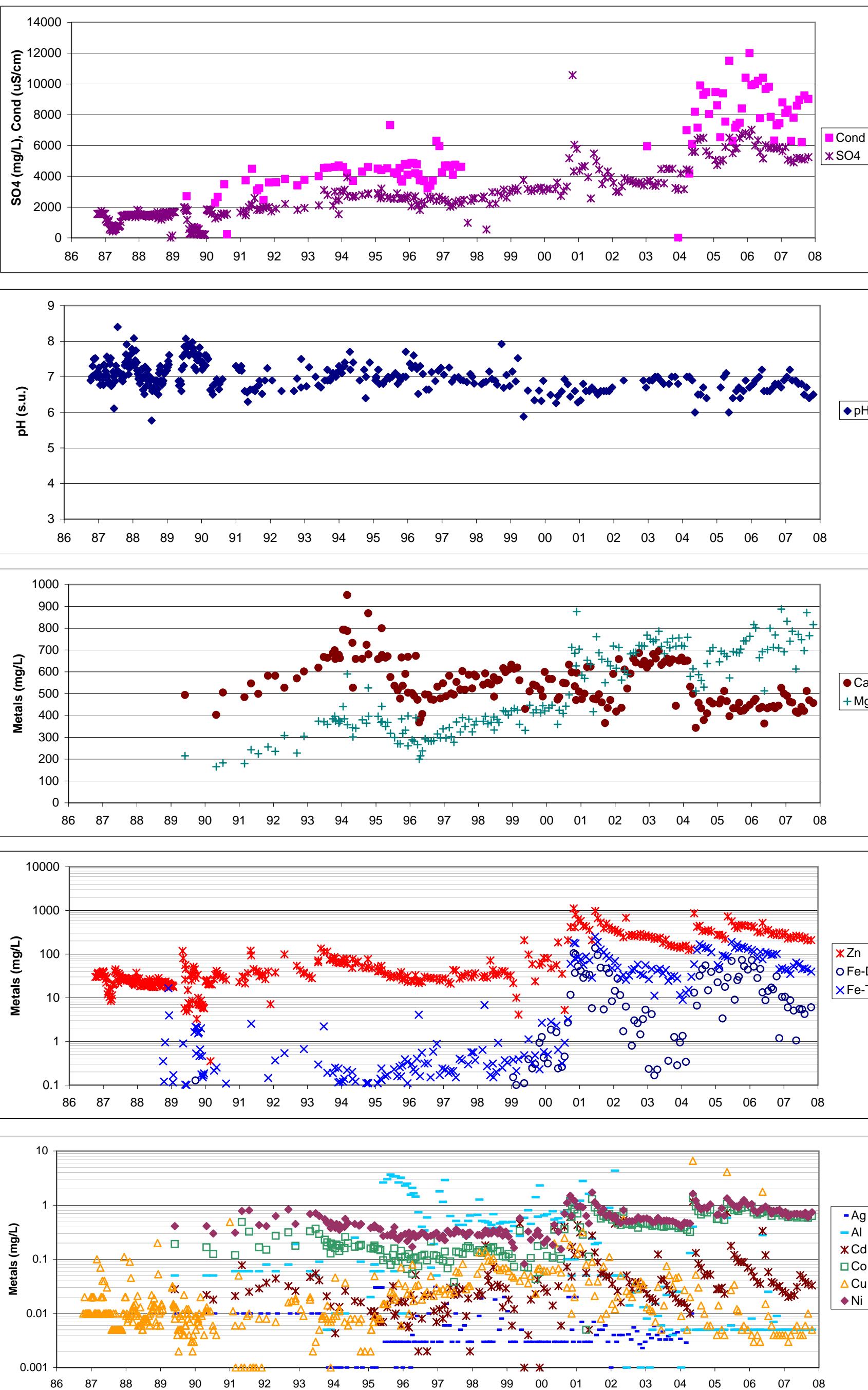


Contour interval: 2m  
 Date of photography: 03/07/25  
 Scale of Photography: 1:20000  
 Survey control derived from existing 1:20000 photography  
 Survey control based on: UTM Projection, NAD27  
 Compiled by The ORTHOSHOP, Calgary, September 2003  
 WO 8856

0 50 100 150 200 250  
 Scale 1: 5000

● Seep Sample Location

### X23



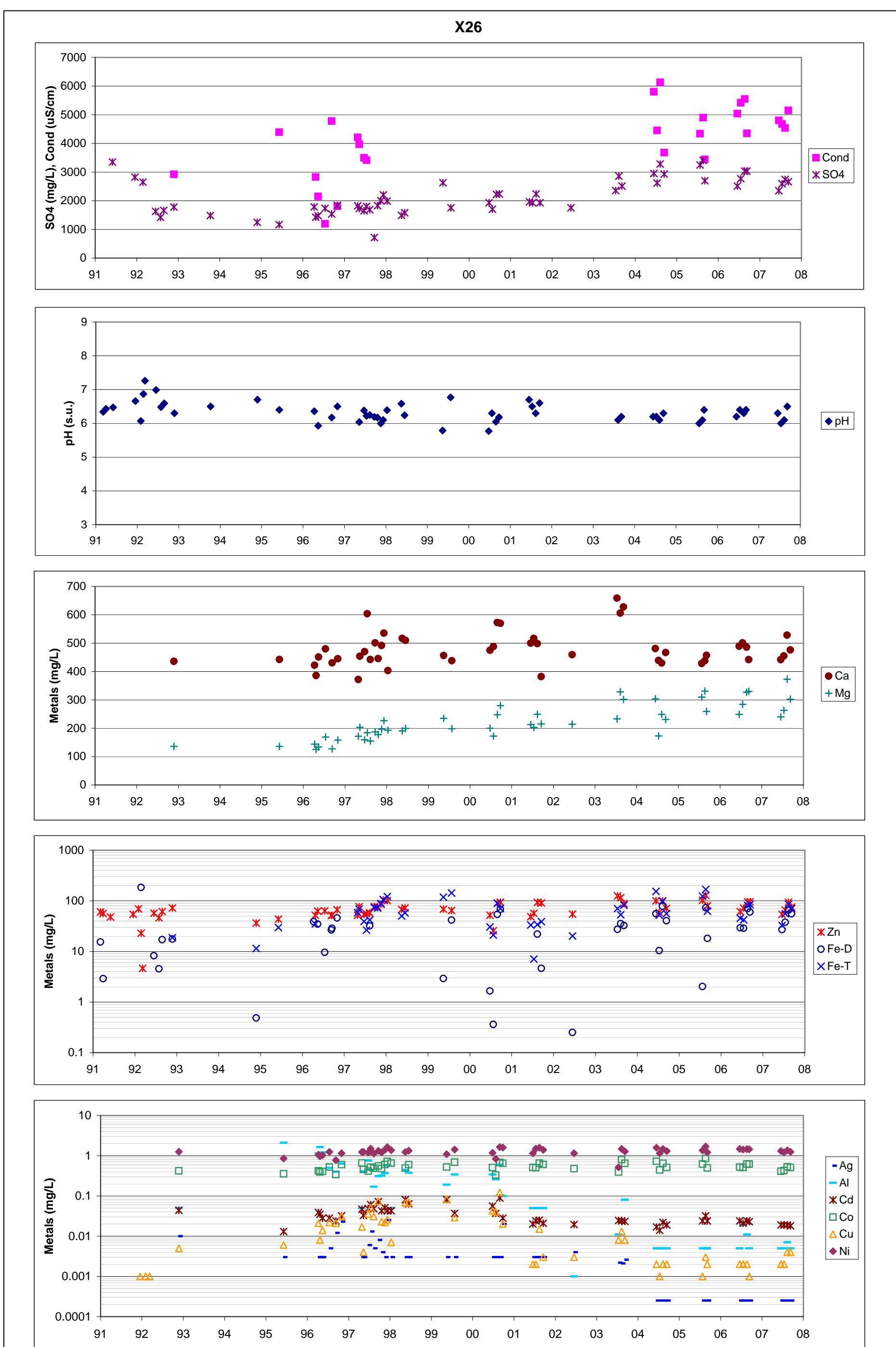
**SRK Consulting**  
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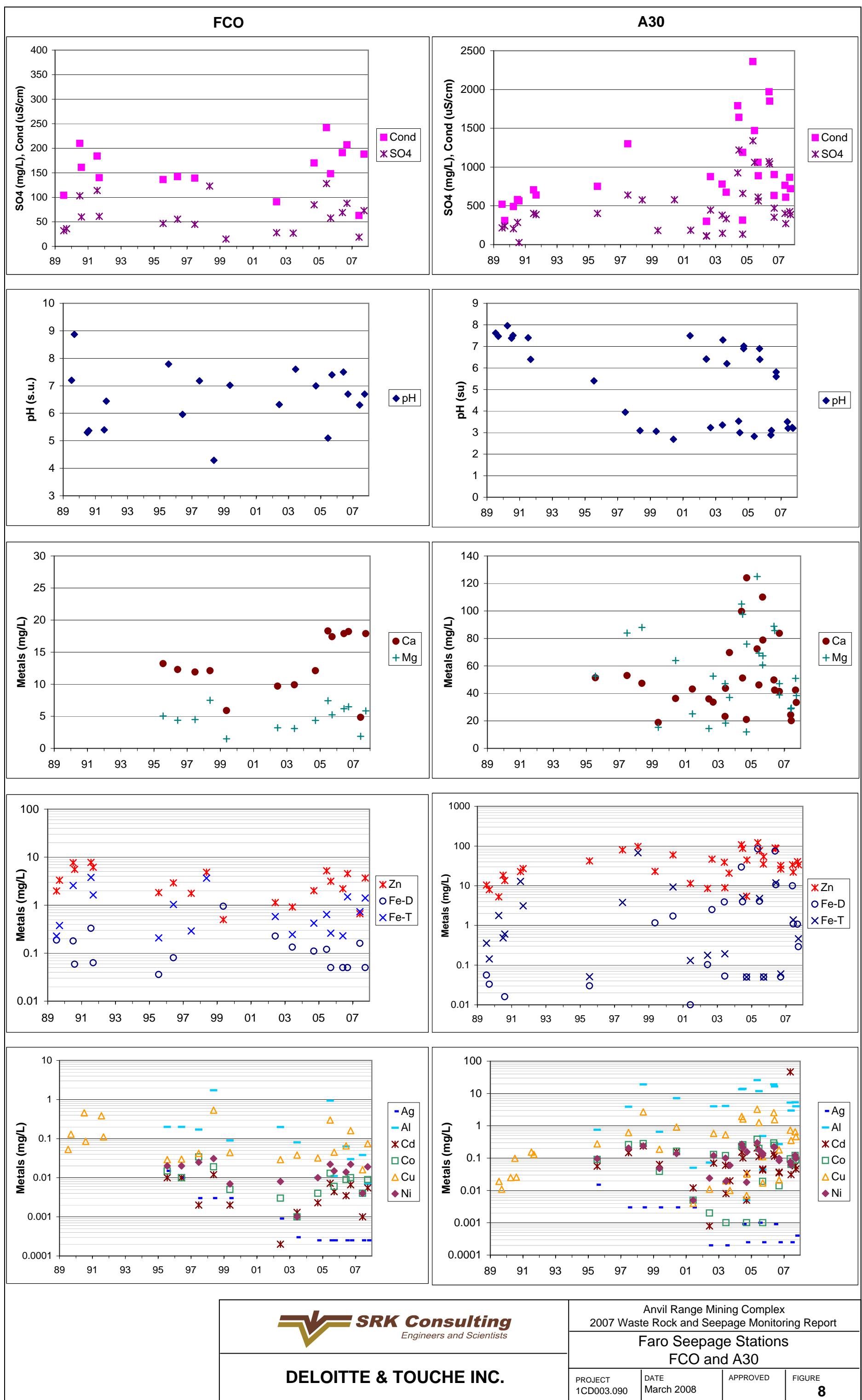
Anvil Range Mining Complex  
2007 Waste Rock and Seepage Monitoring Report

Faro Seepage Station  
X23

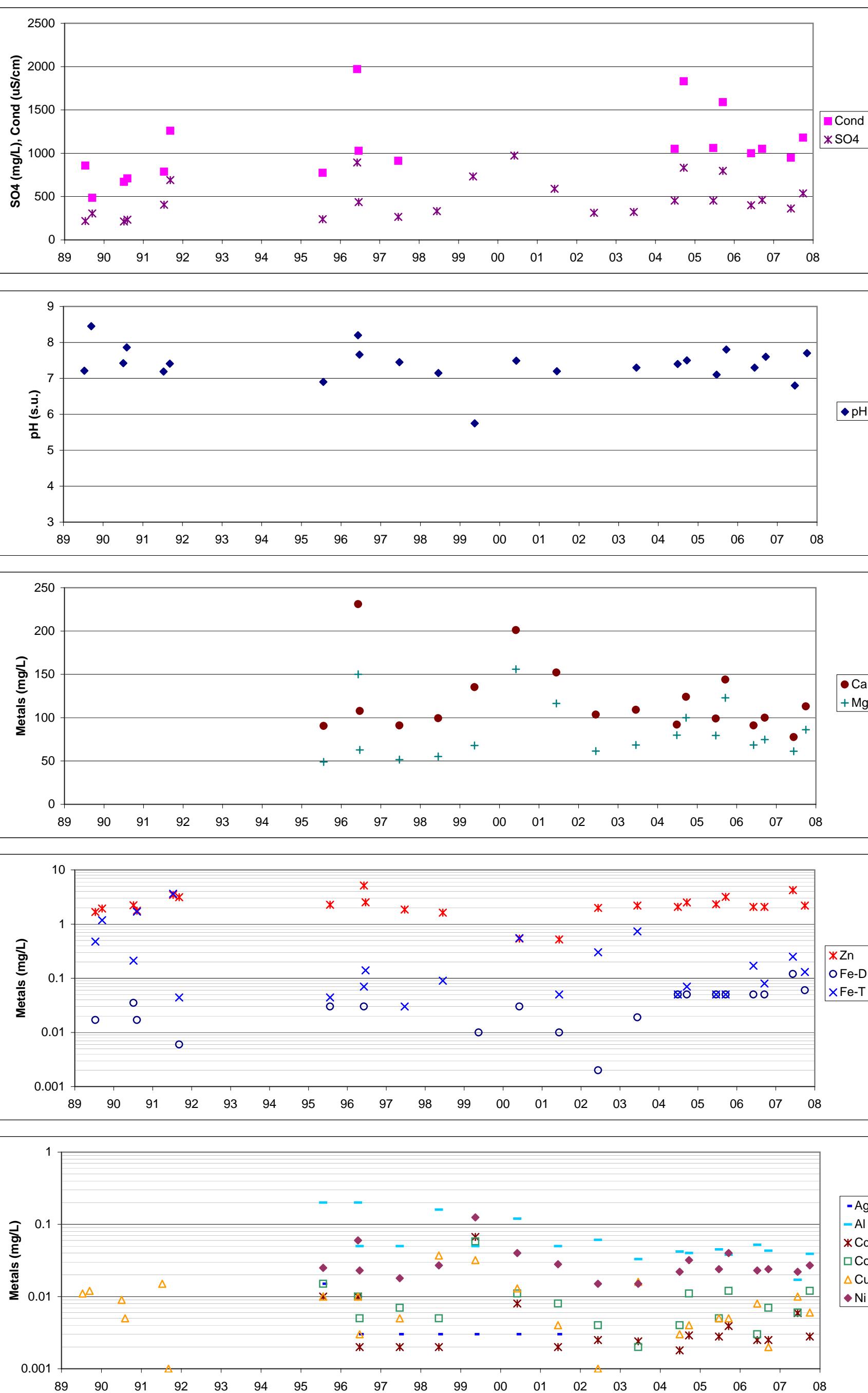
DELOITTE & TOUCHE INC.

PROJECT 1CD003.090	DATE March 2008	APPROVED	FIGURE <b>6</b>
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## SP5-6



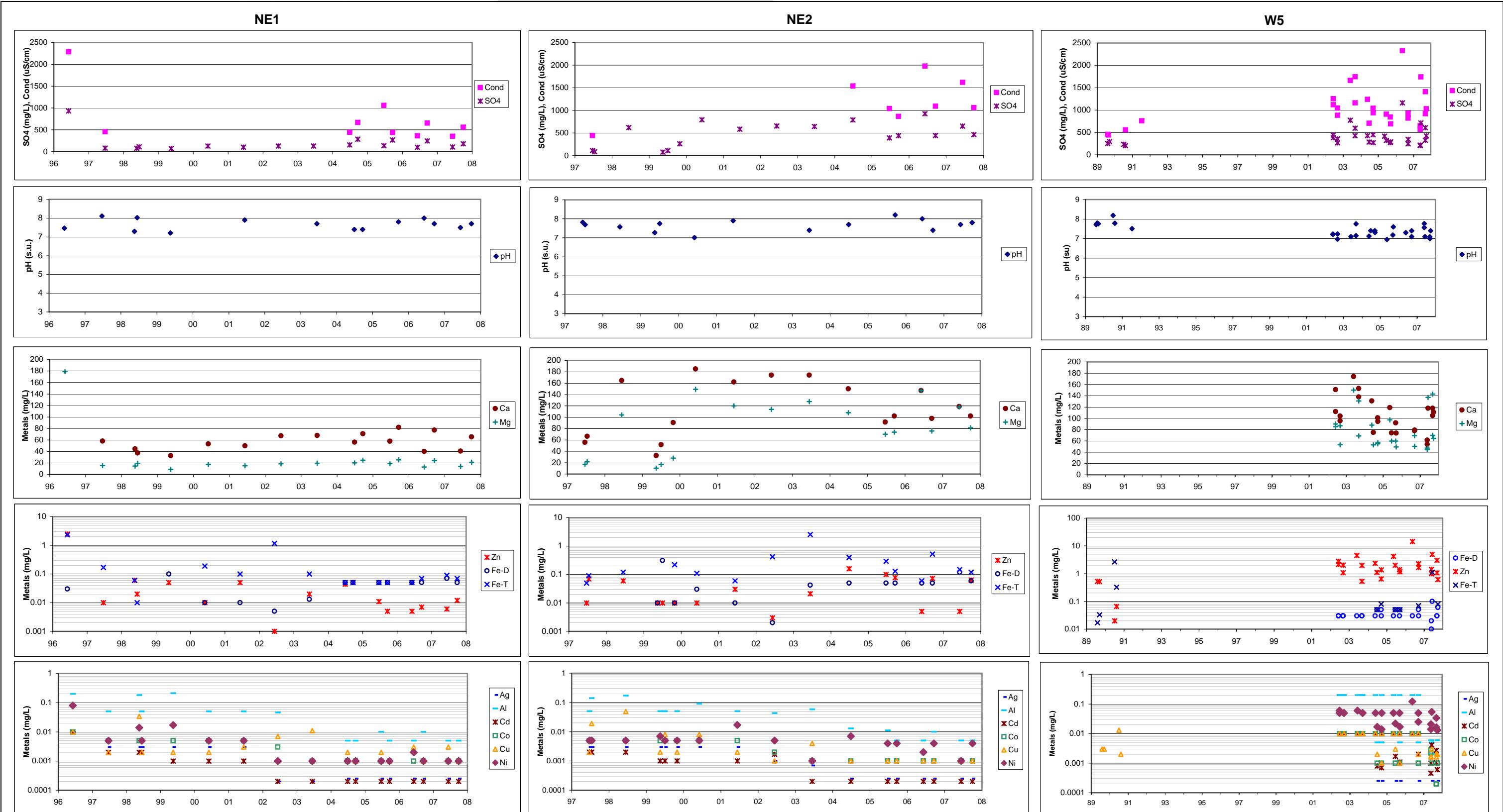
**SRK Consulting**  
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Faro Seepage Station  
SP5-6

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PROJECT 1CD003.090	DATE March 2008	APPROVED	FIGURE 9
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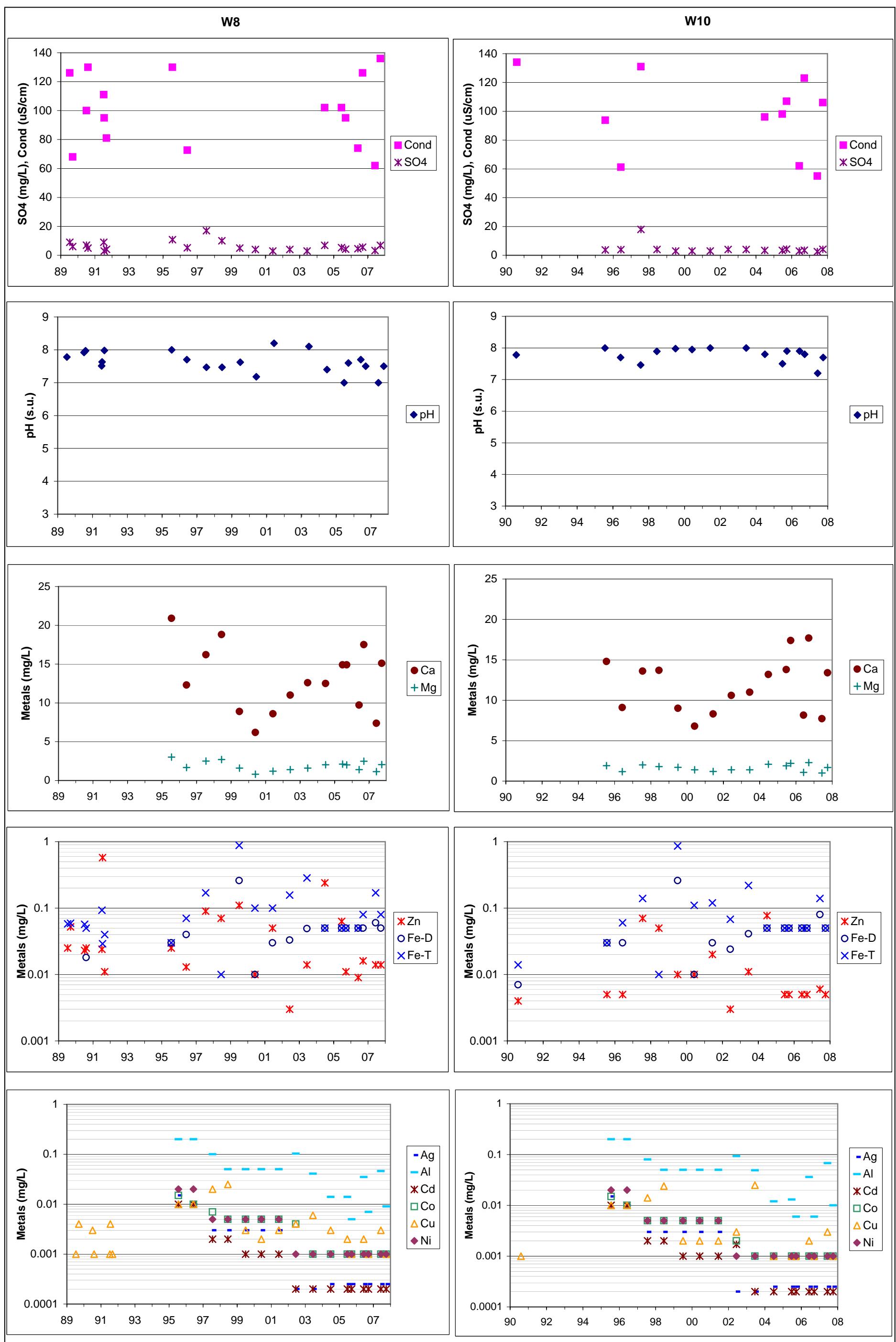


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Anvil Range Mining Complex  
2007 Waste Rock and Seepage Monitoring Report

Faro Seepage Stations  
NE1, NE2, and W5

PROJECT 1CD003.090	DATE March 2008	APPROVED	FIGURE 10
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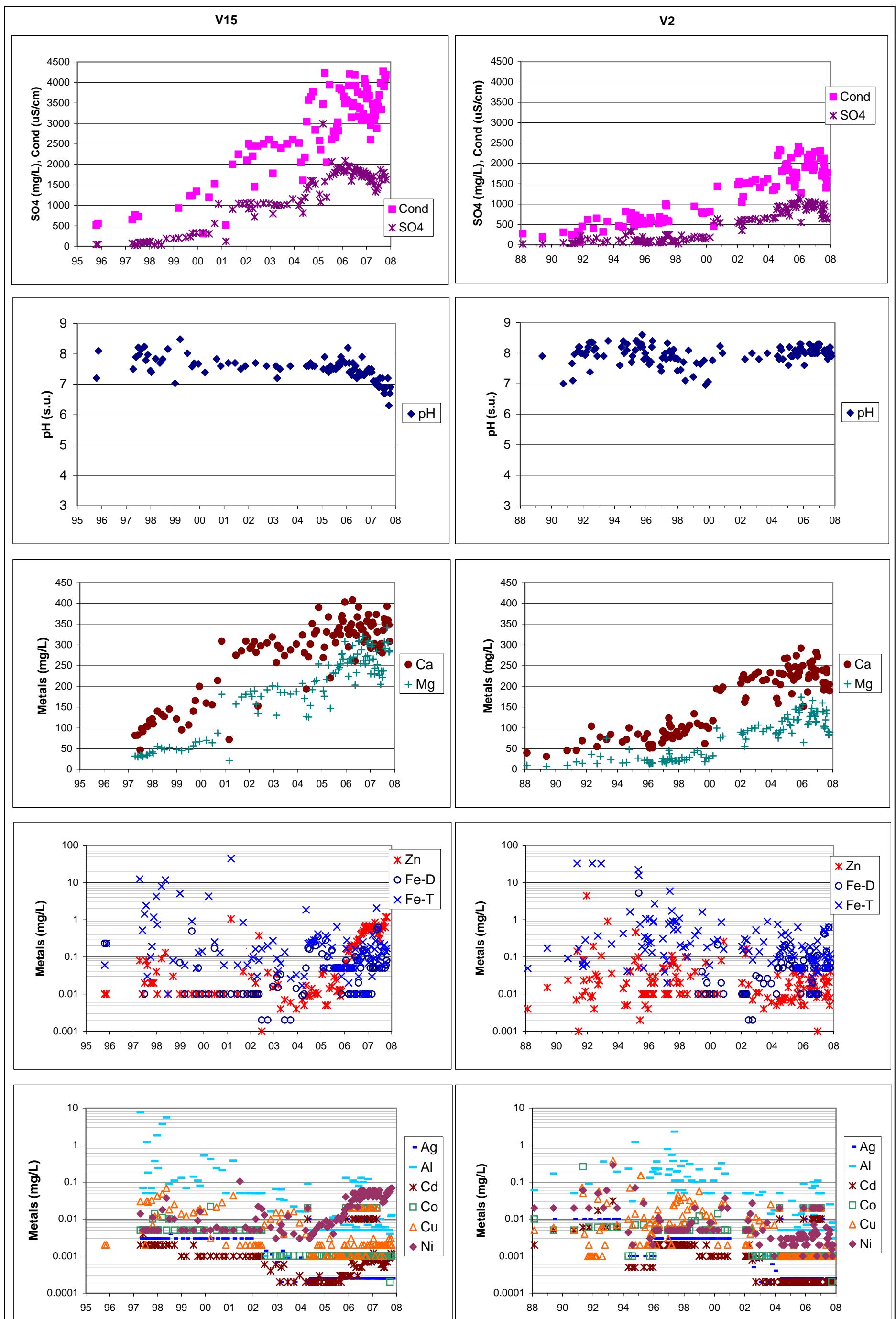
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Faro Seepage Stations  
W8 and W10

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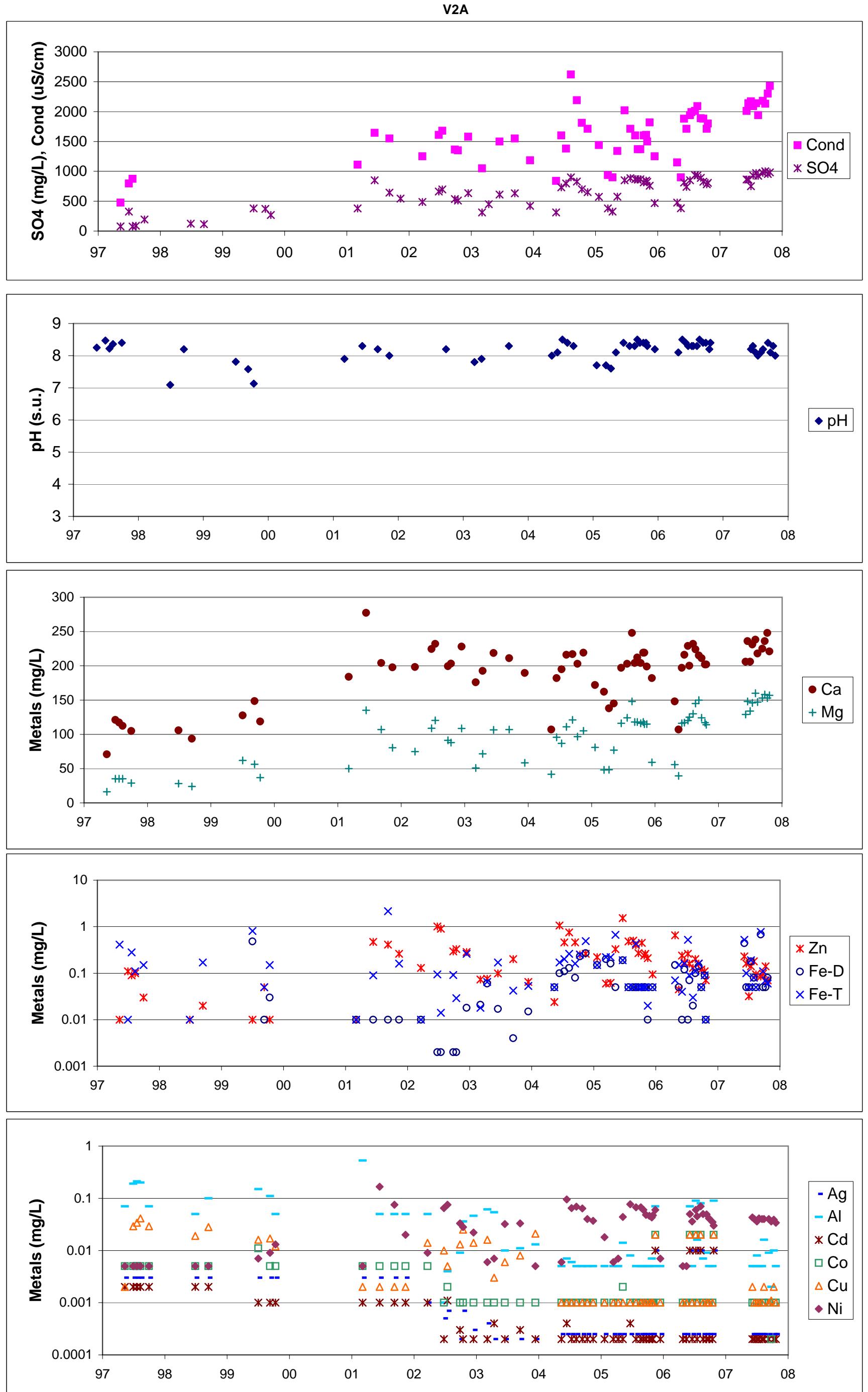
Grum Seepage Stations  
V15 and V2

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1CD003.090

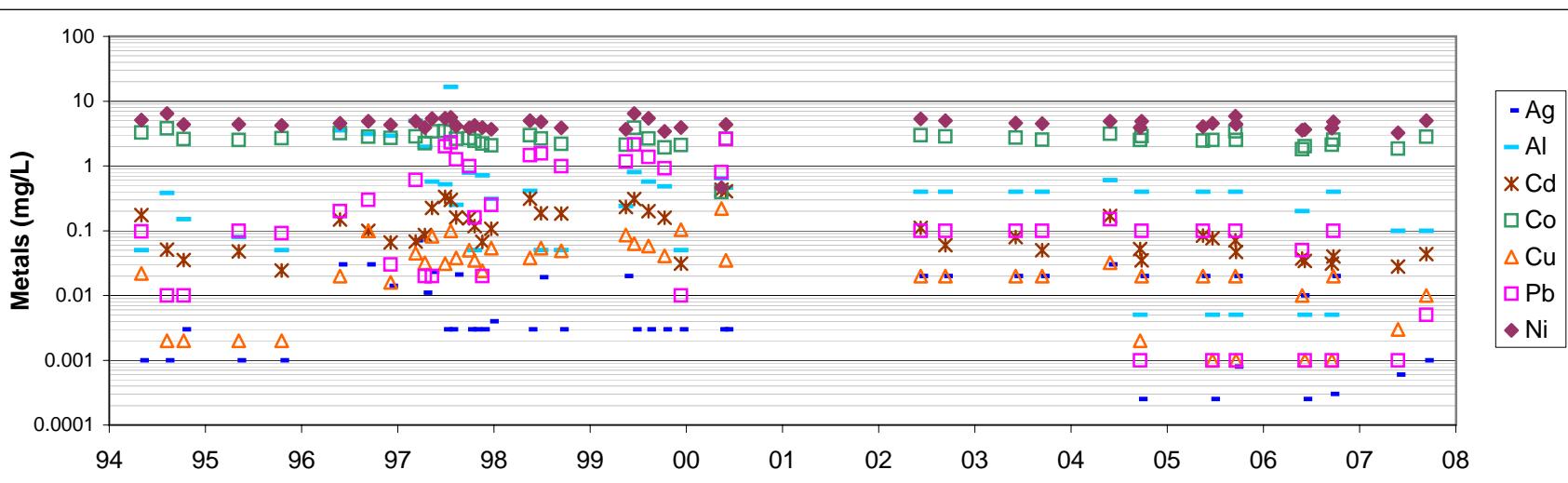
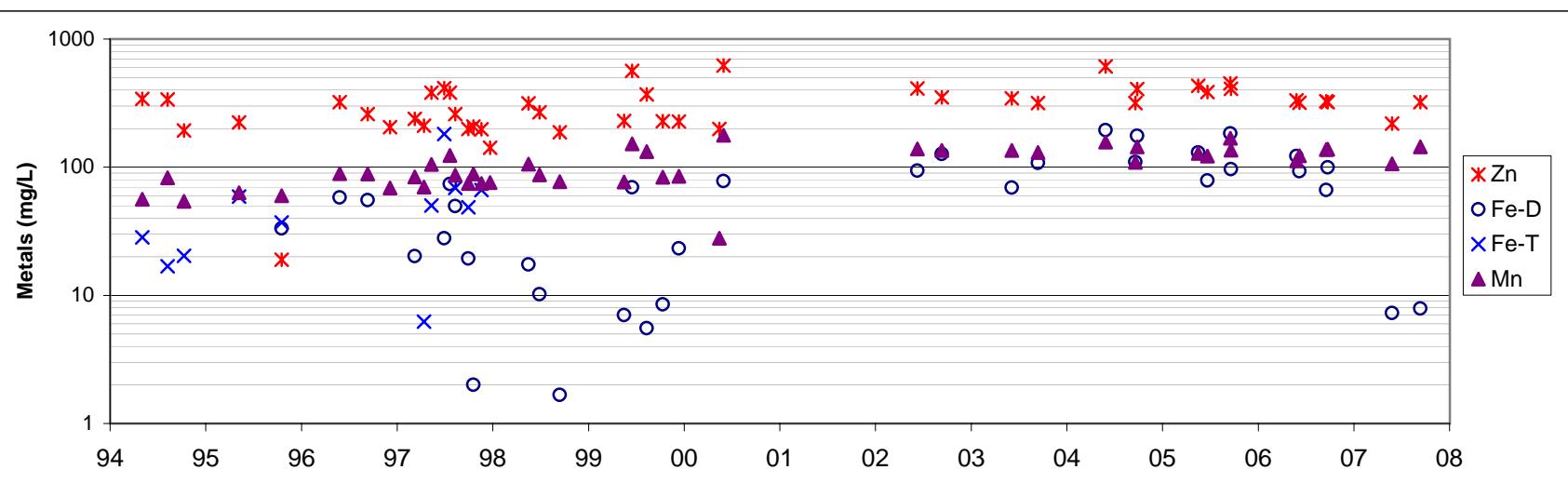
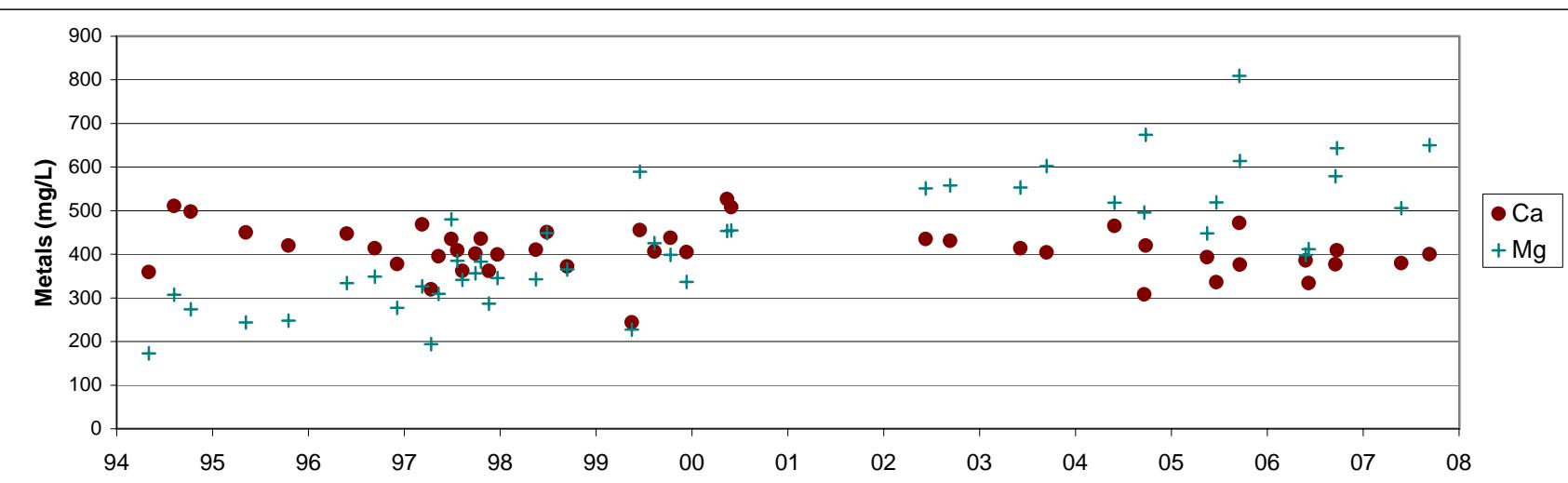
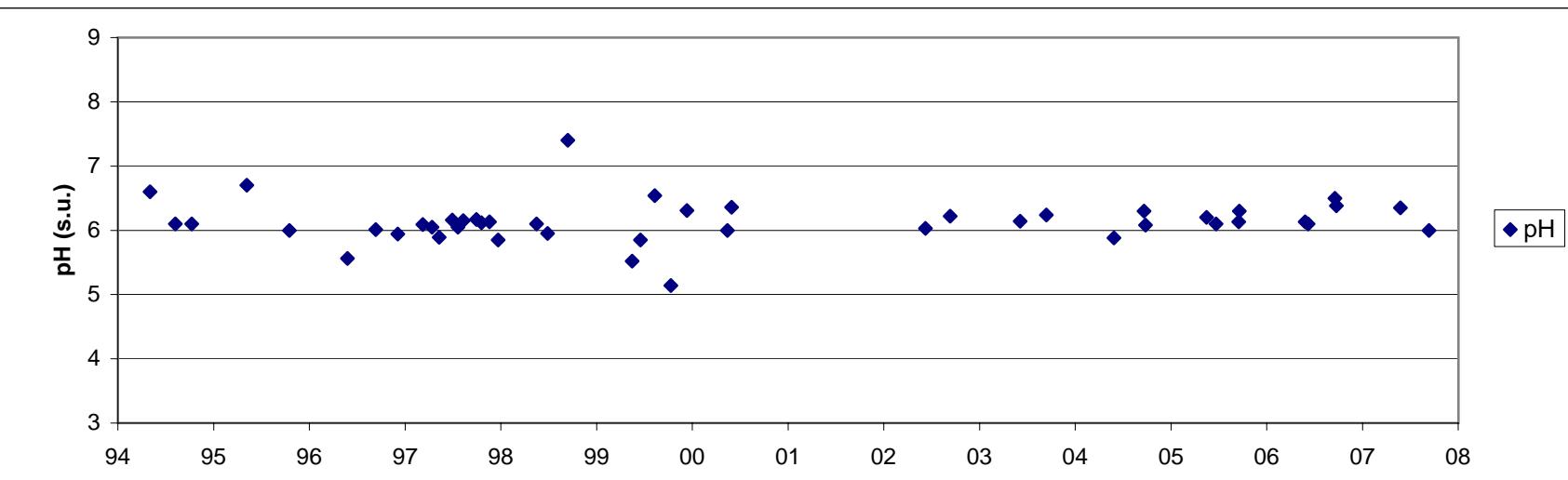
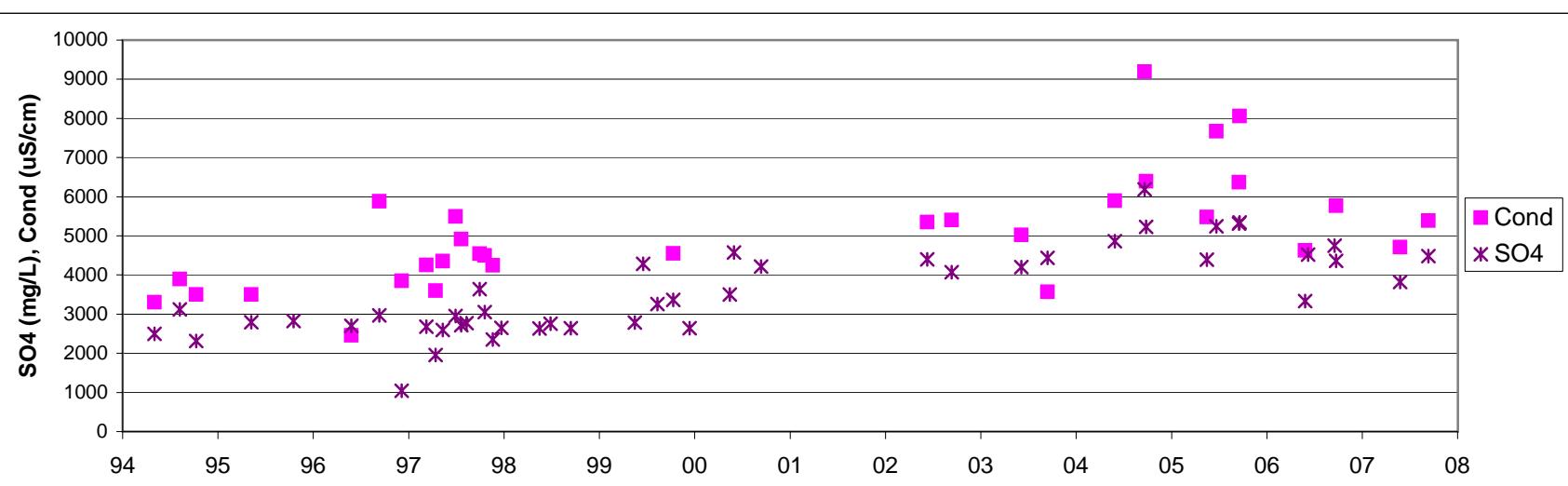
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FIGURE  
**12**



V30



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Vangorda Seepage Station  
V30

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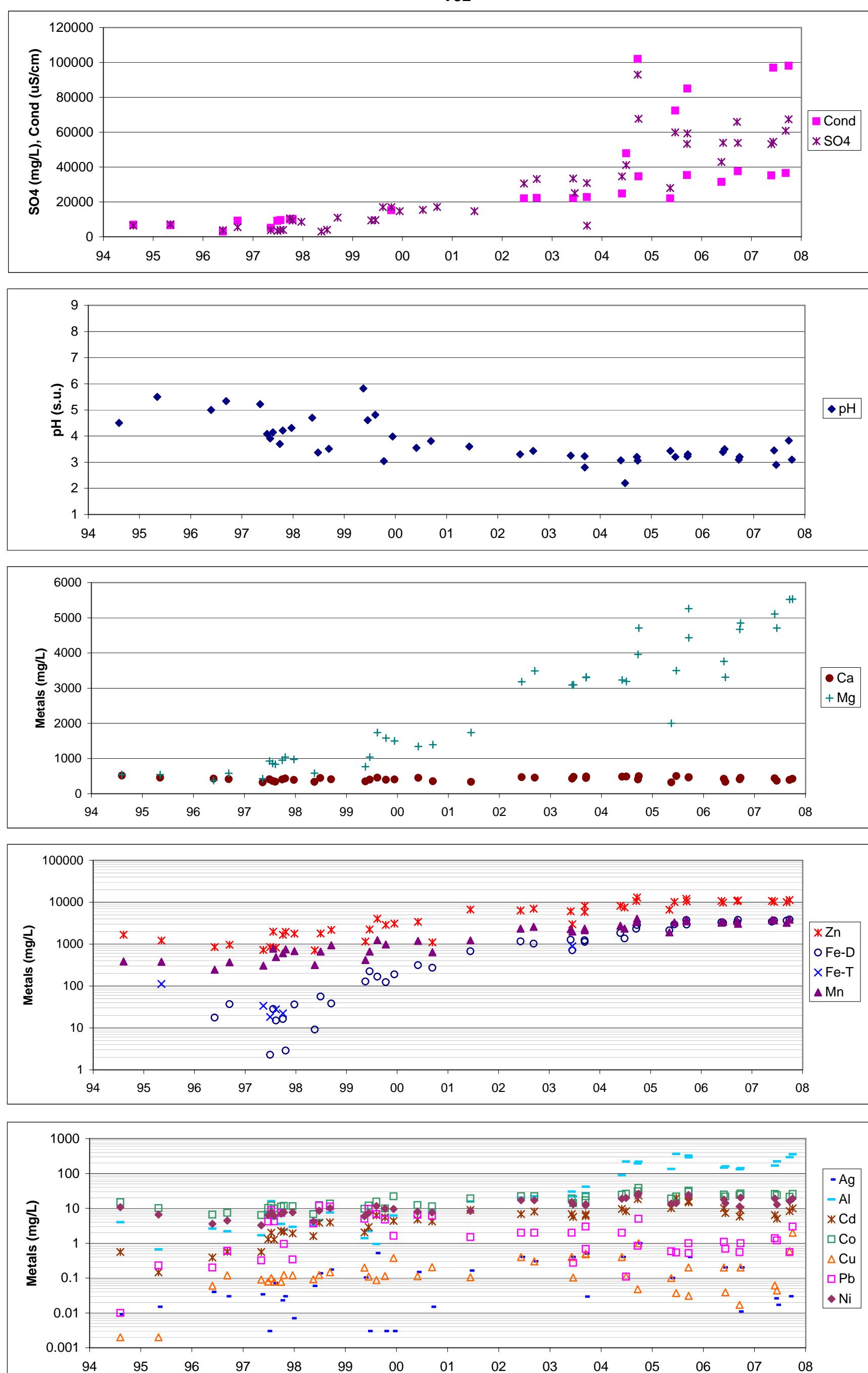
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FIGURE  
**14**

V32



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V32

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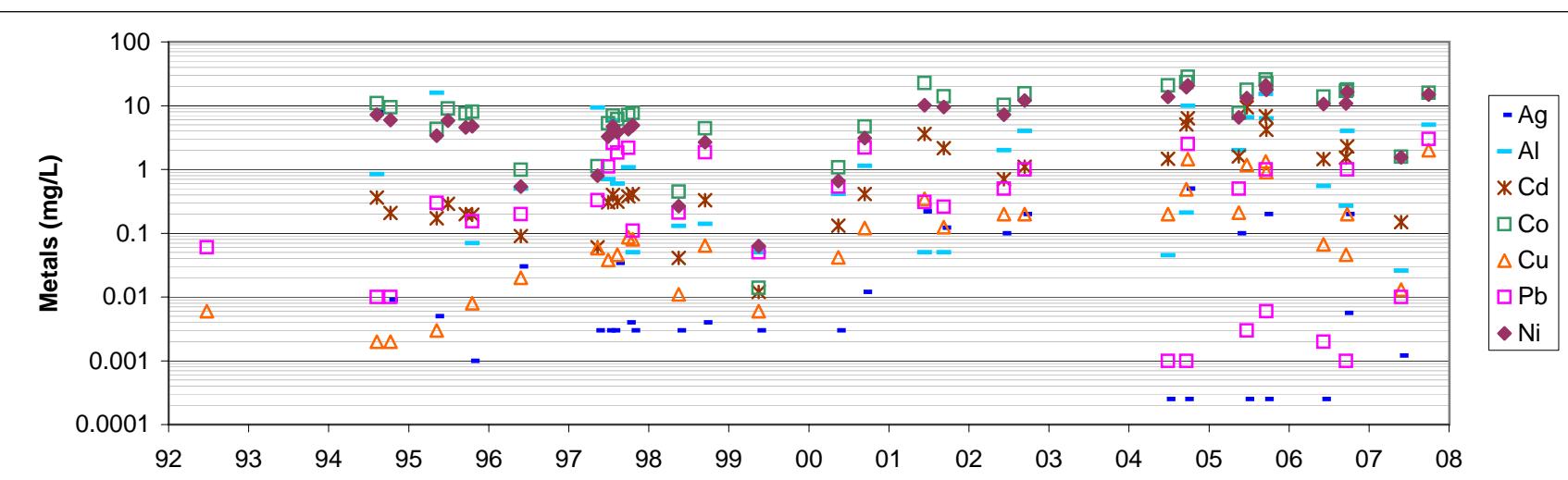
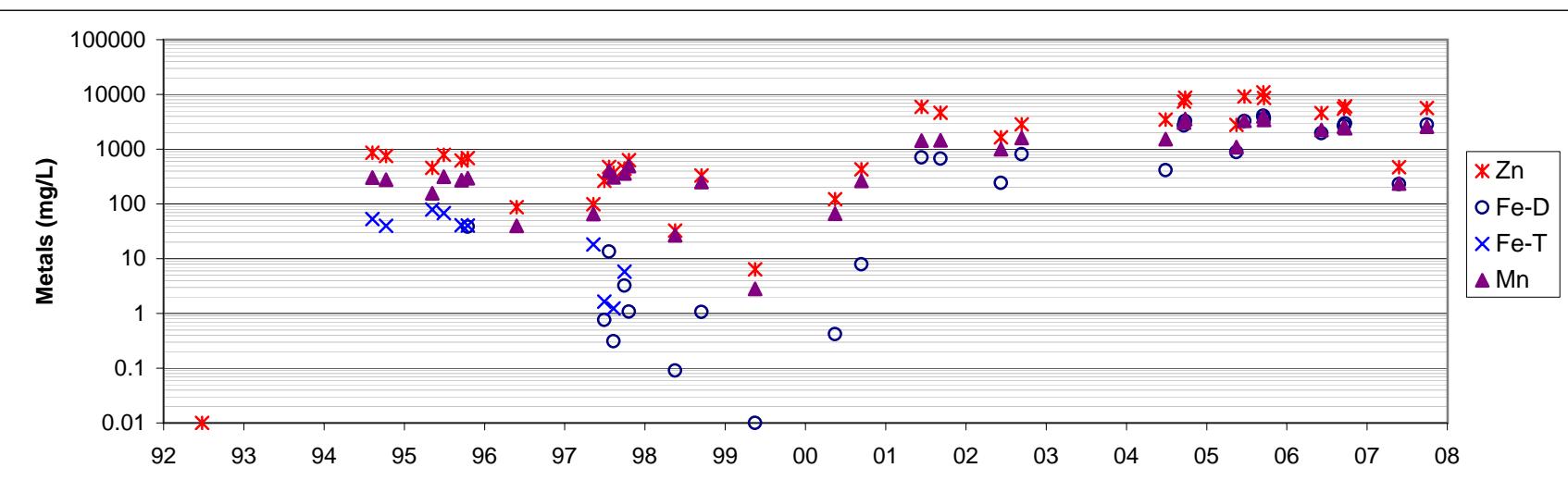
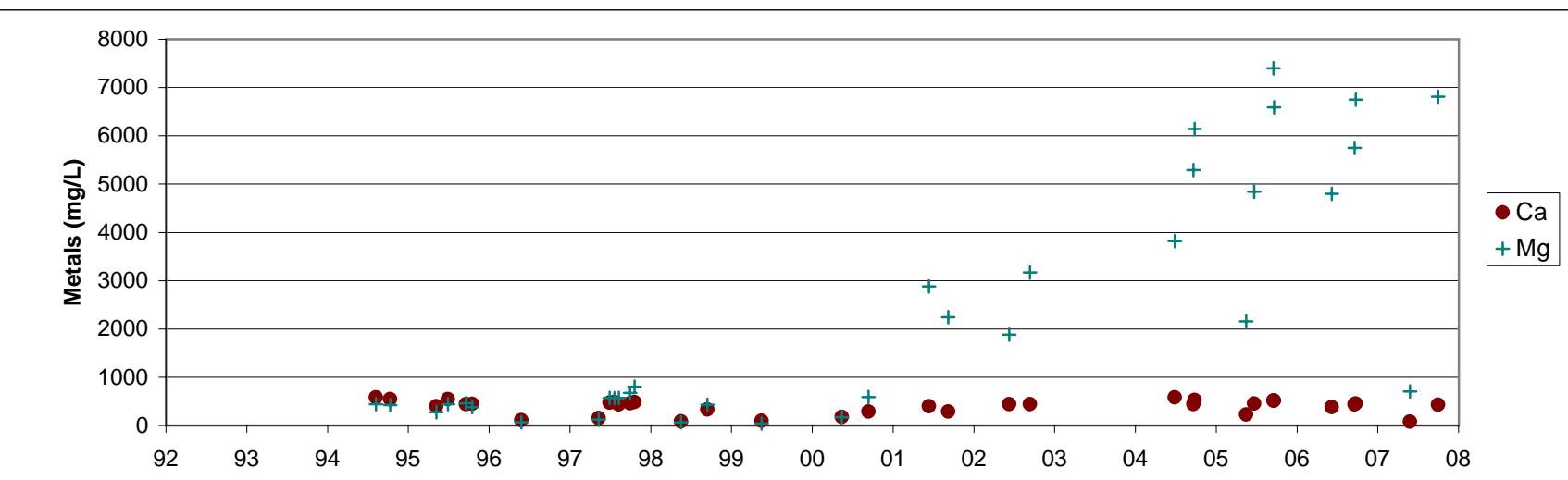
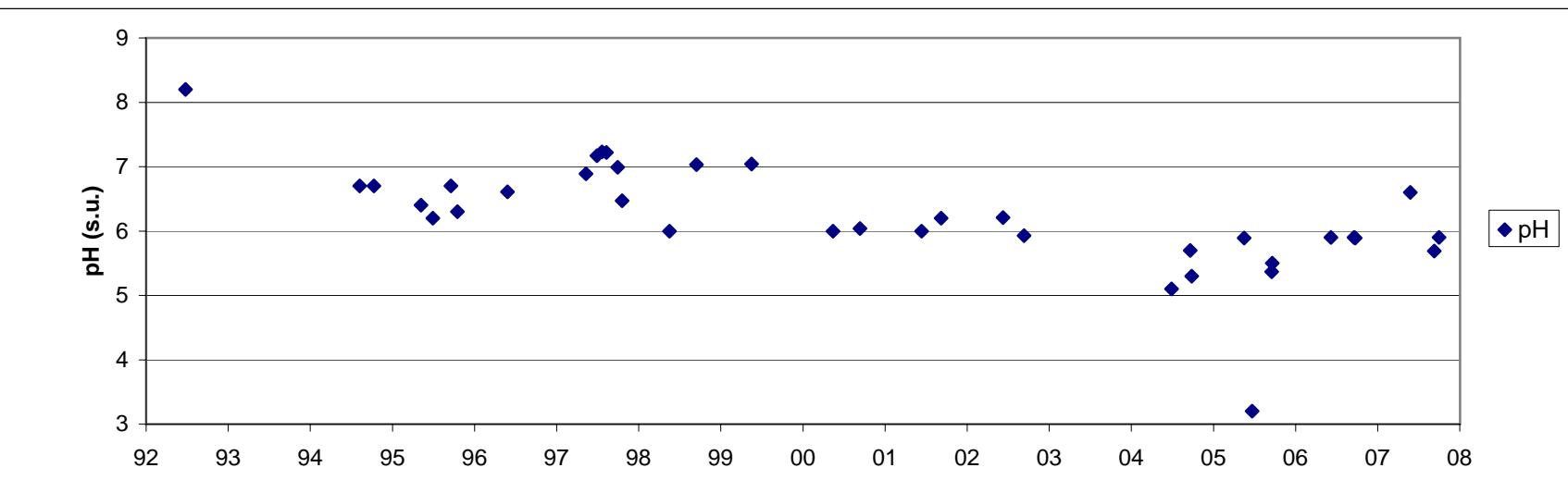
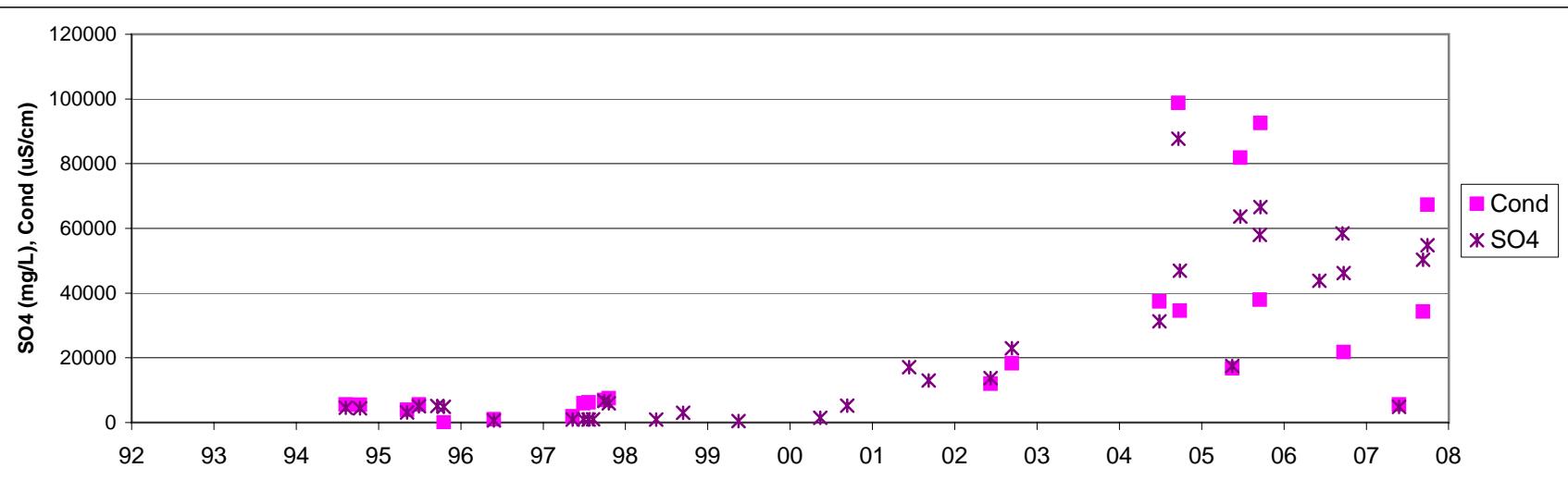
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FIGURE  
**15**

### V33



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Vangorda Seepage Station  
V33

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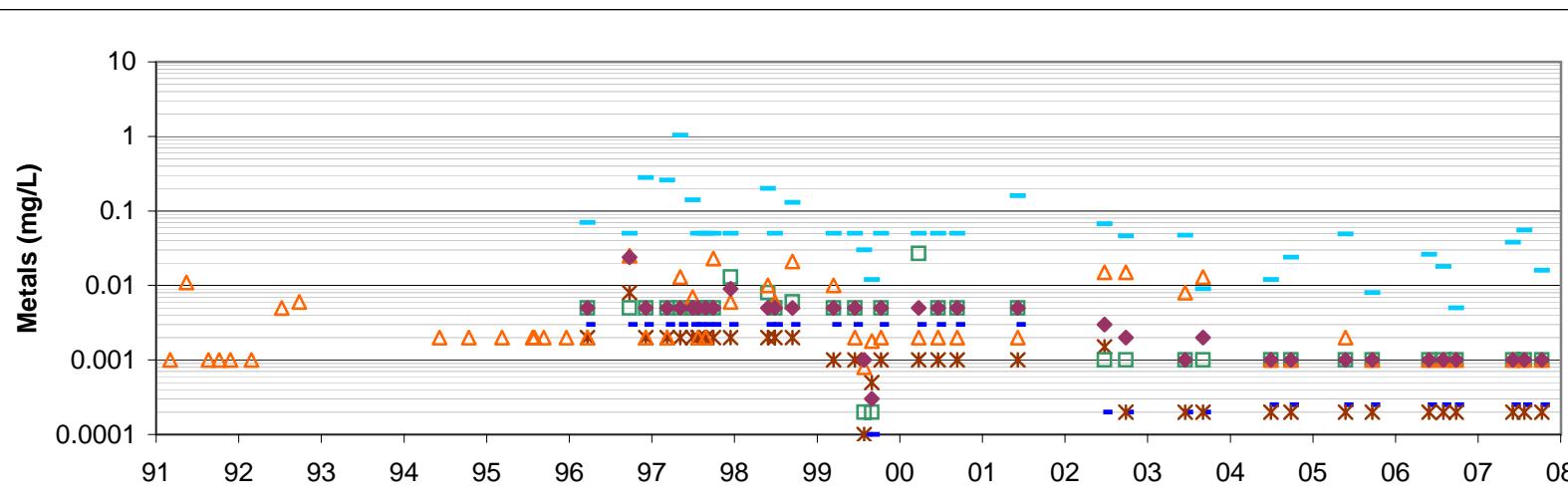
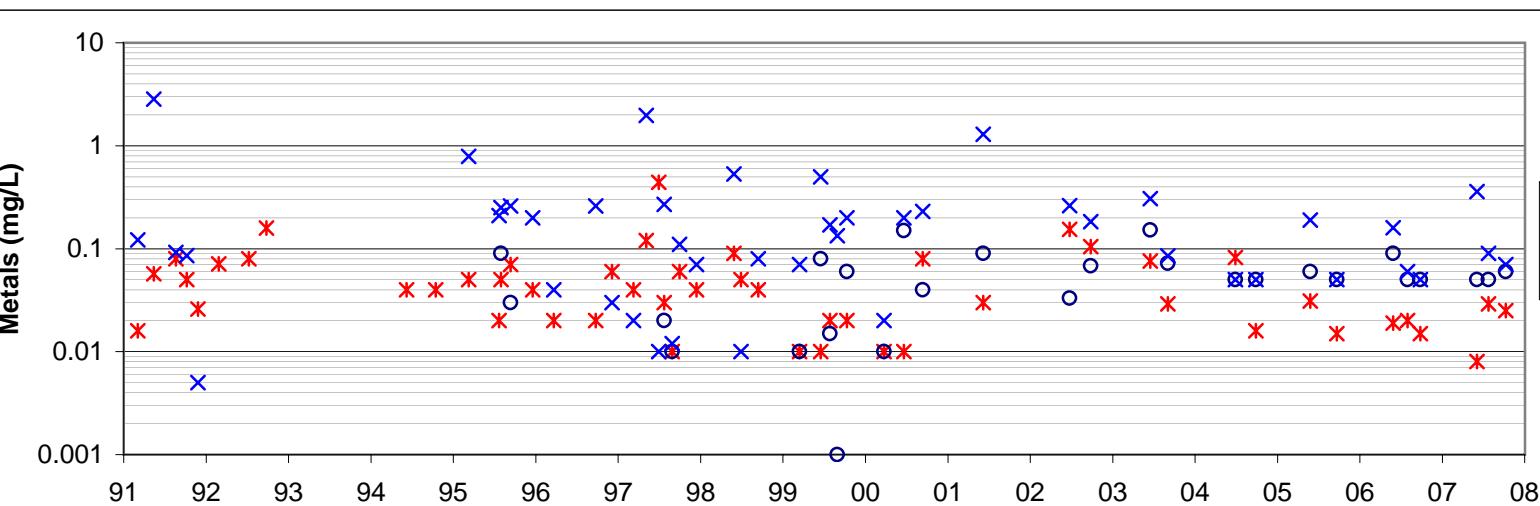
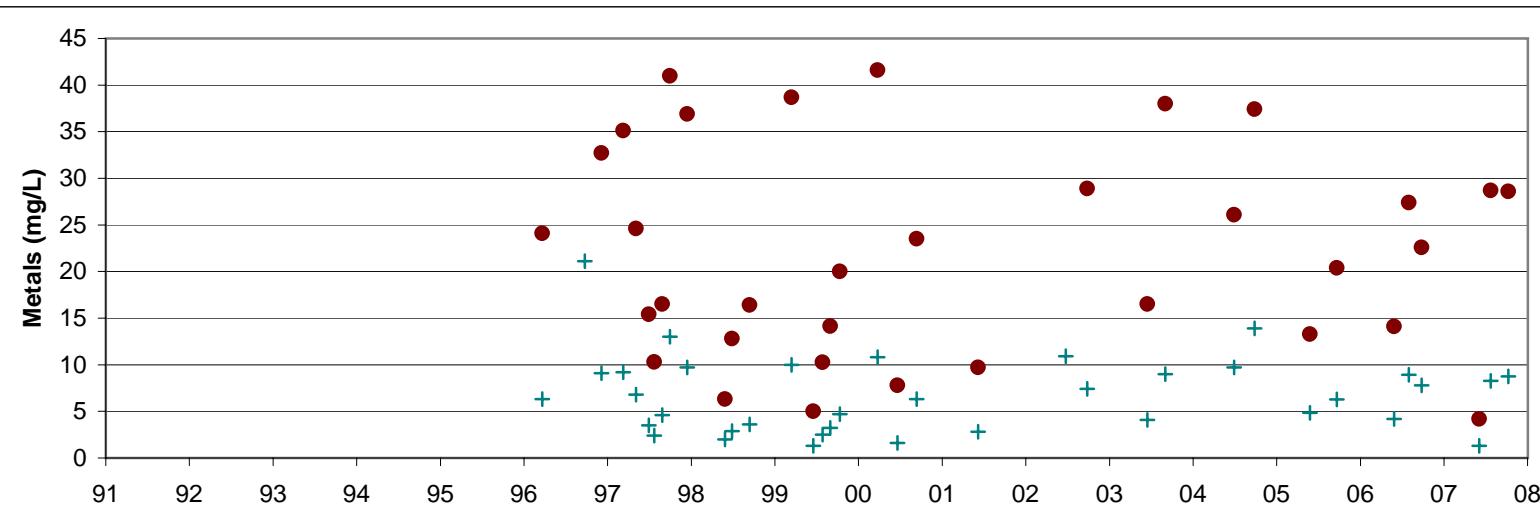
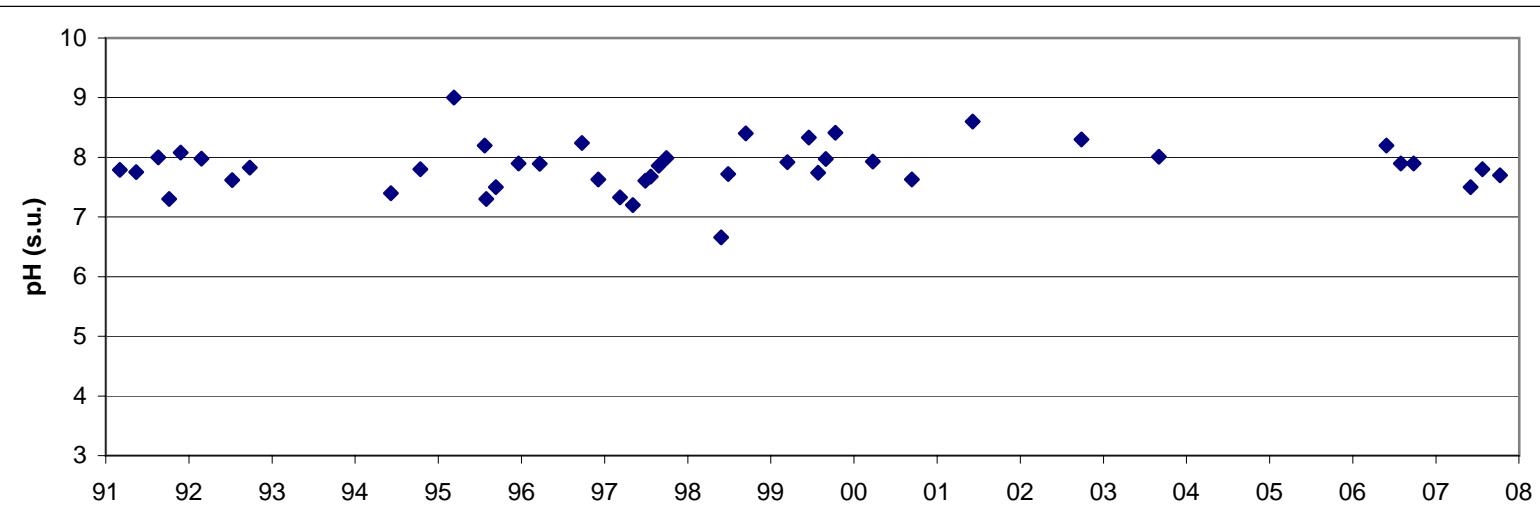
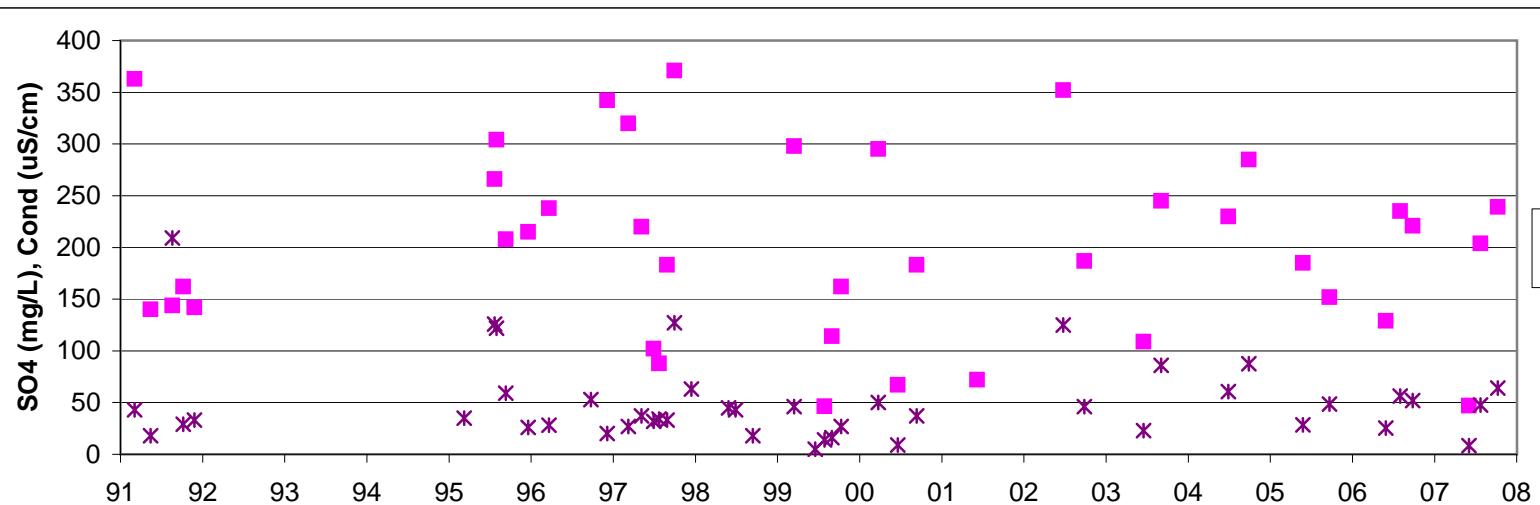
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FIGURE  
**16**

V27



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V27

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FIGURE  
**17**

## **Appendices**

**Appendix A**  
**Waste Rock Seepage Monitoring Results**















## **Appendix B**

### **Pit Seepage Monitoring Results**







## **Appendix C**

### **Thermal and Pore Gas Monitoring Results**











